



# **Soil and crop management practices and the water regulation functions of soils: a synthesis of meta-analyses relevant to European agriculture**

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1 **Abstract.** Adopting soil and crop management practices that conserve or enhance soil structure is critical for  
2 supporting the sustainable adaptation of agriculture to climate change, as it should help maintain agricultural  
3 production in the face of increasing drought or water excess without impairing environmental quality. In this  
4 paper, we evaluate the evidence for this assertion by synthesizing the results of 34 published meta-analyses of  
5 the effects of such practices on soil physical and hydraulic properties relevant for climate change adaptation in  
6 European agriculture. We also review an additional 127 meta-analyses that investigated synergies and trade-offs  
7 or help to explain the effects of soil and crop management in terms of the underlying processes and mechanisms.  
8 Finally, we identify how responses to alternative soil-crop management systems vary under contrasting agro-  
9 environmental conditions across Europe. This information may help practitioners and policymakers to draw  
10 context-specific conclusions concerning the efficacy of management practices as climate adaptation tools.

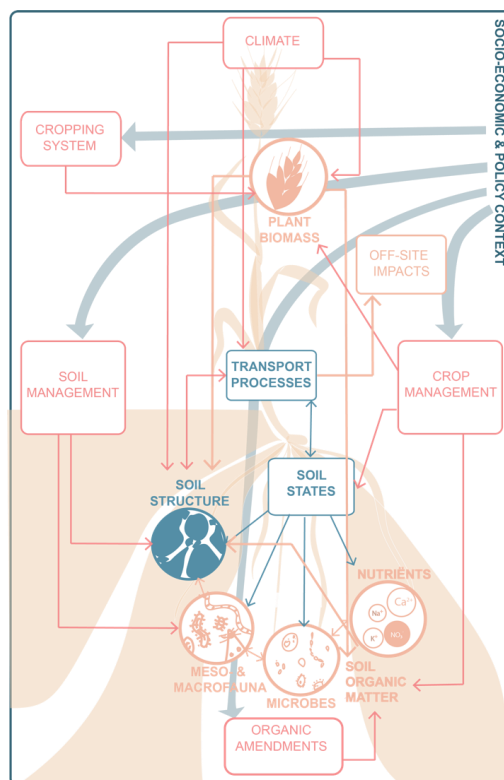
11 Our synthesis demonstrates that organic soil amendments and the adoption of practices that maintain  
12 “continuous living cover” result in significant benefits for the water regulation function of soils, mostly arising  
13 from the additional carbon inputs to soil and the stimulation of biological processes. These effects are clearly  
14 related to improved soil aggregation and enhanced bio-porosity, both of which reduce surface runoff and  
15 increase infiltration. One potentially negative consequence of these systems is a reduction in soil water storage  
16 and groundwater recharge, which may be problematic in dry climates. Some important synergies are reductions  
17 in nitrate leaching to groundwater and greenhouse gas emissions for non-leguminous cover crop systems. The  
18 benefits of reducing tillage intensity appear much less clear-cut. Increases in soil bulk density due to traffic  
19 compaction are commonly reported. However, biological activity is enhanced under reduced tillage intensity,  
20 which should improve soil structure, infiltration capacity, and reduce surface runoff and the losses of agro-  
21 chemicals to surface water. However, the evidence for these beneficial effects is inconclusive, while significant  
22 trade-offs include yield penalties and increases in greenhouse gas emissions and the risks of leaching of  
23 pesticides and nitrate.



## 24 **1 Introduction**

25 The occurrence of extreme weather events, such as high temperatures, summer droughts, waterlogging and  
26 flooding, will most probably increase in many parts of Europe as a consequence of on-going climate change  
27 (IPCC 2021). An urgent task is to develop guidance on management practices that help farmers adapt to these  
28 extreme weather situations.

