Title page

Back to the future? Conservative grassland management can preserve soil health in the changed landscapes of Uruguay

Ina Säumel[1], Leonardo R. Ramírez[1], Sarah Tietjen[2], Marcos Barra[3] and Erick Zagal[4]

[1] Integrative Research Institute THESys Transformation of Human-Environment-Systems Humboldt-Universität zu Berlin, Unter den Linden 6, Berlin, 10099, Germany.

[2] Leibniz Institute of Vegetable and Ornamental Crops (IGZ) e.V., Theodor-Echtermeyer-Weg 1, 14979 Großbeeren, Germany

[3] Städtisches Klinikum Dessau, Auenweg 38, 06847 Dessau-Roßlau, Germany

[4] Departamento de Suelos y Recursos Naturales, Universidad de Concepción, Campus Chillán, Vicente Méndez 595, Chillán, Chile

Correspondence to: Ina Säumel (ina.saeumel@hu-berlin.de)

Abstract. The 'soils of the anthropocene' are predominately agricultural. To understand them, we analysed agri- and silvicultural intensification of Uruguayan grasslands in a country wide survey on fertility proxies, pH and trace metals in topsoils originating from different land uses. We observed a loss of nutrients, trace metals and organic matter from grassland, crops and timber plantations. As an example, the cation exchange capacity was 160 percent higher in native forests compared to grasslands and lowest in timber plantations, reaching only half of the CEC in grasslands Acidification of topsoils continues as three fourth of all samples are 'extremely acidic' and 'very strongly acidic' and lowest in timber plantations. Topsoils of riverine forests accumulate more trace metals compared to the other uses. We assume an accumulation in the topsoils of riverine forests, where high levels of nutrients, trace metals and organic matter are found. The translocation of nutrients and organic matter across the landscape to the erosion base depends on local land use trajectories. Increasing soil acidification is driving a positive feedback loop, and land use intensification is leading to degradation of local black soils within a few decades. Our data raises questions about the resilience and carrying capacity of Uruguayan soils with regard to currently implemented highly productive management forms, including the use of timber plantation for carbon sequestration, and supports more conservative forms of extensive management on the grassland biome.

1. Introduction

Human activities alter the bio- and pedosphere, leaving a footprint of such a magnitude that it can be verified stratigraphically (Waters et al., 2016). This unprecedented transformational force is intimately related to the expansion of societies and its productive frontiers, causing a loss of biodiversity, habitat and soil degradation and, consequently, to ecosystem modification (Foley et al., 2005, Borrelli et al., 2017). In this context, soil sciences have transitioned from studies on natural soil formation to the science of 'anthropedogenesis' (Richter, 2020), focussing on the 'soils of the anthropocene' that are predominately agricultural (51 Mio. km^2) or urban (1.5 Mio km^2; FAO, 2019).

The temperate grasslands of South America have historically been characterised by rolling plains and low hills that have been extensively exploited for cattle production and its derivatives since the arrival of European colonization. The Río de la Plata grasslands are one of our planet's four major black soil regions (Durán et al., 2011; Liu et al., 2012), and some of the most fertile soils in the world. Playing an important role in the global food production, these are characterized by a thick, humus and base cation rich and a high cation exchange capacity throughout their profile. Maintaining their properties are therefore crucial to developing sustainable and productive agriculture (Durán et al., 2011; Liu et al., 2012).

Today, the 'Uruguayan savanna' is part of the top-three most critically endangered biomes (Veldman et al., 2015). In recent decades, however, the grassland has decreased due to the expansion of cash crops and *Eucalyptus* plantations, both of which are promoted by national legislation including land grabbing and trans-nationalization (Piñeiro, 2012). This land use intensification, with its increased input of energy, nutrients and pesticides, leads to an overall loss of soil fertility and increasing toxicity related to acidification, salinization and contaminants (Liu et al., 2012; Borrelli et al., 2017). Ecological, economic and cultural functions of soils are severely degraded, and the degradation of black soils in South America is of particular concern because they have only been heavily exploited for a comparatively short period of time (Durán et al., 2011).

Since the first decades of the twentieth-century, compared to other disciplines, soil sciences have received an extraordinary amount of attention in Uruguayan academia, governance, the productive sector, and also in the general public, resulting in a national soil inventory program in 1965. The subsequent classification of Uruguayan soils and their productivity (CONEAT Index) remains an important source for today’s land taxation and for management plans by the legal conservation regulations, and provides a detailed classification that takes into account soil type, texture, natural vegetation, altitude and geology (Lanfranco and Sapriza, 2011).

As soil degradation is extremely relevant for countries like Uruguay, which are socioeconomically dependent on their soils (Zubriggen et al. 2020), it is a topic of discussion for local farmers, academia, and the public. An actualization of the state of the art of soils and related processes is needed (García-Préchac et al. 2004; De Faccio et al. 2021), particularly as there has been little study of the impacts of the Uruguayan grassland intensification on soils properties (Beretta-Blanco et al., 2019). At the same time, while a new paradigm for grassland intensification with a wide set of means including fertilization has been proposed to increase economic and environmental sustainability (Jaurena et al., 2021), it is urgent to get more insights into the dynamics of nutrient in soils of Uruguay and their availability for crops (Beretta-Blanco et al., 2019).

Soil classifications are mainly based on subsoils. However, we focus on topsoil as the most relevant and very responsive interface for ecological processes and farmer's management, since understanding the state of the art of topsoils and its processes is crucial for developing recommendations for sustainable land management practices. Due to the diversity of perspectives on soil quality and health and related ecosystem services, operational procedures for evaluation of soil functioning are still lacking (Bonfante et al. 2021). We contribute to a better understanding of globally occurring degradation processes in the field of tension between desired soil productivity, yield limits, especially in erosion sensitive soils, and necessary soil conservation.

We therefore explored soil parameters describing current chemical conditions of topsoils that are parts of different soil groups and orders, and Uruguayan soil categories. Specifically, in order to explore the gains and losses of macro and micro-nutrients and soil organic carbon across landscapes and to determine the impact of land use change on acidification and trace metal mobility and related trade-offs with soil degradation and conservation. In detail we address the following question: i) how do fertility proxies such as soil organic carbon and content of nutrients, acidification (pH) and trace metals accumulation in topsoils vary across different land uses? Thus, we expand the knowledge across land uses from more natural to strongly modified uses and discuss the results in light of different degradation processes such as erosion, depletion of nutrients or carbon, acidification and accumulation of pollutants and in the light on current debates on intensification.

2. Material and Methods

2.1 Study area and design

Uruguay covers about 176,000 km^2, and has a population of 3.5 million, mainly in urban areas. The Western half of Uruguay is dominated by Mollisols, developed on a wide range of sediments from the Devonian to the Cenozoic era (partially associated with Vertisols), while the Eastern plains and wetlands have, in addition to Mollisols, other soil types on the slopes, rocks and on flood plans (i.e. Alfisols, Ultisols, Entisols and Inceptisols; Durán et al. 2011). During a country-wide survey from December 2015 to March 2016, we collected 280 topsoil samples of 0-10 cm depth from 101 plots (50x50m in grassland and crops; 100x100m in native forests and timber plantations) distributed at 28 monitoring sites throughout Uruguay, South America (Fig. 1a-b), using a stratified random design. In the first step, we randomly selected monitoring sites across the country. In the second step, we contacted landowners to explore their willingness to establish a long-term monitoring site. If the owner agreed, plot selection was stratified by different rural land use types: grassland, timber plantations of *Pinus* and *Eucalyptus* species, native forest, and crops. Native forests cover mainly riverine and park forests. The later are a savanna like transition zones between riverine forests and the open grasslands. We subdivided grassland plots according to the intensity of use: (i) undisturbed grassland (without grazing), (ii) partially grazed grasslands (with sporadic grazing and low animal charge), and (iii) highly grazed grassland (with high animal charge). Land use change from 1986 to 2017 follows basically three different trajectories: i) the expansion of timber plantations over grassland leading to a disaggregation of grassland by timber plantations; ii) cropland expansion where crop cover maintains the open landscape character of former grasslands, grassland conservation where large and regularly interconnected riverine forests in a landscape dominated by grasslands (Ramírez and Säumel 2021) and grassland intensification changing from natural grassland to so called 'improved' or artificial grasslands (Modernel et al. 2016; Jaurena et al. 2021). Fertilization and application of other agrochemicals is standard procedure in timber plantations, artificial grasslands and industrial crops. We sampled top soil three times at each land use at the edges of the plot, and stored samples below 7°C until laboratory processing.

2.2 Analysis of Soil Samples

For gravimetric determination of soil moisture, the topsoils samples were dried at 105 °C until constant weight. Next, lumps in the samples were broken down and the remaining plant material was removed before sieving (2 mm) and ground.

We analysed 280 samples regarding macronutrients, pH and trace metals and 80 samples for soluble cations and micronutrients. Among fertility-related variables, we measured the total amount of the macronutrients phosphorus (P), organic carbon (SOC) and nitrogen (N), so obtaining the C/N ratio. To determine total carbon and nitrogen, the samples were sieved again (0.5 mm) and analyzed using a LECO TruSpec CN (USA) at a combustion temperature of 950 °C with ultra-pure oxygen. In addition, the presence of SOC was tested for by adding concentrated hydrochloric acid, the presence of carbonates was assessed to obtain the amount of SOC.

We determined concentrations of the soluble cations calcium (Ca), magnesium (Mg), potassium (K), sodium (Na), and the micronutrients copper (Cu), zinc (Zn), manganese (Mn) and iron (Fe) by atomic spectroscopy (Unicam AAS Solaar 969, Thermo Electron Corporation, USA), extracting with ammonium acetate (1 mol l^{-1}, pH 7), and with DTPA-$CaCl_2$-TEA at (pH 7.3). We calculated the cation-exchange capacity (CEC). Acidity was measured by adding calcium chloride (0.01M) to the samples in a 2.5:1 proportion, and after shaking and two hours rest, read with a pH meter (HI2550 meter, Hanna Instruments, USA)). For these variables, each tenth sample was duplicated. We categorized acidity using the USDA Natural Resources Conservation Service classification (Kellogg, 1993).

To determine the concentrations of arsenic (As), cadmium (Cd), chromium (Cr) and lead (Pb), samples were further sieved as before (0.5 mm), weighed out into a digesting container, and extracted with a mixture of nitric acid (70%) and hydrochloric acid (30%) in a 1:3 proportion. The digestion took place in a Titan MPS (Perkin Elmer, USA) microwave at a programme suitable for this digestion (Method 3051A). During the proceedings, only deionized water Type I ASTM1193 (EC max 0.06 to 0.1 μS/cm) was used. Reagents were used to eliminate traces of other materials and to avoid contamination of the samples. The trace metals were determined by inductively coupled plasma optical emission spectroscopy using the Optima 8000 (Perkin Elmer) with metal-associated wavelengths of 193.696 nm for As, 228.802 nm for Cd, 267.716 nm for Cr and 220.353 nm for Pb.

Total P concentration of phosphorus was determined calorimetrically after microwave-assisted digestion with Unicam spectrometer at a wavelength of 660 nm. For all these variables, a repetition for each round of the microwave digestion was made.

2.3 Soil classification

We intersected the coordinates of the centre of the plots with maps containing geospatial information on the classification of the Uruguayan soils using ArcGIS 10.3 (ESRI, 2018). For Soil Groups classification, we used the of the World Reference Base for Soil Resources (WRB; IUSS Working Group, 2015); for Soil Orders, the USDA soil taxonomy (Soil Survey Staff, 1999); and for the local Uruguayan classification, Soil CONEAT (Comisión Nacional de Estudio Agronómico de la Tierra) Groups categories, which include productive capacity of cattle and sheep (MGAP, 2020). The CONEAT groups are defined by their productive capacity in terms of beef, sheep and wool expressed by an index relative to the average productive capacity of the country, to which the index 100 corresponds. The classification is based on photo-interpretation at a scale of 1:40,000, field verifications and physico-chemical analysis of the soils. The productivity indices correspond to soil groups. The CONEAT groups have been defined by the dominant and associated soils according to the Soil Classification of Uruguay. The groups are related to the units of the Soil Reconnaissance Chart of Uruguay at a scale of 1:1,000,000. For each group, some important soil properties and associated landscape characteristics are indicated. The nomenclature of the CONEAT groups correlates with the Soil Use and Management Zones of Uruguay. The Soil Groups are superimposed on the rural parcel and are represented in the CONEAT cartography at a scale of 1:20,000 (for more details see MGAP, 2020).