29 The ecosystem services a soil can deliver depend profoundly on its structure, which we define here as the spatial  
30 arrangement of the soil pore space. Mediated by various biological (e.g. faunal and microbial activity) and  
31 physical processes (e.g. traffic compaction, wet-dry and freeze-thaw cycles), soil structure is constantly evolving  
32 at time scales ranging from seconds to centuries, driven by weather patterns as well as changes in climate and  
33 land management practices (figure 1). In turn, soil structure strongly affects all life in soil as well as the balance  
34 between infiltration and surface runoff, as well as drainage and soil water retention and therefore the supply of  
35 water and nutrients to crops. Practices commonly adopted in “conservation agriculture” (Palm et al., 2014) are  
36 thought to enhance soil structure and should therefore help to maintain agricultural production in the face of  
37 severe droughts or heavy rain. Conservation agriculture to improve soil structure rests on three fundamental  
38 principles (Palm, et al., 2014): i.) minimizing mechanical soil disturbance, ii.) maintaining soil cover by plants as  
39 much as possible and for as long as possible (i.e. aspects of both spatial and temporal coverage), and iii.)  
40 diversifying cropping. Other more recently coined and partially related terms are “regenerative agriculture”,  
41 which acknowledges past failures to preserve soil health (Schreefel et al., 2020) and “climate-smart agriculture”,  
42 which is defined by FAO (2010) as “... *agriculture that sustainably increases productivity, enhances resilience,*  
43 *reduces greenhouse gases, and enhances achievement of national food security and development goals*”.



44

45 **Figure 1. Schematic diagram of drivers, agents and processes governing the dynamics of soil structure and**  
46 **its effects on the soil-plant system.**

47 The effects of soil and crop management practices on soil properties, soil hydrological and biological  
48 functioning and crop performance have been studied in many long-term field trials throughout the world. In  
49 addition to narrative reviews (e.g. Palm et al., 2014), many quantitative meta-analyses synthesizing the findings  
50 of individual experiments have also been published. This is especially the case in the last few years (Beillouin et  
51 al., 2019a,b), probably because the number of field experiments that have been running for a sufficient length of  
52 time has only recently reached the critical mass required to enable these kinds of quantitative analyses. Indeed,  
53 the increase in the number of meta-analyses published on topics related to conservation agriculture has been so  
54 dramatic that four over-arching syntheses of these meta-analyses have also recently been published. Bolinder et  
55 al. (2020) evaluated the effects of organic amendments and cover crops on soil organic matter (SOM) storage,  
56 while Schmidt et al. (2021) focused on the effects of biochar on crop performance. Beillouin et al. (2019b) and  
57 Tamburini et al. (2020) carried out even more ambitious and comprehensive reviews of meta-analyses of the  
58 effects of conservation agriculture and crop diversification strategies on a wide range of ecosystem services.



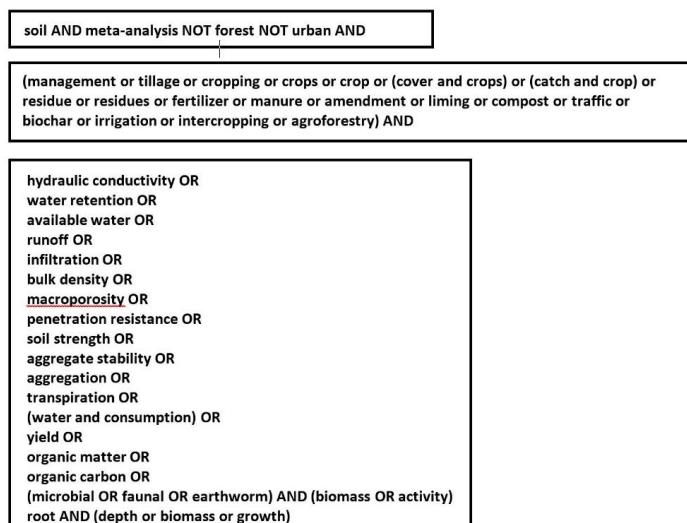
59 Tamburini et al. (2020) concluded that diversification practices most often resulted in a ‘win-win’ situation for  
60 ecosystem services including crop yields, but that the often large variability in responses and the occurrence of  
61 trade-offs highlighted the need to analyze the context-dependency of outcomes, something which was only  
62 possible to do to a limited extent with their broad-brush treatment. These previous syntheses of meta-analyses on  
63 the benefits of conservation agriculture have placed very little emphasis (Tamburini et al., 2020) or none at all  
64 (Beillouin et al., 2019a,b; Bolinder et al., 2020) on soil hydrological functioning even though this is key for  
65 climate change adaptation. In their synthesis, Tamburini et al. (2020) included 17 meta-analyses (involving 31  
66 effects-size comparisons) relevant to water regulation, but most of these concerned water quality issues rather  
67 than hydrological functioning *per se*. Beillouin et al. (2019b) concluded that ... “*our review reveals that a*  
68 *significant knowledge gap remains, in particular regarding water use*”.

69 In this study, we focus on the implications of agricultural management practices for soil hydrological functioning  
70 for climate change adaptation under European agro-environmental conditions. We do this by identifying and  
71 synthesizing existing meta-analyses of the response of soil physical/hydraulic properties and hydrological  
72 processes relevant for climate change adaptation to soil and crop management practices. This evaluation  
73 highlights where consensus has been established and identifies remaining knowledge gaps. In those cases where  
74 the information is available, we summarize knowledge of context-specific effects of relevance for the range of  
75 agro-environmental conditions found in Europe, and as far as possible, explain these variations in terms of  
76 individual driving processes and mechanisms. This kind of information may explain local praxis in agricultural  
77 management (i.e. farmer behavior) and will enable practitioners and policymakers to draw context-specific  
78 conclusions concerning the efficacy of management practices as climate adaptation tools.

## 79 **2 Materials and Methods**

### 80 **2.1 Literature search**

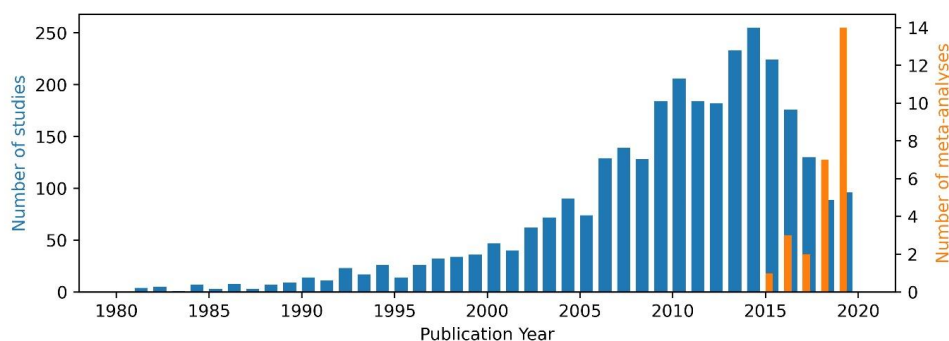
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83 **Figure 2. Search string used to identify relevant meta-analyses.**

84 The text string shown in Figure 2 was used to search the published literature using Web of Knowledge in May  
85 2021. This search returned 663 results. All search results were manually assessed for their relevance to the  
86 objectives of our study. Meta-analyses that only included studies carried out outside Europe were not retained.  
87 Our search identified 34 relevant meta-analyses focusing on the effects of soil and crop management on soil  
88 physical properties and hydrological processes using effects ratios (Appendix 2). Figure 3 shows the number of  
89 primary studies per publication year included in the 34 meta-analyses. A peak is clearly visible in 2014, which is  
90 explained by the fact that all the selected meta-analyses were published after 2015. Our search string was also  
91 designed to identify meta-analyses of management effects on soil organic matter and biological variables (e.g.  
92 microbial biomass), since these help to explain the observed effects on physical/hydraulic properties and  
93 hydrological processes, as well as other studies that analyzed target variables representing potential “trade-offs”  
94 or synergies. Among these, we focused primarily on the impacts of management practices on crop yields,  
95 greenhouse gas emissions and water quality. An additional 127 published meta-analyses of this kind were  
96 identified by our literature search. These studies are listed in the supplementary file (“*Supporting studies.xlsx*”).



97

98 Figure 3. Number of primary studies included in the 36 selected meta-analyses published per year and the  
 99 publication year of these meta-analyses.