2.4 Data Analysis

In a first step, we explored and prepared our database for further analysis. Exploring the distribution of the soil parameters in R (R Core Team 2021), ruled out the normality of the data using the Shapiro-Wilk Test and the homogeneity of variances with the Fligner-Killeen Test. We tested for outliers using the 1.5-3 IQR threshold and the function *outlierTest* from the R package *car* (Fox and Weisberg 2019), reviewing the flagged observations case by case in the experimental context. The variables on soils characteristics showed generally positive skewed distributions, in some cases multimodal, and tests showed evidence contrary to assumptions of spatial autocorrelation, homoscedasticity and normality in most cases (Supplementary Material: Table S1-S2; Fig. S1-S2).

Spearman's rank correlations (ρ) were calculated to explore linear associations between soil parameters across all single samples and within different land uses. We used the *adonis* function of the R package *vegan v2.5-7* (Oksanen et al., 2020) with a Euclidean dissimilarity matrix from our normalized soil data to perform PERMANOVA tests with 9999 permutations to analyse the multivariate homogeneity of group dispersions based on differences on soil parameters between land uses.

To compare the general effects of land use type on topsoil parameters, non-parametric Kruskal-Wallis tests were carried out in R. When significant ($p \leq 0.05$), we used Pairwise Wilcoxon Rank Sum tests with Benjamini & Hochberg correction to evaluate pairwise differences among land uses.

We used non-metric multidimensional scaling (NMDS) as a robust unconstrained ordination method to visualize patterns of top soil characteristics among all samples and within subsamples (intersected with different soil Orders) across different land uses using the Bray-Curtis dissimilarity matrix. The matrix was constructed with the *vegan* package to depict patterns of all soil parameters in two dimensions (Fig.3a, c-f) and for the dataset without soluble cations and micronutrients variables comparing subcategories within single land use types (i.e. within grasslands in 'undisturbed', 'partially grazed' and 'highly grazed' plots; timber plantations in '*Eucalyptus*' or '*Pinus*' plots; Fig. 3b).

3. Results

Nearly 75% of the sampled 101 plots intersected with the soil Group category 'Luvic Phaeozem' (WRB; IUSS Working Group, 2015). A quarter of our plots were distributed among four other Groups (i.e. Haplic Luvisols, Eutric Vertisols, Planosols, Lithic Leptisols; Fig. 1c). Half of our plots intersected with the Orders 'Argiudolls' or 'Argiudolls & Hapluderts' (Soil Survey Staff, 1999), and about a quarter with 'Hapludalfs & Hapludults' or 'Argiudolls, Hapludolls & Hapludalfs' (Fig. 1d). Our plots intersected with 32 different CONEAT categories (MGAP, 2021). The most frequent categories were '10.2', '9.3' and '2.12' (25% of our plots) and another quarter with '9.1', '8.8', '9.6' and 'G03.11' (Fig. 1e).

3.1 General characteristics of Uruguayan topsoils

The measured topsoil parameters vary widely across Uruguay, between the different land uses and classification to different soil orders (Table 1, Table A1-A4). The soil organic matter in topsoils ranged between 0.8 to 16 percent, and was highest in native forests and lowest in timber plantations. The mean of the C/N ratio was about 13, and lowest in crop soils. Phosphorous ranged between 43 to 1009 mg kg^{-1}. We also observed a high variability for the micro- and macronutrients and trace metals. The average cation-exchange capacity (CEC) was 12.41 cmol kg^{-1}, highest in topsoils of native forest, followed by crops and grasslands and timber plantations (Table 1, Table A1-A2).

For the whole data set, high correlation was found between P with SOC and Zn (ρ=0.82 and 0.76, respectively), and between Mg with Ca and Na (ρ=0.82 and 0.76, respectively; Fig. 2). Similar results were observed within particular land uses, although in native forests, a negative moderate correlation between Ca and Fe was observed (Fig. S3). There was high correlation between pH and Ca (ρ=0.89; Fig. 2). In native forests, we also found similar correlation with other soluble cations like Mg and K (Fig. S3). In park forests topsoils, there was a negative correlation between pH and As and Pb (ρ=-0.81 and ρ=-0.84, respectively). In highly grazed pastures and crops pH was highly correlated with Cr, and in crops also with As (ρ=0.93; Fig. S3). Among trace metals, Spearman correlation was moderate between As with Cr and Pb and high between Cr and Pb (Fig. 2). These correlation trends increased in grasslands, timber plantations and crops. We also found a high correlation between cadmium and Cr in crops (ρ=0.81). Phosphorus was highly correlated with Cr and As in pine plantations, while in soybean crops the correlation was between phosphorus with Cr and Pb (Fig. S3).

3.2 Topsoil characteristics clustered by land use

We found differences in multivariate distribution of samples according to the different land uses (i.e., grassland, timber plantations or native forests; Fig. 3a-f; Table S3-S8). All pairwise comparison of land uses showed significant differences with all variables (p=0.0001; Fig. 3a; Table S3). We analysed subcategories within a land use type using the dataset without soluble cations and micronutrients variables, only finding significant differences between *Eucalyptus* and *Pinus* stands (*p*=0.0001; Fig. 3b; Table S4) but not among different grassland subtypes. We also found differences analysing subsamples of different soil order classification to the different land uses (Fig. 3c-f; Table S5-S8). The samples intersected with 'Argiudolls' included all plots on crops, and we found significant differences in all pairwise comparisons of land uses except between grasslands and timber plantation (*p*=0.0004; Fig. 3c; Table S5). We further found differences between timber plantations and native forests at soils of the 'Argiudolls & Hapluderts' Orders (*p*=0.0009; Fig. 3d; Table S6) and between timber plantations and grasslands or native forests in soils of the 'Argiudolls, Hapudolls & Hapludalf' Orders (*p*=0.0284; Fig. 3e; Table S7). Results for samples in 'Hapludalfs & Hapludults' soils were similar to those obtained at country scale (*p*=0.0001; Fig. 3f; Table S8).

3.3 Differences in fertility proxies

We found significantly higher fertility proxies for SOC, P, Ca, Mg and Zn in topsoils from native forests compared to grasslands and timber plantations (Fig. 4b-e, h). Phosphorous was significantly higher in topsoils of grassland compared to those of timber plantations (Fig. 4c). Potassium was significantly higher in topsoils of native forests compared to timber plantations (Fig. 4f). At subcategories level, we found significantly higher amounts of K in partially used grasslands in comparison to samples from highly grazed pastures, and higher SOC (*p*=0.002), P (*p*=0.059), Na (*p*=0.043), K (*p*=0.012) and Zn (*p*=0.048) in *Eucalyptus* compared to *Pinus* plantations (Fig. S4). Among native forests, samples from riverine forests contain more Mg (*p*=0.023) and Na (*p*=0.023) in comparison to park forests.

Considering all samples within the Order ‘Argiudolls', C/N ratio in agricultural topsoils are lower compared to all other land uses. Phosphorus was higher in topsoils of native forests and crops compared to grasslands and timber plantations. Soil organic matter was highest in native forests (Fig. 5a-o; Table A3-A4).

3.4 Soil Acidification

Topsoil samples showed a markedly acid profile (median pH=4.66), with nearly 75% classified as 'extremely acidic' and 'very strongly acidic' (Fig. 6a). We found significant differences across land uses, with less acidic native forests, and lower pH in timber plantations (Fig. 6b). Comparing between land uses, we found more samples with neutral acidity in grasslands and more with higher acidity in timber plantations (Fig. 6b-c). In addition, we found lower pH in samples from *Pinus* compared to *Eucalyptus* stands (p=0.018). Results of analysis inside soil Orders showed similar variations observed at country scale, with timber plantations being more acid and native forest closer to neutral pH (Fig. 6d-f).

3.5 Trace metal accumulation across land uses

For As, Cd, Cr and Pb, we found significantly higher concentrations in topsoils originating from native forests compared to the grassland and timber plantation samples (Fig. 4l-o and 5l-o; Table A1-A4). At the same time, samples from *Eucalyptus* plantations had higher concentrations of both Cr ($p<0.005$) and Pb ($p<0.005$) than *Pinus* topsoils, while the same was observed for Cr ($p<0.001$) and Pb ($p<0.05$) in riverine forests compared to park forests (Fig. S5).

4. Discussion

The vicious circle between the wish to stop soil degradation and concurrent increases in land productivity to satisfy the increasing demand for food, fibres and energy has not been broken since green revolution. Socio-economic and conventional management practices that drive soil degradation have generated several traps, such as the ‘inputs trap' where a reduced yield per area is followed by higher fertilizer application or the ‘credit or poverty trap' where economic pressure forces the farmer to intensification (Gomiero, 2016). Caught in this loop, the soils of the temperate grasslands of Uruguay have suffered strong degradation from erosion, acidification, contamination, salinification and compaction (Zurbriggen et al., 2020). This is clearly reflected in the results of our topsoil survey, which also adds interesting insights from timber plantations, grasslands and native forests to an existing database consisting mainly of crops and pastures samples from 2002-2014, which demonstrated the loss of organic matter by 25% and an increasing loss of nutrients (Beretta-Blanco et al. 2019). We contribute deeper insights on fertility, acidification and trace metals accumulation in topsoils from a wide range of different land uses, which is, to our knowledge, unique for the region since the CONEAT classification (CONEAT Index, 1976).

4.1 Translocation of elements in topsoils within landscape

Our data demonstrate the high amounts of organic carbon, nutrients and trace metals in topsoil samples from riverine forests, suggesting transport of soil particles from the surrounding land uses (e.g., grasslands, crop or timber plantations) to the borders of rivers, streams and creeks. Therefore, we assume that organic carbon, nutrients and trace metals are displaced within the landscape and accumulate in the floodplains. Regional soil erosion models estimate the loss of 2-5 tons ha^{-1} $year^{-1}$ for a third of the country depending on precipitation, topography, soil erodibility and land management (Carrasco-Letelier and Beretta-Blanco 2017). One possible direct impact is the increasing eutrophication reported for larger local rivers, although the models used by these authors did not link Chlorophyll-a concentrations with agricultural land use (Beretta-Blanco and Carrasco-Letelier 2021 and replies). However, other vertical processes and differences across land uses can be hidden deeper in the soil profile, and have not been analysed in this study.

Organic matter content and the exchangeable cations are strongly reduced in topsoils of grasslands, timber plantations and crops compared to native forests (Fig. 4b, d-h and 5b, d-h). Of all the fertility proxies assessed, phosphorus in topsoil was most significantly affected by different land uses, being highest in native forests (Fig. 4c and 5c). The cation exchange capacity (CEC) was highest in native forests (more than 160 percent of the CEC in grasslands) and lowest in timber plantations, reaching only half of the CEC in grasslands (Table A1-A2). Lower average nutrient concentrations and corresponding CEC reported for two timber plantations in the East and one in the Norwest of Uruguay (Hernández et al., 2009 and 2016; Céspedes-Payret et al., 2012) may result from a combined effect of topsoil degradation in timber plantations due to management practices and to soil texture (Sandoval-Lopez et al., 2020). The trees' uptake and the general export of nutrients from fast growing timber plantations through harvesting is higher than the natural input into those systems (Merino et al., 2005). This effect is particularly relevant because timber plantations in Uruguay are on 'forestry priority soils', which are generally soils with low fertility, superficial to moderate depth and good drainage (OAS, 1994), so afforestation might reduce even more their soil fertility.

Comparing neighbouring sites of grassland and afforested grasslands, our data set shows no significant depletion of nutrients by timber plantations ($p=0.208$) but a slightly higher average of CEC at grasslands (Table A1). In general, plants are expected to compensate the extraction of cations from upper soil by the 'uplift of nutrient' from deeper horizons (Jobbagy and Jackson, 2004). This has been questioned for *Eucalyptus* plantations on sandy soils, where, for example, an increase of potassium in the topsoil was not observed compared to the neighbouring grassland (Céspedes-Payret et al., 2012). Phosphorous and cations decline with sand content, so physicochemical factors such as the percentage of sand and organic matter, influence soil fertility (Sandoval-Lopez et al., 2020). A study of *Eucalyptus* plantations in South-East Brazil did not show a significant depletion of nutrients and carbon; in contrast, carbon, potassium and calcium increased in the topsoil after twelve years over one or two harvest and fertilization cycles (McMahon et al., 2019). To sum up, the patterns observed in the various sites in our survey and the other studies indicate very complex interactions of numerous factors. Removal of nutrients by high-yield timber plantations may exceed the capacity of nutrient exchange or turnover of forest litter and wastes. Both stand management and environmental conditions (e.g., precipitation) influence nutrient and carbon stocks (McMahon et al., 2019; Sandoval-Lopez et al., 2020).

The C/N ratio in topsoils ranged within values reported for grassland and timber plantations of the region (see Berthrong et al., 2012). In contrast to that study, however, we observed no differences between grassland and timber plantations (Fig. 4a), but between topsoils of timber plantations, native forests and crops (Fig. 5a). Organic matter and nitrogen decrease in topsoils after grassland afforestation (Berthrong et al., 2009) and C/N ratio increases with plantation age and decreases with precipitation (Berthrong et al., 2012). The topsoils of timber plantations have, on average, lower N concentrations compared to other land uses (Table A2) and data in literature (Jobbagy and Jackson 2003).

Although in cropland, nutrients are regularly compensated for by increased application of fertilizers (Beretta-Blanco et al., 2019), timber plantations are usually fertilized only in the first year of planting (e.g., Binkley et al., 2017; Sandoval-Lopez et al., 2020). The CEC and the average ratio between $K^+(Ca^{2+} + Mg^{2+})^{-1/2}$ in crops and grasslands are in the ranges reported by Beretta-Blanco et al. (2019) with lower availability of potassium for crops. A shortage of potassium and calcium is also most likely for future timber plantations after harvest, especially if the logging process do not include bark stripping on site (Hernández et al., 2009) and due to short time spans between harvest and new planting in the same area (Sandoval-Lopez et al., 2020), but see contrasting results of McMahon et al. (2019). The extraction of nutrients and biomass due to grazing (Fernandez et al., 2017), timber (Hernández et al., 2009 and 2016) and crop production (Beretta-Blanco et al., 2019) lowers organic matter content and exchangeable cations, affects the physical, chemical and biological soil properties, and drives degradation.

4.2 Acidification in Uruguayan topsoils across land uses

A further dimension of the soil degradation directly linked to the cation extraction is the acidification of soils. This has been demonstrated for crops and pastures (Beretta-Blanco et al., 2019), and a study site with *Eucalyptus* (Céspedes-Payret et al., 2012) and is now broadly supported by our topsoil samples originating from a wide range of different land uses across Uruguay. The pH of our topsoil samples are mainly in the category of very strongly to extremely acidic and lowest in timber plantations (Fig. 6), which are below the means reported so far (Jobbagy and Jackson, 2003; Céspedes-Payret et al., 2012). Moreover, as our data on topsoil pH fall short compared to the values estimated by the Food and Agriculture Organization of the United Nations and the Intergovernmental Technical Panel on Soils (Caon and Vargas, 2017), acidification and the deterioration of topsoil quality continues. Acidification results from intensified land uses with nitrogen fertilization, with biological N fixation by the legumes both used in the improved pastures (Modernel et al., 2016), or with cation extraction by crop or timber harvest (Jobbagy and Jackson, 2003). So far, although forest soils tend to be more acidic than agricultural soils due to acid-neutralizing treatments in the latter (Baize and van Oort, 2014), we found no differences between topsoils of crops and native forests (Fig. 6d) as the high organic matter content in native forests buffers the process to a certain extent.

4.3 Riverine Forest soils as sink for trace metals

In general, the average concentration of Zn, Cu, Cd, Cr, As and Pb in our topsoils were within the expected ranges of samples from crops, pastures and grassland in the region (Lavado et al., 1998, 2004; Roca, 2015). Some samples, especially from orchards and crops, exceeded the background values of copper, cadmium and arsenic (Table A2; Roca, 2015). Little is known about the pedo-geochemical background of the Pampean soils (Roca, 2015): for example, high arsenic concentration in Uruguayan ground waters has been hypothesized to be due to quaternary ash deposits (Wu et al., 2021). In addition, the lateral heterogeneity of Pampean soils over short distances makes separating geochemical and anthropic signatures difficult

(Roca, 2015). However, that the main risk of soil contamination in the region is from the application of fertilizers and agrochemicals is uncontested (Kaushal et al., 2018).

To our knowledge, there has been no regional study of trace metals in the native riverine forests or timber plantations. Our work thus expands the evidence base for these land uses.. The topsoils of riverine forests accumulate more trace metals compared to those of timber plantations and crops (Fig. 4l-o). The higher amount of soil organic matter in riverine forests favours the retention of cations, including trace metals. Although the origin of potentially harmful elements in forest soils have been primarily attributed to atmospheric deposition (Baize and van Oort, 2014), atmospheric deposition only plays a major role in vicinity of urban or industrial development, and our data from rural sites suggests a different entry path from the surrounding land uses to the riverine forests. High acidification and low amounts of organic matter reduce the retention of trace metals in the soil of timber plantations, and elements leach out of the soil towards the water table (Baize and van Oort, 2014). The acidification strongly contributes to the overall mobility of base cations into the 'chemical cocktail of the Anthropocene' (Kaushal et al., 2018), including trace metals. We thus observe positive feedback in already impoverished soils with high acidity favouring cations solubility, in addition to the uptake by trees intensifies this effect. Timber plantations extract trace metals from soil and also accumulate it in the bark or leaves, so they have been used for phytoremediation (Li et al., 2020). This may explain the higher concentration of cadmium in grassland compared to timber plantations (Fig. 4m). Differences between *Eucalyptus* and *Pinus* stands may be related to different age classes, as the later may have extracted more lead and chromium from the soil due to their older stand age with rotation periods of about 20 years (Li et al., 2020).

4.4 Carbon storage in topsoils of *Eucalyptus* plantations?

Our study provides evidence that the loss of soil organic matter limits not only the productivity of the crops, but also potential carbon sequestration in the region (Beretta-Blanco et al., 2019).

Grassland conversion to cropland decreases soil organic carbon storage compared to adjacent native grasslands (Hernandez-Ramirez et al., 2021).

Afforestation of croplands has been also discussed as a carbon sequestration measure to proactively address and effectively mitigate ongoing climate change within a person's lifetime (Hernandez-Ramirez et al., 2021; Mayer et al. 2020). Although carbon stocks of four *Eucalyptus* stands in the Brazilian Cerrados increased but did not change in four *Eucalyptus* stands within the Atlantic Forest (McMahon et al., 2019), this may not be the case for short rotation of *Eucalyptus* plantation on Campos grasslands. In our topsoil samples originating from 28 different stands across Uruguay, organic matter is lowest in topsoils of timber plantations (Fig. 4b). Similar amounts have been reported for a timber plantation in the East of Uruguay (Céspedes-Payret et al., 2012), and in North-Eastern Argentine (Sandoval-Lopez et al., 2020). Our data therefore provide clear evidence that rather than contributing to carbon sequestration in the topsoil, the carbon release from the transformation of native grasslands to plantation with these fast-growing species has several adverse effects depending on precipitation and soil type (reviewed by Mayer et al., 2020).

Several trade-offs between carbon sequestration through afforestation and local water yield and soil fertility have been demonstrated, including nutrient and soil organic matter depletion, acidification, and biodiversity loss and corresponding challenges for landscape conservation (e.g., Jackson et al., 2005; Veldman et al., 2015; Friggens et al., 2020). Soil carbon changes dynamically during the first decade of afforestation. Remaining grassland carbon declines, while tree carbon gain starts (e.g. Paul et al., 2002). Although a net gain of carbon is expected when the new forests approached equilibrium after decades (Hernández-Ramirez et al., 2021), in contrast to long-lasting forests, *Eucalyptus* harvest in Uruguay takes place after less than a decade (appr. 7-10 years). Soil organic matter does not differ between *Eucalyptus* plantations and neighbouring grasslands (Appendix A; Table A1 and A2). Another study near our sites 12 and 13 also did not find a significant difference between the soil organic carbon of the upper soil (0-30cm) compared to afforested versus native grasslands (Hernández et al., 2016). A study near our sites 7, 8, 10 reported higher top soil carbon of grassland compared with afforested sites in their vicinity only if sand content is lower than sixty percent (Sandoval-Lopez et al., 2020). McMahon et al. (2019) identified a greater carbon gain under *Eucalyptus* stands compared to (potential) carbon losses in neighbouring degraded Cerrado grasslands. There is evidence that the retention of residues after harvest increases the carbon stock in soil (Mayer et al., 2020). Long term monitoring of carbon stocks is needed to verify the influences of management and environmental changes. A regional study on *Eucalyptus* plantations across different biomes in Brazil shows both decreases and no changes depending on precipitation in the dry season, clay content and on the initial stock of carbon in the soil (Cook et al., 2016)

The simplistic solution of huge tree plantations to compensate anthropogenic CO_2 emissions has been challenged in the last decade, and some crucial lessons learnt have been summarized (Di Sacco et al., 2020). Since grasslands are more sustainable carbon sinks compared to climate-vulnerable monocultures such as timber plantations (Dass et al. 2018), avoiding afforestation of previously non-forested lands is important. This is the case for *Eucalyptus* afforestation in the originally forestless grasslands of Uruguay: models on carbon sequestration and dynamics in Mollisols and Oxisols under South American grasslands estimated a higher carbon retention efficiency under grasslands compared to afforested sites, suggesting that silvopastoral systems are a potential solution for soil carbon sequestration in tropical soils (Berhongaray and Alvarez, 2019).

Connecting or expanding existing forest and using native species for plantings (Di Sacco et al., 2020) is also recommended (see also Alvaro et al. 2020, Ortiz et al. 2020). Our topsoil data indicates that carbon sequestration occurs mainly in the topsoils of native riverine forests that cover less than five percent of the Uruguayan territory. Consequently, the expansion of native forests and the use of native species in forestry project for long term establishment can reduce adverse effects of timber plantations. While there is preliminary evidence that N-fixing tree species can help increase local C stocks of afforestation, because of their potential for invasion, exotic N-fixers should be avoided (Mayer et al. 2020).

Recent studies indicated that novel techniques such as perennial grain cropping can help to turn around cropland degradation into beneficial cropland aggradation by using the advantage of perennial vegetation of conserving and even enhancing short term and long-term soil carbon storage and other ecosystems services (Kim et al. 2022).

4.5. Back to more conservative grassland management?

Our soil survey data shows strong soil degradation of Uruguayan black soils from erosion, acidification and contamination, and suggests a translocation of nutrients and organic matter across the landscape from grassland, timber and crop plantations to the riverine forests. The potential of grasslands as cropland reserve have been largely overestimated (Lambin et al., 2013), and they have already degraded during the last decades by inappropriate land management techniques (Jaurena et al., 2021, De Faccio et al., 2021), lack of mainstreaming soil conservative techniques (García-Préchac, 2004; Fernandez et al., 2017), decoupling of crop and livestock (De Faccio et al, 2021) plus climate change impacts with storm water events and drought, all of which trigger soil erosion (Wingeyer et al. 2015). From the although very limited point of view on topsoils, the concept of conserving 'old growth grasslands' with extensive use (Veldman et al. 2015) appears a more promising strategy to put the 'grasslands at the core' in the Campos region than the use intensification strategies envisioned by Jaurena et al. (2021).

5. Conclusions

The land use intensification in Uruguay associated with increasing inputs of energy, nutrients and pesticides leads to an overall loss of soil fertility and increasing toxicity related to acidification, salinization and trace metal contaminants. Our data demonstrate the high amounts of organic carbon, nutrients and trace metals in topsoil samples from riverine forests, suggesting transport of soil particles from the surrounding grasslands, crop or timber plantations to the borders of rivers, streams and creeks. Of all the fertility proxies assessed, phosphorus in topsoil was most significantly affected by different land uses, being highest in native forests. Cation exchange capacity was also highest in native forests and lowest in timber plantations, where only half that of grasslands was measured. Our study highlights that soil acidification is ongoing and probably also mobilizing trace metals and their accumulation in riverine forest topsoils.