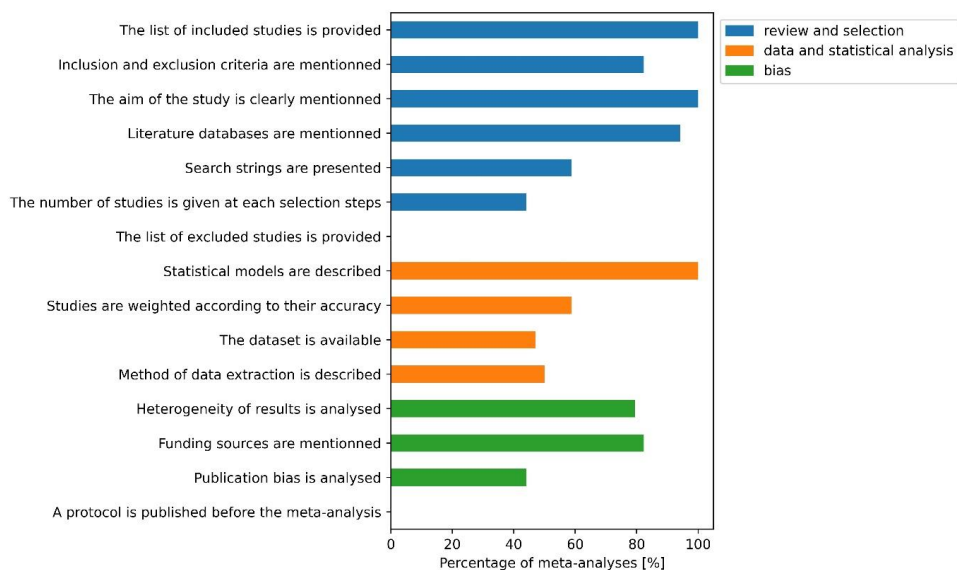
100 The target variables (e.g. soil physical and hydraulic properties) and drivers (i.e. soil and crop management  
 101 practices) included in the 34 meta-analyses were then classified into a limited number of groups. The target  
 102 variables were grouped into five classes: pore space properties (e.g. porosity, bulk density), hydraulic properties  
 103 (e.g. saturated hydraulic conductivity, field capacity), mechanical properties (e.g. soil aggregate stability,  
 104 penetration resistance), water flows (e.g. infiltration, surface runoff, drainage) and plant properties (e.g. root  
 105 length density, water use efficiency). Likewise, the management practices were also grouped into five classes:  
 106 soil amendments (e.g. manure, biochar, organic farming systems), cropping practices and systems (e.g. cover  
 107 crops, crop rotations), tillage systems (e.g. no-till), grazing management and irrigation. In total, the 34 meta-  
 108 analyses reported 104 effects ratios comparing the impacts of a management practice to a control treatment for a  
 109 particular response variable. The effects of these treatments on the target variables (either positive, negative or  
 110 neutral i.e. non-significant) were read from tables and figures in each of the 34 meta-analyses.

111 **2.2 Quality assessment**

112 FWe performed a quality assessment of the selected 34 meta-analyses using 15 of the criteria proposed by  
 113 Beillouin et al. (2019a). Figure 4 presents a summary of the quality of the selected meta-analyses according to  
 114 these criteria. Nearly half of the meta-analyses included datasets in the paper, while only ca. 44% investigated  
 115 the important issue of publication bias (Philibert et al., 2012). The authors of these studies used simple statistical  
 116 techniques such as frequency distributions of effects sizes or “funnel plots” of sample sizes against effect sizes to  
 117 investigate whether experiments with non-significant effects are under-represented in the literature. For both of  
 118 these methods, symmetry of the distributions is taken to indicate a lack of bias. Two studies detected evidence of  
 119 publication bias (e.g. Basche and deLonge, 2019; Shackelford et al., 2019) using this method, although in both



120 cases the effects on the overall conclusions of the studies were considered marginal. Basche and deLonge (2019)  
 121 also investigated the sensitivity of the outcome to the exclusion of individual studies, which is another important  
 122 aspect of publication bias. They found mostly robust results for the impacts of management practices on  
 123 infiltration, especially for no-tillage and cover crops.