Data availability

All data generated or analysed in this study are included in this article (and its supplementary materials). Further data are available from the corresponding authors upon reasonable request.

Author contributions

IS, LRR, ST and MB conceptualized the study, performed the soil sampling and data analysis. EZ advised the laboratory analyses. IS wrote the initial draft. LRR, ST, MB and EZ contributed to generating and reviewing the subsequent versions of the manuscript. IS received the funding for the study.

Competing interests

The authors declare that they have no conflict of interest

Acknowledgement

In alphabetical order, we thank Juan Barreneche, Lucia Gaucher, Sören Miehe, Nicolas Silvera and Matias Zarucki for their assistance in field work. We thank Manuel Garcia and Meica Valdivia for the pre-processing of samples. We thank the staff of the soil laboratory from the Department of Soil and Natural Resources, University of Concepción, Chile. We thank Diego de Panis for help with statistical analysis and visualization and valuable comments on a previous version of this manuscript. Thanks to Vera Krause, Serafina Bischoff, Sophia Reitzug, Rhea Rennert and Diego Nicolas Rojas for support with data analysis and plots and maps visualization. We also thank all landowners for access permission to establish our monitoring sites on their land, their hospitality and willingness to discuss land use goals concerning all dimension of sustainability. The study was funded by the German Federal Ministry of Education and Research (BMBF; 01LN1305A). Special thanks go to Amal Chatterjee for improving our English.

References

Alfaro, M., Dube, F. and Zagal, E.: The Influence of Overmature, Degraded Nothofagus Forests with Strong Anthropic Disturbance on the Quality of an Andisol and its Gradual Recovery with Silvopasture in Southwestern South America, in: Agroforestry for Degraded Landscapes, edited by: Dagar, J.C., Gupta, S.R., Teketay, D. Springer, Singapore. https://doi.org/10.1007/978-981-15-6807-7_3, 2020.

Baize, D. and van Oort, F.: Potentially Harmful Elements in Forest Soils, in: PHEs, Environment and Human Health, edited by: Bini, C. and Bech, J., Springer Netherlands, Dordrecht, 151–198, https://doi.org/10.1007/978-94-017-8965-3_4, 2014.

Beretta-Blanco A, Pérez O, Carrasco-Letelier L. Soil quality decrease over 13 years of agricultural production. Nutr Cycling Agroecosyst. 2019; 114:45-55.

Beretta-Blanco, A.; Carrasco-Letelier, L. Relevant factors in the eutrophication of the Uruguay River and the Río Negro. Sci. Total Environ. 2021, 761,143299; DOI: 10.1016/j.scitotenv.2020.143299

Berhongaray, G; Alvarez R. Soil carbon sequestration of Mollisols and Oxisols under grassland and tree plantations in South America - A review. Geoderma Regional 18 (2019) e00226.

Berthrong, S. T., Jobbágy, E. G., and Jackson, R. B.: A global meta-analysis of soil exchangeable cations, pH, carbon, and nitrogen with afforestation, Ecological Applications, 19, 2228–2241, https://doi.org/10.1890/08-1730.1, 2009.

Berthrong, S. T., Piñeiro, G., Jobbágy, E. G., and Jackson, R. B.: Soil C and N changes with afforestation of grasslands across gradients of precipitation and plantation age, 22, 76–86, https://doi.org/10.1890/10-2210.1, 2012.

Binkley, Dan et al. The interactions of climate, spacing and genetics on clonal *Eucalyptus* plantations across Brazil and Uruguay. Forest Ecology And Management. Amsterdam: Elsevier Science Bv, v. 405, p. 271-283, doi.org/10.1016/j.foreco.2017.09.050

Bonfante, A., Basile, A., and Bouma, J.: Targeting the soil quality and soil health concepts when aiming for the United Nations Sustainable Development Goals and the EU Green Deal, SOIL, 6, 453–466, https://doi.org/10.5194/soil-6-453-2020, 2020.

Borrelli, P., Robinson, D. A., Fleischer, L. R., Lugato, E., Ballabio, C., Alewell, C., Meusburger, K., Modugno, S., Schütt, B., Ferro, V., Bagarello, V., Oost, K. V., Montanarella, L., and Panagos, P.: An assessment of the global impact of 21st century land use change on soil erosion, Nat Commun, 8, 2013, https://doi.org/10.1038/s41467-017-02142-7, 2017.

Caon, L., & Vargas, R. (2017). Threats to soils: Global trends and perspectives. A Contribution from the Intergovernmental Technical Panel on Soils, Global Soil Partnership Food and Agriculture Organization of the United Nations. In G. Pierzynski & Brajendra (Eds.), Global land outlook working paper, 28 pp.

Carrasco-Letelier, L. and A. Beretta-Blanco. 2017. Soil erosion by water estimated for 99 Uruguayan basins. Cien. Inv. Agr. 44(2): 184-194.

Céspedes-Payret et al. 2012 Science of the Total Environment 438 (2012) 549–557. http://dx.doi.org/10.1016/j.scitotenv.2012.08.075

Cook, R. L., Binkley, D., and Stape, J. L.: Eucalyptus plantation effects on soil carbon after 20years and three rotations in Brazil, Forest Ecology and Management, 359, 92–98, https://doi.org/10.1016/j.foreco.2015.09.035, 2016.

Dass, P. et al 2018 Grasslands may be more reliable carbon sinks than forests in California. Environ. Res. Lett. 13 074027. https://doi.org/10.1088/1748-9326/aacb39

De Faccio Carvalho, P. C., Savian, J. V., Chiesa, T. D., Souza Filho, W. D., Terra, J. A., Pinto, P., Martins, A. P., Villarino, S., Da Trindade, J. K., De Albuquerque Nunes, P. A., and Piñeiro, G.: Land-use intensification trends in the Rio de la Plata region of South America: toward specialization or recoupling crop and livestock production, Front. Agr. Sci. Eng., 8, 97, https://doi.org/10.15302/J-FASE-2020380, 2021.

Di Sacco, A.; Hardwick, K.; Blakesley, D.; Brancalion, P.H.S.; Breman, E.; Cecilio Rebola, L.; Chomba, S.; Dixon, K.; Elliott, S.; Ruyonga, G.; Shaw, K.; Smith, P.; Smith, R.J.; Antonelli, A. Ten Golden Rules for Reforestation to Optimise Carbon Sequestration, Biodiversity Recovery and Livelihood Benefits. https://doi.org/10.1111/gcb.15498. 2020

Durán, A., Morrás, H., Studdert, G., and Liu, X.: Distribution, properties, land use and management of Mollisols in South America, Chin. Geogr. Sci., 21, 511, https://doi.org/10.1007/s11769-011-0491-z, 2011.

ESRI 2018. ArcGIS Desktop: Release 10.3. Redlands, CA: Environmental Systems Research Institute.
https://www.nrcs.usda.gov/Internet/FSE_DOCUMENTS/nrcs142p2_051232.pdf

FAO 2019. Land Use. Published online at OurWorldInData.org. Retrieved from: 'https://ourworldindata.org/land-use' [Online Resource]

Fernández, R., Frasier, I., Noellemeyer, E., and Quiroga, A.: Soil quality and productivity under zero tillage and grazing on Mollisols in Argentina – A long-term study, Geoderma Regional, 11, 44–52, https://doi.org/10.1016/j.geodrs.2017.09.002, 2017.

Fox, J., Weisberg, S. (2019) An R Companion to Applied Regression, Third Edition, Sage.

Friggens, N. L., Hester, A. J., Mitchell, R. J., Parker, T. C., Subke, J., and Wookey, P. A.: Tree planting in organic soils does not result in net carbon sequestration on decadal timescales, Glob Change Biol, 26, 5178–5188, https://doi.org/10.1111/gcb.15229, 2020.

García-Préchac, F., Ernst, O., Siri-Prieto, G., and Terra, J. A.: Integrating no-till into crop–pasture rotations in Uruguay, Soil and Tillage Research, 77, 1–13, https://doi.org/10.1016/j.still.2003.12.002, 2004.

Gomiero, Tiziano. 2016. "Soil Degradation, Land Scarcity and Food Security: Reviewing a Complex Challenge" Sustainability 8, no. 3: 281. https://doi.org/10.3390/su8030281

Hernández J, del Pino A, Salvo L, Arrarte G. Nutrient export and harvest residue decomposition patterns of a *Eucalyptus dunnii* Maiden plantation in temperate climate of Uruguay. Forest Ecology Management 2009; 258(2): 92-99. 10.1016/j.foreco.2009.03.050

Hernández, J., del Pino, A., Vance, E. D., Califra, Á., Del Giorgio, F., Martínez, L., and González-Barrios, P.: *Eucalyptus* and *Pinus* stand density effects on soil carbon sequestration, Forest Ecology and Management, 368, 28–38, https://doi.org/10.1016/j.foreco.2016.03.007, 2016.

Hernandez-Ramirez, G., Sauer, T. J., Chendev, Y. G., and Gennadiev, A. N.: Nonlinear turnover rates of soil carbon following cultivation of native grasslands and subsequent afforestation of croplands, SOIL, 7, 415–431, https://doi.org/10.5194/soil-7-415-2021, 2021.

IUSS Working Group WRB. 2015. World Reference Base for Soil Resources 2014, update 2015 International soil classification system for naming soils and creating legends for soil maps. World Soil Resources Reports No. 106. FAO, Rome. https://www.fao.org/3/i3794en/I3794en.pdf

Jackson et al 2005 Trading Water for Carbon with Biological Carbon Sequestration. Science 310, 1944 (2005); DOI: 10.1126/science.1119282

Jackson, R.: Soil nutrient stocks are maintained over multiple rotations in Brazilian Eucalyptus plantations | Elsevier Enhanced Reader, 13, n.d.

Jaurena M, Durante M, Devincenzi T, Savian JV, Bendersky D, Moojen FG, Pereira M, Soca P, Quadros FLF, Pizzio R, Nabinger C, Carvalho PCF and Lattanzi FA (2021) Native Grasslands at the Core: A New Paradigm of Intensification for the Campos of Southern South America to Increase Economic and Environmental Sustainability. Front. Sustain. Food Syst. 5:547834. doi: 10.3389/fsufs.2021.547834

Jobbagy, E. G., and R. B. Jackson. 2003. Patterns and mechanisms of soil acidification in the conversion of grasslands to forests. Biogeochemistry 64:205–229.

Jobbagy, E. G., and R. B. Jackson. 2004. The uplift of soil nutrients by plants: biogeochemical consequences across scales. Ecology 85:2380–2389.

Kaushal, S. S., Gold, A. J., Bernal, S., Johnson, T. A. N., Addy, K., Burgin, A., Burns, D. A., Coble, A. A., Hood, E., Lu, Y., Mayer, P., Minor, E. C., Schroth, A. W., Vidon, P., Wilson, H., Xenopoulos, M. A., Doody, T., Galella, J. G., Goodling, P., Haviland, K., Haq, S., Wessel, B., Wood, K. L., Jaworski, N., and Belt, K. T.: Watershed ‘chemical cocktails’: forming novel elemental combinations in Anthropocene fresh waters, Biogeochemistry, 141, 281–305, https://doi.org/10.1007/s10533-018-0502-6, 2018.

Kellogg, C. E. "Soil survey division staff: soil survey manual. Chapter 3" United States Department of Agriculture, Washington (1993).

Kim, K., Daly, EJ, M. Gorzelak, Hernandez-Ramirez, G. 2022. Soil organic matter pools response to perennial grain cropping and nitrogen fertilizer. Soil and Tillage Research. 220, 105376 https://doi.org/10.1016/j.still.2022.105376

Lambin, E.F.; Gibbs, H.K.; Ferreira, L.; Grau, R.; Mayaux, P.; Meyfroidt, P.; Morton, D.C.; Rudel, T.K.; Gasparri, I.; Munger, J. Estimating the world’s potentially available cropland using a bottom-up approach. Glob. Environ. Chang. 2013, 23, 892–901.

Lanfranco, B. and Sapriza, G.: El índice CONEAT como medida de productividad y valor de la tierra., 57, n.d.

Lavado, R.S., Rodrírguez, M.B., Scheiner, S.D., Taboada, M.A., Rubio, G., Alvarez, R., Alconada, M., Zubillaga, M.S., 1998. Heavy metals in soils of Argentina: comparison between urban and agricultural soils. Communi- cations in Soil Science and Plant Analysis 29, 1913–1917.

Lavado, R.S., Zubillaga, M.S., Alvarez, R., Taboada, M.A., 2004. Baseline levels of potentially toxic elements in pampas soils. Soil and Sediment Contamination: An International Journal 13, 329–339.

Li, H., Jiang, L., You, C., Tan, B., and Yang, W.: Dynamics of heavy metal uptake and soil heavy metal stocks across a series of Masson pine plantations, Journal of Cleaner Production, 269, 122395, https://doi.org/10.1016/j.jclepro.2020.122395, 2020.

Liu, X., Lee Burras, C., Kravchenko, Y. S., Durán, A., Huffman, T., Morras, H., Studdert, G., Zhang, X., Cruse, R. M., and Yuan, X.: Overview of Mollisols in the world: Distribution, land use and management, Can. J. Soil. Sci., 92, 383–402, https://doi.org/10.4141/cjss2010-058, 2012.

Mayer M., Prescott C.E., Abaker W.E.A., Augusto L., Cécillon L., Ferreira G.W.D., James J., Jandl R., Katzensteiner K., Laclau J.-P., Laganière J., Nouvellon Y., Paré D., Stanturf J.A., Vanguelova E.I., Vesterdal L. Influence of forest management activities on soil SOC stocks: A knowledge synthesis. Forest Ecology and Management, Volume 466, 2020

McMahon, D. E., Vergütz, L., Valadares, S. V., Silva, I. R. da, and Jackson, R. B.: Soil nutrient stocks are maintained over multiple rotations in Brazilian Eucalyptus plantations, Forest Ecology and Management, 448, 364–375, https://doi.org/10.1016/j.foreco.2019.06.027, 2019.

Merino, A., Balboa, M. A., Rodríguez Soalleiro, R., and González, J. G. Á.: Nutrient exports under different harvesting regimes in fast-growing forest plantations in southern Europe, Forest Ecology and Management, 207, 325–339, https://doi.org/10.1016/j.foreco.2004.10.074, 2005.

MGAP (Ministerio de Ganadería, Agricultura y Pesca) 2020. CONEAT. Online: https://www.gub.uy/ministerio-ganaderia-agricultura-pesca/politicas-y-gestion/coneat

Modernel, P., Rossing, W. A. H., Corbeels, M., Dogliotti, S., Picasso, V., and Tittonell, P.: Land use change and ecosystem service provision in Pampas and Campos grasslands of southern South America, Environ. Res. Lett., 11, 113002, https://doi.org/10.1088/1748-9326/11/11/113002, 2016.

OAS (Organization of American States, Department of Sustainable Development) 1994. Uruguay - Proyecto Regional de Alternativas para la Inversión Forestal. Available online: http://www.oas.org/dsd/publications/unit/oea20s/oea20s.pdf.

Oksanen, J., Blanchet, F. G., Friendly, M., Kindt, R., Legendre, P., McGlinn, D., Minchin, P. R., O'Hara, R. B., Simpson, G. L., Solymos, P., Stevens, M. H. H., Szoecs, E., and Wagner, H.: vegan: Community Ecology Package, 2020.

Ortiz, J., Dube, F., Neira, P., Panichini, M., Stolpe, N.B., Zagal, E. and Martínez-Hernández, P.A. Soil Quality Changes within a (*Nothofagus obliqua*) Forest Under Silvopastoral Management in the Andes Mountain Range, South Central Chile. Sustainability; 12(17):6815. https://doi.org/10.3390/su12176815, 2020.

Paul, K. I., Polglase, P. J., Nyakuengama, J. G., and Khanna, P. K.: Change in soil carbon following afforestation, Forest Ecology and Management, 168, 241–257, https://doi.org/10.1016/S0378-1127(01)00740-X, 2002.

Piñeiro, D. E.: Land grabbing: concentration and "foreignisation" of land in Uruguay, Canadian Journal of Development Studies / Revue canadienne d'études du développement, 33, 471–489, https://doi.org/10.1080/02255189.2012.746216, 2012.

R Core Team: R: A Language and Environment for Statistical Computing, R Foundation for Statistical Computing, Vienna, Austria, available at: https://www.R-project.org/, last access: 2 March 2021.

Richter, D. D.: Game Changer in Soil Science. The Anthropocene in soil science and pedology., J. Plant Nutr. Soil Sci., 183, 5–11, https://doi.org/10.1002/jpln.201900320, 2020.

Roca, N.: Heavy metal background levels in rural soils: a case study in Pampean soils (Argentina), 10, 2015.

Sandoval López, D.M., Arturi, M.F., Goya, J.F. et al. *Eucalyptus grandis* plantations: effects of management on soil carbon, nutrient contents and yields. J. For. Res. 31, 601–611 (2020). https://doi.org/10.1007/s11676-018-0850-z

Soil Survey Staff. 1999. Soil taxonomy: A basic system of soil classification for making and interpreting soil surveys. 2nd edition. https://www.nrcs.usda.gov/Internet/FSE_DOCUMENTS/nrcs142p2_051232.pdf

Veldman, J. W., Overbeck, G. E., Negreiros, D., Mahy, G., Le Stradic, S., Fernandes, G. W., Durigan, G., Buisson, E., Putz, F. E., and Bond, W. J.: Where Tree Planting and Forest Expansion are Bad for Biodiversity and Ecosystem Services, BioScience, 65, 1011–1018, https://doi.org/10.1093/biosci/biv118, 2015.

Waters, C. N., Zalasiewicz, J., Summerhayes, C., Barnosky, A. D., Poirier, C., Ga uszka, A., Cearreta, A., Edgeworth, M., Ellis, E. C., Ellis, M., Jeandel, C., Leinfelder, R., McNeill, J. R., Richter, D. d., Steffen, W., Syvitski, J., Vidas, D., Wagreich, M., Williams, M., Zhisheng, A., Grinevald, J., Odada, E., Oreskes, N., and Wolfe, A. P.: The Anthropocene is functionally and stratigraphically distinct from the Holocene, Science, 351, aad2622–aad2622, https://doi.org/10.1126/science.aad2622, 2016.

Wingeyer, A.B.; Amado, T.J.C.; Pérez-Bidegain, M.; Studdert, G.A.; Perdomo Varela, C.H.; Garcia, F.O.; Karlen, D.L. Soil Quality Impacts of Current South American Agricultural Practices. Sustainability 2015, 7, 2213–2242.

Wu, R., Alvareda, E., Polya, D., Blanco, G., and Gamazo, P.: Distribution of Groundwater Arsenic in Uruguay Using Hybrid Machine Learning and Expert System Approaches, Water, 13, 527, https://doi.org/10.3390/w13040527, 2021.

Zurbriggen, C., González-Lago, M., Baraibar, M., Baethgen, W., Mazzeo, N. and Sierra, M., 2020. Experimentation in the Design of Public Policies: The Uruguayan Soils Conservation Plans. Iberoamericana – Nordic Journal of Latin American and Caribbean Studies, 49(1), pp.52–62. DOI: http://doi.org/10.16993/iberoamericana.459

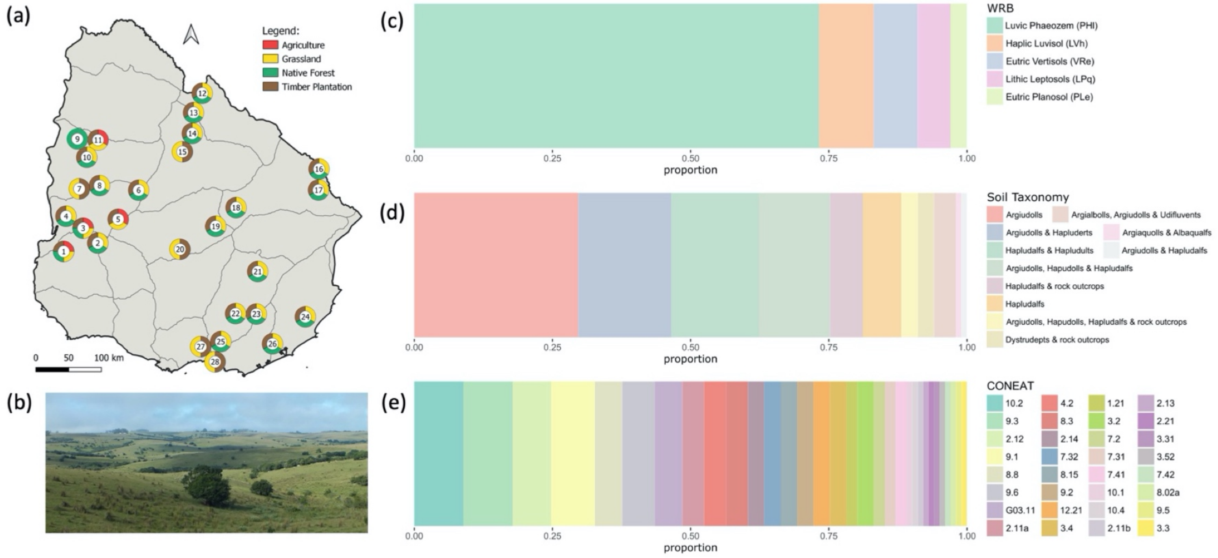

Figure 1: Study design and sampling. (a) Location of the 28 monitoring sites across Uruguay including land use types sampled (grassland (b); timber plantations, native forest and agricultural land. Proportion of plots with particular category of soil classification according to the World Reference Base (WRB, c), Soil Taxonomy (d) and CONEAT (Comisión Nacional de Estudio Agronómico de la Tierra, e). Photo: RuralFutures.

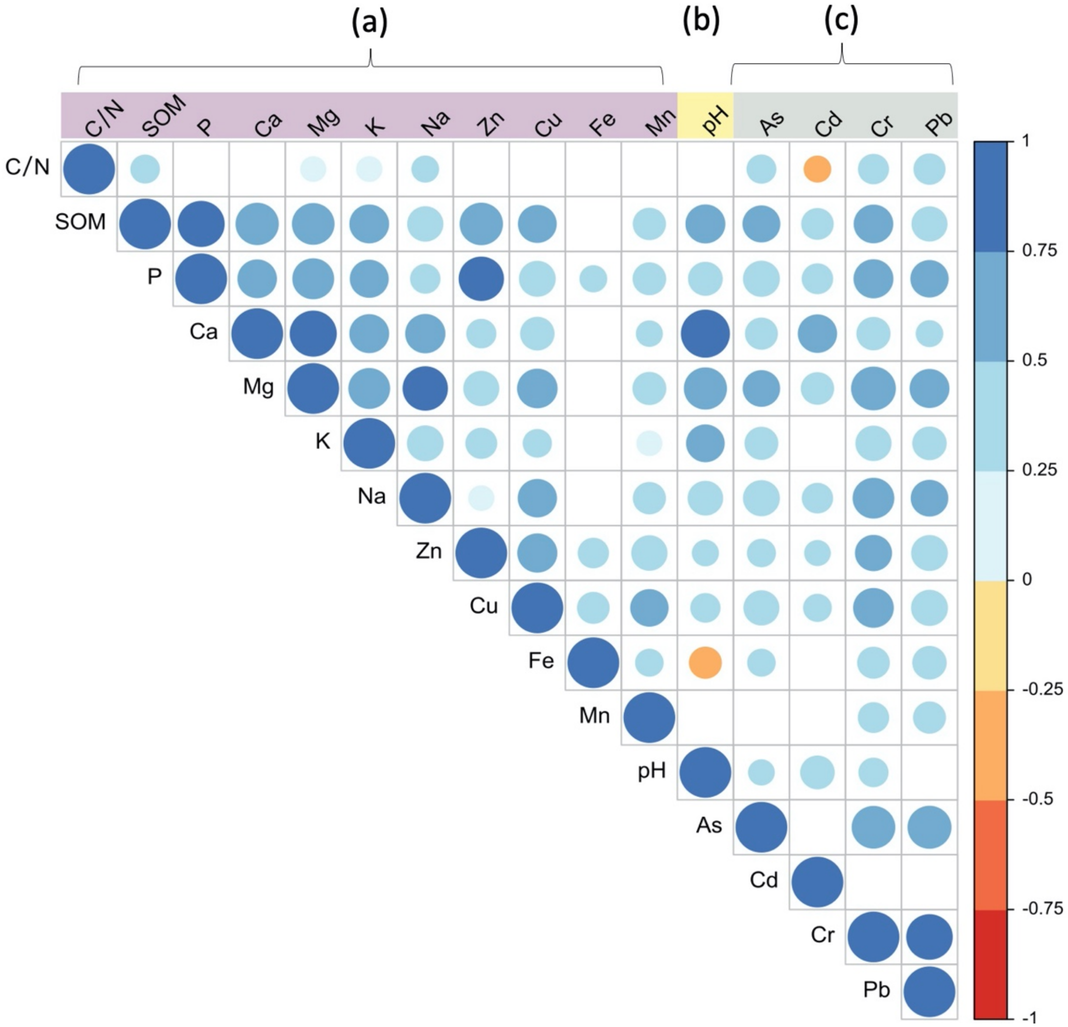

Figure 2. Spearman's rank correlations for parameters of topsoils (n=80) regarding (a) fertility proxies, (b) acidity and (c) trace metals. Colour intensity and the size of the circle are proportional to the correlation coefficients (ρ). Empty slots show correlations with $p>0.05$.