124

125 **Figure 4. Proportion of the quality criteria defined by Bellouin et al. (2019a) that are met by the selected**  
 126 **meta-analyses in this study.**

127 **2.3 Redundancy**

128 We performed a redundancy analysis to identify the proportion of common primary or source studies among the  
 129 meta-analyses following the methodology of Beillouin et al. (2019a). For each of the 34 selected meta-analyses,  
 130 the references to the studies used were extracted from the supplementary materials. Each reference contained at  
 131 least the name of the first author, the year of publication, the title, the journal and – if available – the DOI. Of the  
 132 3142 unique primary studies, 437 had no DOI. Old publications or publications not written in English were  
 133 usually found to have no DOI. In some cases, the title and DOI were not available so we had to manually check  
 134 these references based on contextual information supplied in the supplementary material. In most cases,  
 135 however, the title was provided in the meta-analysis and the DOI could be extracted automatically from the  
 136 Cross-Ref database. We then manually checked if the title of the paper matched the one found on Cross-ref, to  
 137 confirm the DOI assignment. The results of the redundancy analysis are presented in the Appendix 1 (Figures  
 138 A1-A3) as well as in the notebook at <https://github.com/climasoma/review-of-meta->



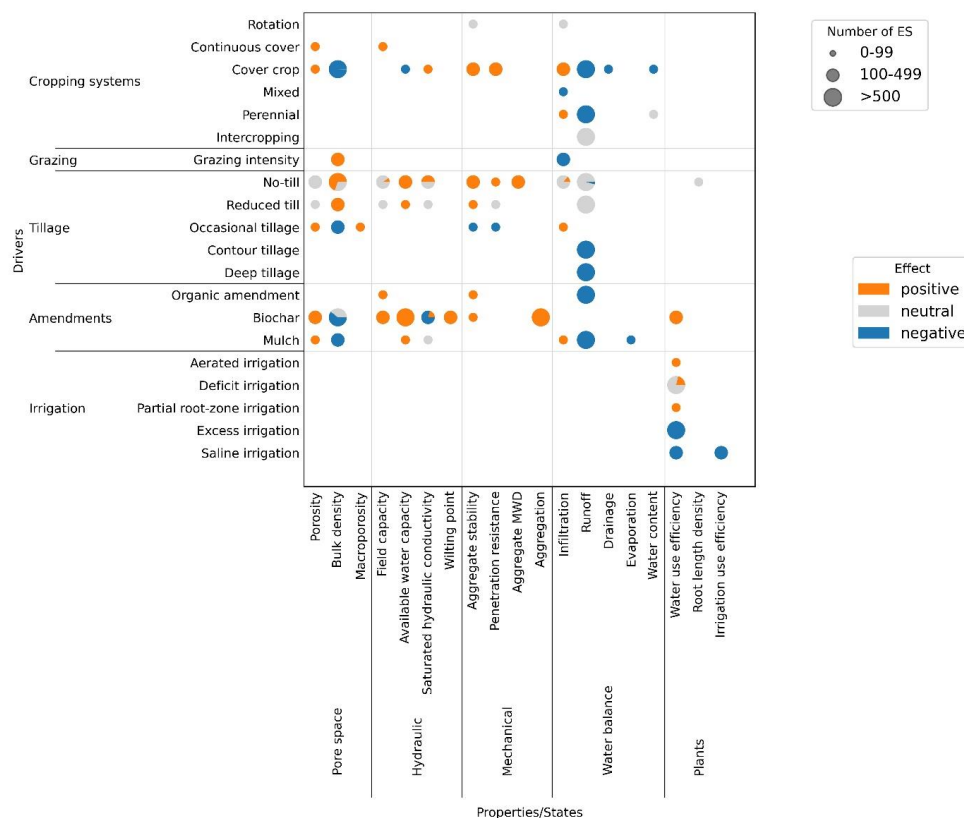


139 [analyses/blob/main/notebooks/redundancy.ipynb](#). The main outcome of this analysis is that redundancy  
140 is only an important factor for a few meta-analyses on biochar that were published almost simultaneously (e.g.  
141 Edeh, et al., 2020; Rabbi, et al., 2021).