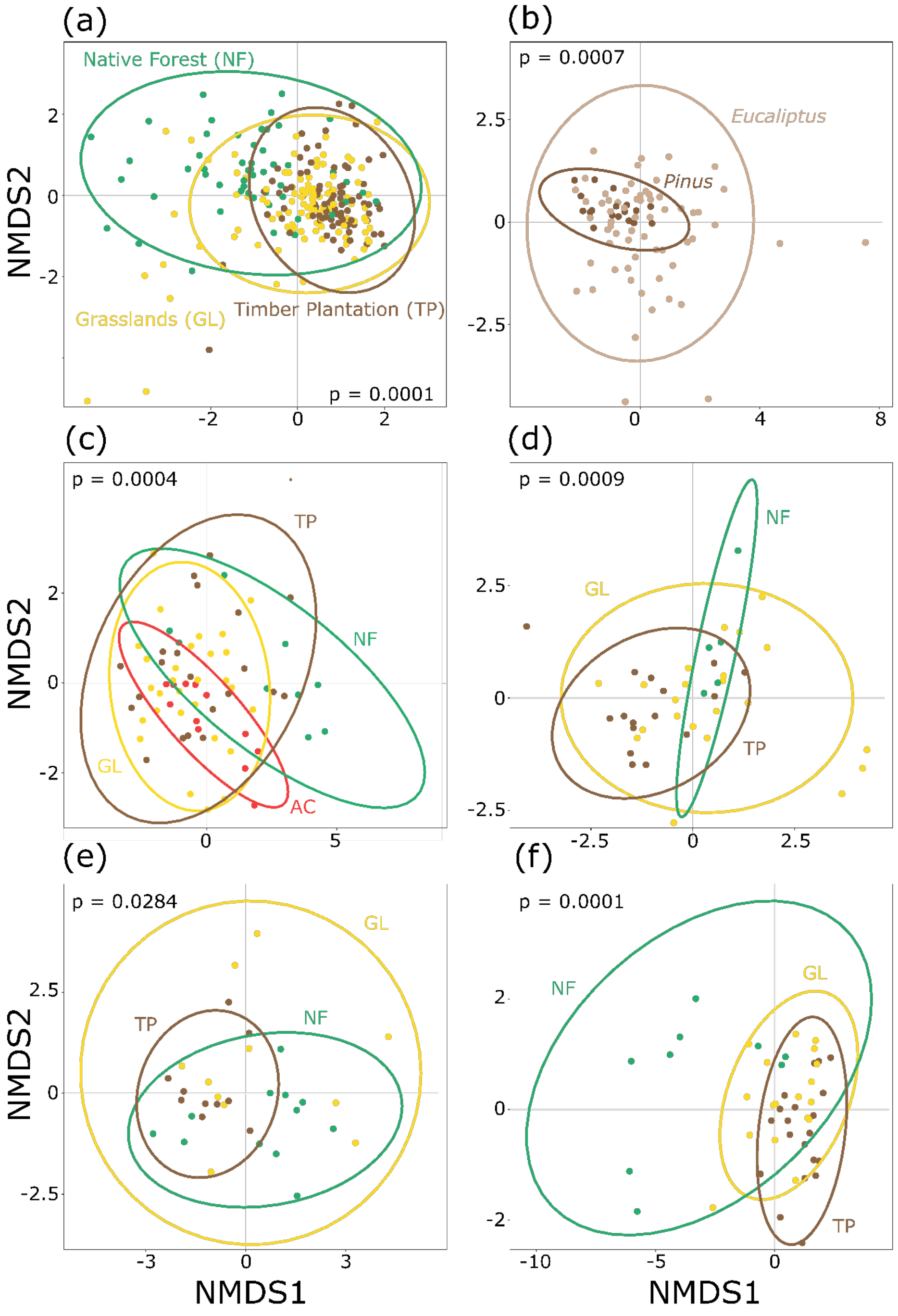
(a)
Native Forest (NF)
Grasslands (GL)
Timber Plantation (TP)
p = 0.0001
(b)
p = 0.0007
Eucaliptus
Pinus
(c)
p = 0.0004
TP
NF
GL
AC
(d)
p = 0.0009
NF
GL
TP
(e)
p = 0.0284
GL
TP
NF
(f)
p = 0.0001
NF
GL
TP
NMDS1
NMDS2
-2
0
2
2.5
-2.5
4
8
5
-3
3
-10
-5

Figure 3. Non-metric multidimensional scaling showing significant clustering differences among samples from (a) grassland (GL), timber plantations (TP) and native forests (NF); (b) among samples from *Pinus* and *Eucalyptus* plantations; (c) among land uses (including agriculture AC) in Argiudolls; (d) among land uses in Argiudolls & Hapluderts; (e) among land uses in Argiudolls, Hapudolls & Hapludalf and (f) among land uses in Hapludalfs & Hapludults.

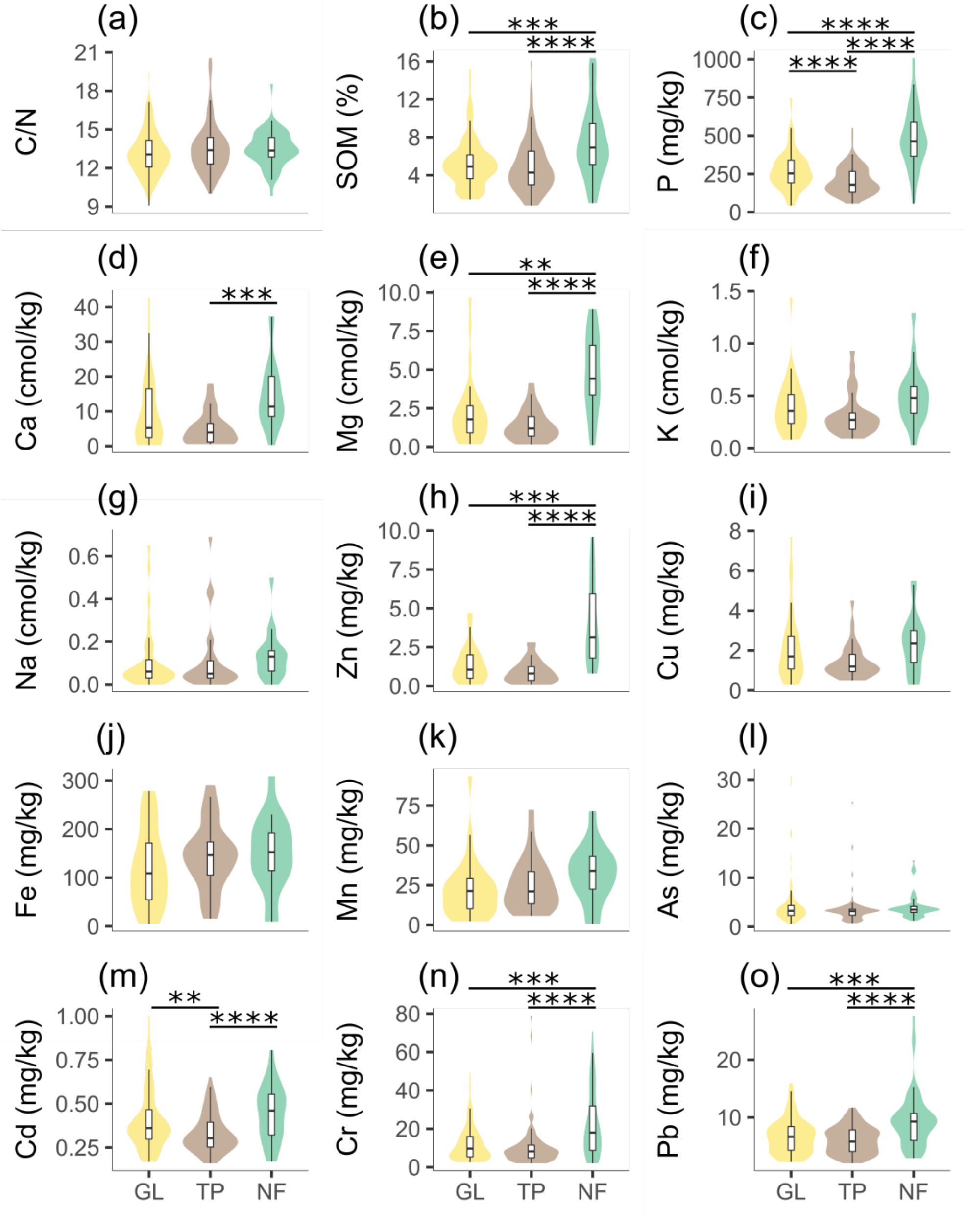
(a)
C/N
21
18
15
12
9
(b)
***
****
SOM (%)
16
12
8
4
(c)
****
****
****
P (mg/kg)
1000
750
500
250
0
(d)
***
Ca (cmol/kg)
40
30
20
10
0
(e)
**
****
Mg (cmol/kg)
10.0
7.5
5.0
2.5
0.0
(f)
K (cmol/kg)
1.5
1.0
0.5
0.0
(g)
Na (cmol/kg)
0.6
0.4
0.2
0.0
(h)
***
****
Zn (mg/kg)
10.0
7.5
5.0
2.5
0.0
(i)
Cu (mg/kg)
8
6
4
2
0
(j)
Fe (mg/kg)
300
200
100
0
(k)
Mn (mg/kg)
75
50
25
0
(l)
As (mg/kg)
30
20
10
0
(m)
**
****
Cd (mg/kg)
1.00
0.75
0.50
0.25
GL
TP
NF
(n)
***
****
Cr (mg/kg)
80
60
40
20
0
GL
TP
NF
(o)
***
****
Pb (mg/kg)
20
10
GL
TP
NF

Figure 4. Violin box plots for significant Kruskal-Wallis Tests across evaluated land uses (i.e. grassland (GL), timber plantation (TP) and native forest (NF)) for each variable. Significance in posterior Wilcoxon pairwise comparisons is depicted as $p<0.05$ (*), $p<0.01$ (**), $p<0.001$ (***), $p<0.0001$ (****).

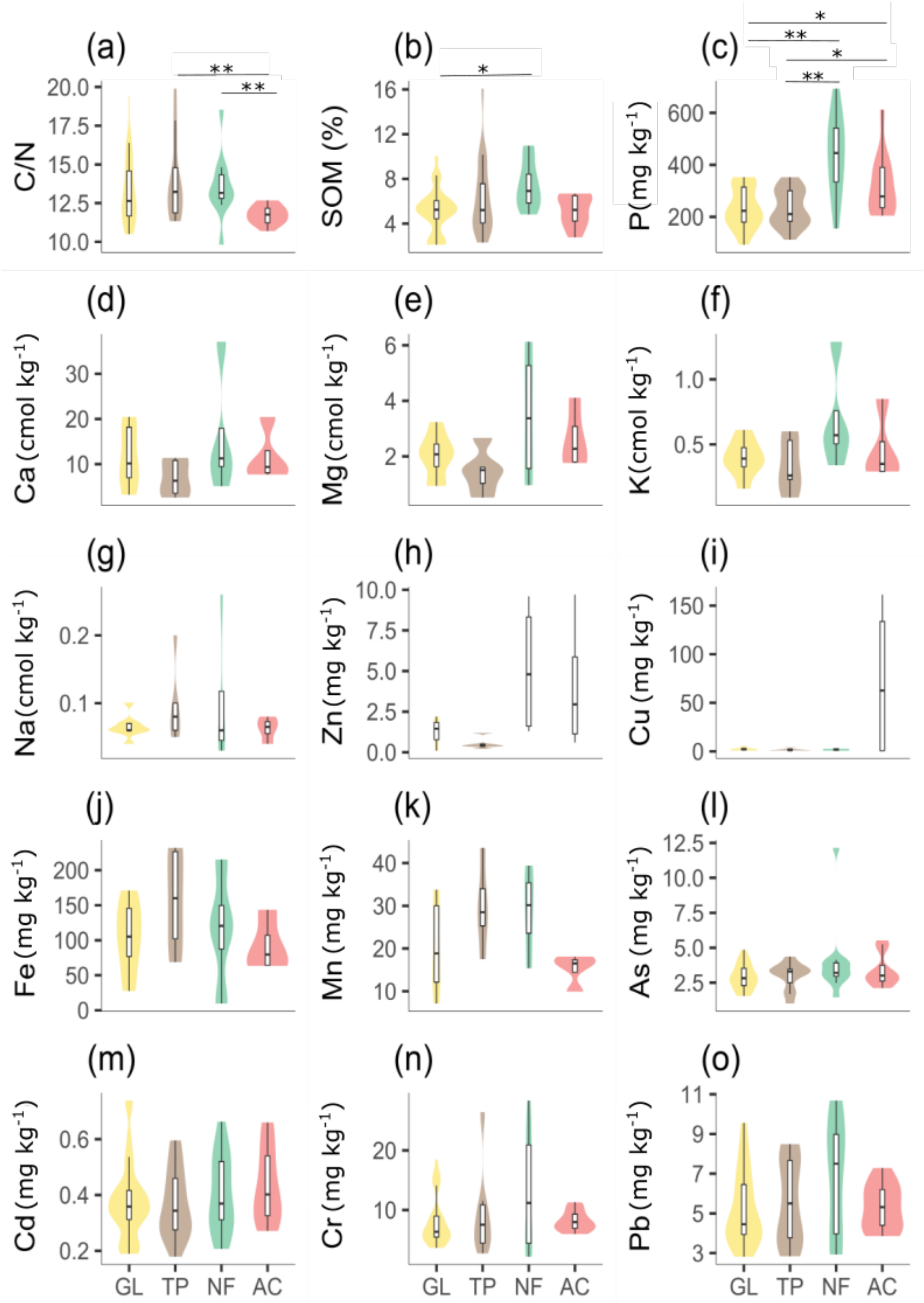
(a)
C/N
20.0
17.5
15.0
12.5
10.0
**
**
(b)
SOM (%)
16
12
8
4
*
(c)
P(mg kg^{-1})
600
400
200
*
**
*
**
(d)
Ca (cmol kg^{-1})
30
20
10
(e)
Mg (cmol kg^{-1})
6
4
2
(f)
K (cmol kg^{-1})
1.0
0.5
(g)
Na (cmol kg^{-1})
0.2
0.1
(h)
Zn (mg kg^{-1})
10.0
7.5
5.0
2.5
0.0
(i)
Cu (mg kg^{-1})
150
100
50
0
(j)
Fe (mg kg^{-1})
200
150
100
50
0
(k)
Mn (mg kg^{-1})
40
30
20
10
(l)
As (mg kg^{-1})
12.5
10.0
7.5
5.0
2.5
(m)
Cd (mg kg^{-1})
0.6
0.4
0.2
(n)
Cr (mg kg^{-1})
20
10
(o)
Pb (mg kg^{-1})
11
9
7
5
3
GL
TP
NF
AC
GL
TP
NF
AC
GL
TP
NF
AC

Figure 5. Violin box plots for significant Kruskal-Wallis Tests in 'Argiudolls' Soil Taxonomy category for fertility variables across available land uses (GL: Grassland, TP: Timber plantation, NF: Native forest, AC: Agriculture). Significance in posterior Wilcoxon pairwise comparisons is depicted as p<0.05 (*), p<0.01 (**), p<0.001 (***), p<0.0001 (****).

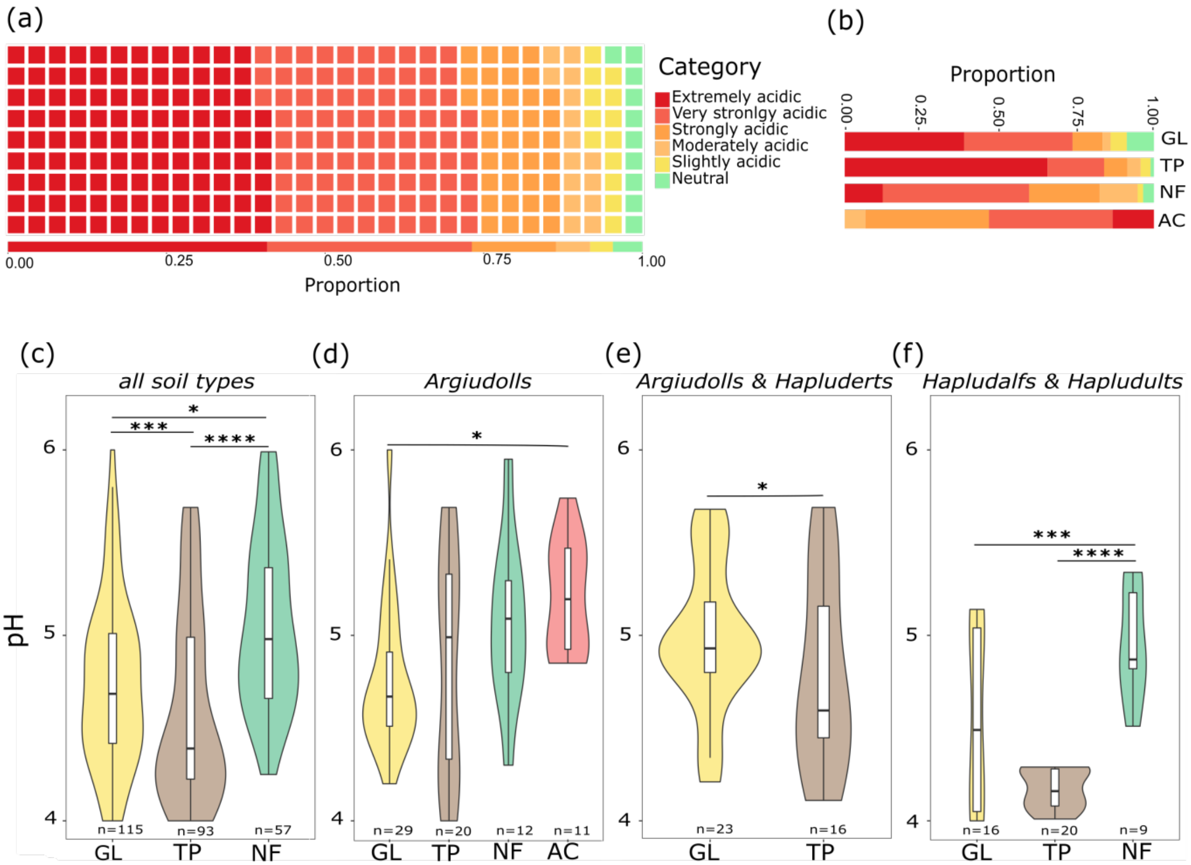

Figure 6. Acidity of all topsoils samples (a) and (b) according to different land use (GL: Grassland, TP: Timber plantation, NF: Native forest, AC: Agriculture). Proportion of samples with a given acidic category (Kellogg, 1993). pH of topsoil given as violin box plots for

significant Kruskal-Wallis Tests across different land uses independent from soil Order (c) and separated per soil Orders (d-f). Sample number (n) is given. Significance in posterior Wilcoxon pairwise comparisons is depicted as $p<0.05$ (*), $p<0.01$ (**), $p<0.001$ (***), $p<0.0001$ (****).

Table 1. General characteristics of Uruguayan topsoils: descriptive statistics for the parameters of all single samples (n) across different land uses and classification to different soil types (for details on different land uses i.e. grassland, timber plantations, native forests and crops see Appendix A1-A2 and on different soil types see Appendix A3-4). SD = standard derivation.

Variable	n	Mean	SD	Minimum	Maximum
Soil moisture (%)	280	18.9	10.8	1.34	51.5
N_total (%)	280	0.2	0.14	0.04	1.2
C_total (%)	280	3.0	2.3	0.4	25.7
C/N ratio	279	13.3	1.8	9.1	20.6
SOC (%)	267	5.6	3.0	0.8	16.4
P (mg kg^{-1})	278	295	166	43	1008
Ca (cmol kg^{-1})	82	9.4	8.9	0.3	42.7
Mg (cmol kg^{-1})	82	2.5	2.2	0.1	9.7
K (cmol kg^{-1})	82	0.4	0.25	0.03	1.44
Na (cmol kg^{-1})	82	0.1	0.14	0	0.69
Zn (mg kg^{-1})	82	1.9	2.1	0.1	9.7
Cu (mg kg^{-1})	82	5.4	22.1	0.3	161.2
Fe (mg kg^{-1})	82	134	75	5	309
Mn (mg kg^{-1})	81	25	17	0.7	93
pH	279	4.8	0.8	3.6	7.34
As (mg kg^{-1})	279	3.9	3.6	0.6	30.7
Cd (mg kg^{-1})	274	0.4	0.2	0.2	1
Cr (mg kg^{-1})	280	13.6	12.6	2	78.9
Pb (mg kg^{-1})	280	7	3.4	2	27.6
CEC (cmol kg^{-1})	82	12.4	10.7	0.5	50.1
$K^+/(Ca^{2+}+Mg^{2+})$	82	0.07	0.08	0.01	0.43
$K^+/(Ca^{2+}+Mg^{2+}+Na^+)$	82	0.07	0.07	0.01	0.42

Appendices

Appendix A

Table A1. Descriptive statistics of topsoil variables for all single soil types at grassland and native forests. Number of samples (n), mean, standard deviation (SD), minimum (Min) and maximum (Max) of each variable are given.