### 142 **3 Results and discussion**

#### 143 **3.1 Knowledge gaps**

144 Figure 5 summarizes the statistical relationships found between the drivers and target variables in the selected  
145 meta-analyses. Gaps in the scatter plots shown in figure 5 indicate particular combinations of drivers and target  
146 variables that have not been the subject of meta-analysis according to our search criteria. Inspection of these  
147 figures suggests that there are several significant “knowledge gaps”. Although several meta-analyses have  
148 focused on the effects of irrigation management or organic amendments on water use efficiency, none have been  
149 published specifically on the effects of management practices on water supply to crops. Such information should  
150 be critical to support policies and practices for effective adaptation of farming systems to future climates with  
151 more frequent and severe summer droughts. We can therefore only make inferences about the effects of soil  
152 management on crop transpiration from other terms in the soil water balance. Other knowledge gaps may be only  
153 apparent and therefore less serious: macroporosity is rarely studied in the context of meta-analysis, although  
154 infiltration has been much more frequently measured and these two variables should be strongly correlated.  
155 Figure 5 also indicates that some management practices have been less often the subject of field experiments  
156 including, for example, deep tillage, occasional tillage, and crop rotations. Presumably for reasons of cost, many  
157 long-term field experiments often only have simple designs, neglecting potentially interesting combinations of  
158 treatments, for example, no-till combined with the use of cover crops. Similarly, the interactions between soil  
159 and crop management and irrigation or drainage systems and practices do not appear to be a common topic of  
160 field experimentation.



161

162 **Figure 5. Effects of drivers (vertical axis) on target variables (horizontal axis) in the 36 selected meta-**  
 163 **analyses. The coloured pie charts represent the directions of the statistical effects in the different meta-**  
 164 **analyses, while the size of the circle indicates the total number of effects sizes (ES) reported. Note that this**  
 165 **number has not been corrected for redundancy. Blank cells denote that no data was available for this**  
 166 **target variable in any of the selected meta-analyses.**

167 Some additional potential knowledge gaps concerning the effects of soil management on water regulation  
 168 functions are not revealed by inspection of figure 5, since they concern variables that are rarely measured and so  
 169 have not yet been the subject of meta-analysis. Most long-term field trials on the effects of soil and crop  
 170 management practices on hydrological and biological functioning have measured surrogate variables (or proxies)  
 171 for soil structure, such as infiltration rates or soil hydraulic properties (water retention, hydraulic conductivity at  
 172 and near saturation). No meta-analyses have been performed yet for metrics quantifying different aspects of soil  
 173 structure *per se* (Rabot et al., 2018) even though the application of X-ray imaging techniques to quantify soil  
 174 structure is now becoming increasingly common. As a result, the number of X-ray studies published is rapidly  
 175 increasing, so it should not be too long before it will be possible and worthwhile to carry out such an analysis.



176 **3.2 Cropping systems and practices**

177 Broadly speaking, published meta-analyses that have investigated the effects of cropping systems and practices  
178 (figure 5) fall into two categories: i.) studies analyzing the effects of maintaining a more continuous soil surface  
179 cover, either in a temporal (e.g. cover crops in arable rotations) or in a spatial sense (e.g. inter-row cover in  
180 widely-spaced row crops such as vineyards and orchards), and ii.) studies comparing farming systems (e.g.  
181 continuous arable contrasted with either perennial crops or rotations or mixed farming systems with livestock).  
182 In the following, we combine these two aspects, referring to both of them as cropping systems that as far as  
183 possible maintain a “continuous living cover” (Basche and deLonge, 2017).

184 Figure 5 shows that meta-analyses have identified several beneficial effects of such agronomic practices on  
185 important physical and hydraulic properties in soil, such as porosity or bulk density, saturated hydraulic  
186 conductivity and aggregate stability (Basche and deLonge, 2017; Jian et al., 2020). These positive effects are  
187 almost certainly due to a combination of the protective effects of surface cover against the degradation of soil  
188 structure by raindrop impact as well as the enhancement of various biological processes that occurs as a  
189 consequence of plant growth, root production and the additional carbon supply to the soil. In this respect, meta-  
190 analyses have demonstrated that practices that maintain a continuous living cover (e.g. rotations with leys, cover  
191 crops) promote increases in microbial biomass, activity and diversity (Venter et al., 2016; Shackleford et al.,  
192 2019; Jian et al., 2020; Kim et al., 2020; Muhammad et al., 2021) and increase soil organic matter contents in the  
193 long-term (Aguilera et al., 2013; McDaniel et al., 2014; Poeplau and Don, 2015; King and Blesh, 2018; Bai et  
194 al., 2019; Shackleford et al., 2019; Bolinder et al., 2020; Jian et al., 2020; McClelland et al., 2021). This will  
195 both promote stable soil aggregation and reduce soil bulk density (Chenu et al., 2000; Meurer et al., 2020a,b).  
196 The abundance of soil meso- and macro-fauna also increases under long-term cover cropping (Reeleder et al.,  
197 2006; Roarty et al., 2017) and perennial crops such as grass/clover leys (Fraser et al., 1994; Bertrand et al., 2015;  
198 Jarvis et al., 2017). Through their burrowing activity, these “ecosystem engineers” (Jones et al., 1994) create  
199 networks of large biopores in soil (Jarvis, 2007) that greatly increase saturated and near-saturated hydraulic  
200 conductivity and thus infiltration capacity (e.g. Bertrand et al., 2015; Capowiez et al., 2021).