	Grassland					Native forests				
Variable (unit)	n	Mean	SD	Min	Max	n	Mean	SD	Min	Max
Soil moisture (%)	115	20.4	11.5	1.5	51.5	57	25.8	9.3	6.3	50.3
N_total (%)	115	0.22	0.13	0.06	0.80	57	0.30	0.19	0.04	1.20
C_total (%)	115	2.9	1.9	0.7	10.8	57	4.2	3.6	0.5	25.7
C/N ratio	115	13.2	1.7	9.1	19.4	56	13.5	1.4	9.8	18.5
SOC (%)	109	5.3	2.6	1.4	15.2	53	7.5	3.8	1.1	16.4
P (mg/kg)	114	272	124	43	749	57	484	194	56	1009
Ca (cmol kg^{-1})	32	9.5	10.0	0.3	42.7	18	15.1	10.0	0.3	37.3
Mg (cmol kg^{-1})	32	2.2	2.1	0.18	9.7	18	4.8	2.6	0.12	8.9
K (cmol kg^{-1})	32	0.40	0.26	0.08	1.44	18	0.50	0.28	0.03	1.29
Na (cmol kg^{-1})	32	0.12	0.15	0.00	0.65	18	0.14	0.12	0.00	0.50
Zn (mg kg^{-1})	32	1.4	1.2	0.10	4.7	18	4.0	2.7	0.80	9.60
Cu (mg kg^{-1})	32	2.2	1.8	0.30	7.7	18	2.4	1.5	0.30	5.50
Fe (mg kg^{-1})	32	121	81	5	279	18	149	76	10	309
Mn (mg kg^{-1})	32	22	18	2	93	17	33	18	1	72
pH	114	4.9	0.9	3.6	7.3	57	5.1	0.7	3.6	7.2
As (mg kg^{-1})	115	4.3	4.5	0.6	30.7	56	4.1	2.5	1.2	13.6
Cd (mg kg^{-1})	111	0.41	0.17	0.17	1.00	57	0.45	0.15	0.17	0.81
Cr (mg kg^{-1})	115	12.3	9.2	2.6	49.5	57	22.8	17.3	2.0	70.7
Pb (mg kg^{-1})	115	6.8	3.0	2.3	15.9	57	9.5	4.8	2.9	27.7
CEC (cmol/ kg^{-1})	32	12.2	11.5	0.8	50.1	18	20.4	11.4	0.5	44.6
$K^+/(Ca^{2+}+Mg^{2+})$	32	0.08	0.10	0.01	0.43	18	0.03	0.02	0.01	0.07
$K^+/(Ca^{2+}+Mg^{2+}+Na^+)$	32	0.08	0.10	0.01	0.42	18	0.03	0.02	0.01	0.07

Table A2. Descriptive statistics for of topsoil variables for all single soil types at timber plantations and crops. Number of samples (n), mean, standard deviation (SD), minimum (Min) and maximum (Max) of each variable are given.

	Timber plantations					Crops				
Variable (unit)	n	Mean	SD	Min	Max	n	Mean	SD	Min	Max
Soil moisture (%)	93	13.2	7.0	1.7	38.0	15	18.1	12.0	1.3	50.5
N_total (%)	93	0.18	0.10	0.04	0.61	15	0.22	0.06	0.13	0.29
C_total (%)	93	2.5	1.5	0.4	8.4	15	2.6	0.7	1.4	3.3
C/N ratio	93	13.6	2.1	10.0	20.6	15	11.7	0.7	10.5	12.8
SOC (%)	90	4.9	2.8	0.8	16.1	15	5.2	1.3	2.8	6.7
P (mg kg^{-1})	93	206	91	56	551	14	310	120	142	613
Ca (cmol kg^{-1})	27	5.2	4.8	0.6	18	5	10.8	5.5	7.2	20.4
Mg (cmol kg^{-1})	27	1.4	1.0	0.16	4.1	5	2.5	1.0	1.76	4.1
K (cmol kg^{-1})	27	0.32	0.21	0.09	0.93	5	0.46	0.23	0.29	0.85
Na (cmol kg^{-1})	27	0.11	0.16	0.00	0.69	5	0.06	0.02	0.03	0.08
Zn (mg kg^{-1})	27	0.94	0.74	0.10	2.8	5	3.7	3.7	0.60	9.7
Cu (mg kg^{-1})	27	1.5	0.9	0.5	4.5	5	57.7	78.8	0.8	161.0
Fe (mg kg^{-1})	27	143	69	16	290	5	115	62	63	210
Mn (mg kg^{-1})	27	25	16	6	72	5	17	5	10	23
pH	93	4.5	0.7	3.6	6.8	15	5.1	0.4	4.2	5.7
As (mg kg^{-1})	93	3.4	3.0	0.7	25.5	15	3.1	1.2	1.1	5.5
Cd (mg kg^{-1})	91	0.33	0.11	0.16	0.65	15	0.41	0.14	0.21	0.66
Cr (mg kg^{-1})	93	10.6	11.3	2.0	78.9	15	7.8	1.8	5.0	11.3
Pb (mg kg^{-1})	93	6.1	2.3	2.0	11.7	15	5.1	1.2	3.4	7.3
CEC (cmol kg^{-1})	27	7.0	5.6	1.0	20.6	5	13.8	6.5	9.5	25.0
$K^+/(Ca^{2+}+Mg^{2+})$	27	0.08	0.07	0.02	0.25	5	0.04	0.03	0.02	0.09
$K^+/(Ca^{2+}+Mg^{2+}+Na^+)$	27	0.08	0.07	0.02	0.25	5	0.04	0.03	0.02	0.08

Table A3. Descriptive statistics of topsoil variables for Argiudolls and for grassland and native forests within of Argiudolls. Number of single samples (n), mean, standard deviation (SD), minimum (Min) and maximum (Max) of each variable are given.

	Total Argiudolls					Grassland					Native forests				
Variable (Unit)	n	Mean	SD	Min	Max	n	Mean	SD	Min	Max	n	Mean	SD	Min	Max
Soil moisture (%)	77	18.3	10.9	1.3	50.5	31	19.0	9.2	5.0	49.0	12	26.4	13.4	9.0	48.2
N_total (%)	77	0.24	0.15	0.07	1.20	31	0.21	0.11	0.07	0.66	12	0.34	0.28	0.13	1.20
C_total (%)	77	3.3	3.0	1.1	25.7	31	2.9	1.8	1.1	10.8	12	5.3	6.5	1.7	25.7
C/N ratio	76	13.3	2.2	9.8	19.9	31	13.3	2.1	10.5	19.4	11	13.6	2.1	9.8	18.5
SOC (%)	73	5.8	2.5	2.1	16.1	30	5.2	1.9	2.1	10.0	9	7.4	2.1	4.8	11.0
P (mg kg^{-1})	76	280	122	93	693	30	234	78	93	353	12	438	155	155	693
Ca (cmol kg^{-1})	21	11.4	8.0	2.6	37.1	8	11.5	6.7	3.2	20.5	4	16.2	14.3	5.1	37.1
Mg (cmol kg^{-1})	21	2.3	1.4	0.5	6.1	8	2.0	0.8	0.9	3.2	4	3.5	2.5	1.0	6.1
K (cmol kg^{-1})	21	0.45	0.26	0.1	1.3	8	0.4	0.14	0.16	0.61	4	0.69	0.41	0.34	1.29
Na (cmol kg^{-1})	21	0.08	0.05	0.03	0.26	8	0.07	0.02	0.04	0.10	4	0.10	0.11	0.03	0.26
Zn (mg kg^{-1})	21	2.4	3.0	0.1	10	8	1.3	0.8	0.1	2.2	4	5.1	4.3	1.3	10
Cu (mg kg^{-1})	21	15.5	42.7	0.8	161	8	2.6	1.6	1.0	6.0	4	1.9	1.0	0.8	3.0
Fe (mg kg^{-1})	21	117	62	10	232	8	104	50	28	171	4	117	84	10	216
Mn (mg kg^{-1})	21	23.3	10.4	7.1	43.6	8	20.5	10.4	7.1	33.8	4	28.8	10.4	15.4	39.4
pH	77	5.0	0.7	3.8	6.6	31	4.9	0.7	3.9	6.6	12	5.2	0.5	4.3	6.1
As (mg kg^{-1})	76	3.2	1.4	1.0	12.2	31	2.9	0.9	1.5	4.9	11	4.0	2.8	1.5	12.2
Cd (mg kg^{-1})	76	0.4	0.1	0.2	0.7	30	0.4	0.1	0.2	0.7	12	0.4	0.1	0.2	0.7
Cr (mg kg^{-1})	77	9.3	6.2	2.2	28.3	31	8.1	4.1	3.6	18.5	12	13.2	9.5	2.2	28.3
Pb (mg kg^{-1})	77	5.6	2.1	2.8	10.7	31	5.2	1.8	2.8	9.6	12	6.8	2.9	2.9	10.7
CEC (cmol kg^{-1})	21	14.2	9.3	3.4	44.6	8	14.0	7.3	4.5	23.1	4	20.4	16.7	6.4	44.6
K^{+}/(Ca^{2+}+Mg^{2+})	21	0.04	0.02	0.02	0.09	8	0.03	0.02	0.02	0.07	4	0.04	0.01	0.03	0.06
K^{+}/(Ca^{2+}+Mg^{2+}+Na^{+})	21	0.04	0.02	0.02	0.08	8	0.03	0.02	0.02	0.07	4	0.04	0.01	0.03	0.06

Table A4. Descriptive statistics of topsoil variables for Argiudolls and timber plantations and crops within of Argiudolls. Number of single samples (n), mean, standard deviation (SD), minimum (Min) and maximum (Max) of each variable are given.

	Total Argiudolls					Timber plantations					Crops				
Variable (unit)	n	Mean	SD	Min	Max	n	Mean	SD	Min	Max	n	Mean	SD	Min	Max
Soil moisture (%)	77	18.3	10.9	1.3	50.5	22	14.0	7.8	5.5	38.0	12	16.5	13.0	1.3	50.5
N_total (%)	77	0.24	0.15	0.07	1.20	22	0.22	0.10	0.10	0.45	12	0.22	0.06	0.13	0.29
C_total (%)	77	3.3	3.0	1.1	25.7	22	3.2	1.6	1.2	8.0	12	2.6	0.7	1.4	3.3
C/N ratio	76	13.3	2.2	9.8	19.9	22	14.0	2.6	11.3	19.9	12	11.8	0.6	10.7	12.7
SOC (%)	73	5.8	2.5	2.1	16.1	22	6.3	3.3	2.3	16.1	12	5.2	1.4	2.8	6.7
P (mg kg^{-1})	76	280	122	93	693	22	234	70	111	353	12	321	119	204	613
Ca (cmol kg^{-1})	21	11.4	8.0	2.6	37.1	5	6.9	4.1	2.6	11.4	4	11.7	5.9	7.7	20.4
Mg (cmol kg^{-1})	21	2.3	1.4	0.5	6.1	5	1.5	0.8	0.5	2.7	4	2.6	1.1	1.8	4.1
K (cmol kg^{-1})	21	0.45	0.26	0.09	1.29	5	0.34	0.21	0.09	0.60	4	0.46	0.27	0.29	0.85
Na (cmol kg^{-1})	21	0.08	0.05	0.0	0.3	5	0.1	0.06	0.05	0.20	4	0.06	0.02	0.04	0.08
Zn (mg kg^{-1})	21	2.4	3.0	0.1	9.7	5	0.5	0.4	0.2	1.2	4	4.1	4.2	0.6	9.7
Cu (mg kg^{-1})	21	15.5	42.7	0.8	161	5	2.0	1.5	1.0	4.5	4	71.8	83.4	0.8	161
Fe (mg kg^{-1})	21	117	62	10	232	5	158	73	69	232	4	91	38	63	143
Mn (mg kg^{-1})	21	23.3	10.4	7.1	43.6	5	29.8	9.8	17.5	43.6	4	15.3	3.7	9.9	18.1
pH	77	5.0	0.7	3.8	6.6	22	4.9	0.8	3.8	6.4	12	5.2	0.3	4.9	5.7
As (mg kg^{-1})	76	3.2	1.4	1.0	12.2	22	3.0	1.0	1.0	4.4	12	3.4	1.2	2.1	5.5
Cd (mg kg^{-1})	76	0.39	0.13	0.18	0.74	22	0.37	0.12	0.18	0.60	12	0.43	0.13	0.27	0.66
Cr (mg kg^{-1})	77	9.3	6.2	2.2	28.3	22	9.5	7.3	2.7	26.4	12	8.2	1.7	6.0	11.3
Pb (mg kg^{-1})	77	5.6	2.1	2.8	10.7	22	5.6	2.1	2.8	8.5	12	5.3	1.2	3.9	7.3
CEC (cmol kg^{-1})	21	14.2	9.3	3.4	44.6	5	8.8	4.6	3.4	13.6	4	14.8	7.0	9.9	25.0
K$^+$/(Ca^{2+}+Mg^{2+})	21	0.04	0.02	0.02	0.1	5	0.04	0.03	0.02	0.07	4	0.04	0.03	0.02	0.09
K$^+$/(Ca^{2+}+Mg^{2+}+Na$^+$)	21	0.04	0.02	0.02	0.08	5	0.04	0.02	0.02	0.07	4	0.04	0.03	0.02	0.08