201 The changes in soil physical and hydraulic properties brought about by the introduction of continuous living  
202 cover have significant beneficial consequences for the water regulation function of soils. Thus, cover crops  
203 enhance infiltration capacity and reduce surface runoff (Xiong et al., 2018; Basche and deLonge, 2019; Lee et  
204 al., 2019; Jian et al., 2020; Liu et al., 2021). An increased proportion of perennial crops in the rotation and the  
205 presence of ground cover between the rows of perennial crops (e.g. in vineyards) increase soil infiltration and



206 reduce surface runoff (Xiong et al., 2018; Basche and deLonge, 2019; Liu et al., 2021). These positive effects  
207 seem broadly similar regardless of climate (Xiong et al., 2018; Liu et al., 2021).

208 Some negative consequences of mixed farming systems with grazing livestock for soil physical properties have  
209 been noted. Meta-analyses have shown that high grazing intensities result in significantly poorer soil physical  
210 quality, in terms of larger bulk densities (Byrnes et al., 2018) and reduced infiltration rates (deLonge and Basche,  
211 2018; Basche and deLonge, 2019) as a result of compaction by animal trampling. These impacts of intensive  
212 grazing are similar irrespective of soil texture or climate, although they appear to be slightly larger in wetter  
213 climates (Byrnes et al., 2018; deLonge and Basche, 2018). Another significant negative effect is that by  
214 increasing transpiration, systems employing “continuous living cover” may reduce soil water content  
215 (Shackleford et al., 2019) and decrease recharge to groundwater. Thus, for a combined dataset of 36 studies  
216 comprising both experimental and modelling studies, Meyer et al. (2019) found that cover crops reduced  
217 recharge by 27 mm/year on average with no apparent effects of climate, soil type or cropping system. For their  
218 meta-analysis based on a more limited dataset of six studies, Winter et al. (2018) found no significant effects of  
219 inter-row vegetation in vineyards on the soil water balance. Other impacts of cover crops on physical properties  
220 with adverse consequences for crop water supply have been reported, for example an increase in soil penetration  
221 resistance and a reduced available water capacity (Jian et al 2020), although it seems quite difficult to identify  
222 plausible mechanisms for such effects. As noted earlier, “continuous living cover” increases soil organic matter  
223 contents and both long-term field experiments and meta-analyses suggest that soil organic matter generally tends  
224 to increase the plant available water capacity. However, although the magnitude of this effect is still a matter of  
225 debate (Lal, 2020), in most cases it seems relatively small compared with the crop water demand (Minasny and  
226 McBratney, 2018a,b; Libohova et al., 2018).

227 With respect to potential synergies and trade-offs, studies have shown that cover crops mostly have either neutral  
228 or positive effects on main crop yields (Tonitto et al., 2006; Quemada et al., 2013; Valkama et al., 2015; Angus  
229 et al., 2015; Marcillo and Miguez, 2017). However, Shackleford et al. (2019) reported an average 7% reduction  
230 in cash crop yields for systems employing non-legume cover crops in dry Mediterranean climate conditions.  
231 Similarly, in a recent meta-analysis on cover crops grown in climates with less than 500 mm annual rainfall,  
232 Blanco-Canqui et al. (2022) found that cover crops decreased main crop yields in 38% of cases, with no effects  
233 found in 56% of cases and increased yields in 6% of cases. Non-leguminous cover crops significantly reduce  
234 nitrate leaching and, to a lesser extent, N<sub>2</sub>O emissions, although this is clearly not the case for legumes (Tonitto  
235 et al., 2006; Quemada et al., 2013; Basche et al., 2014; Valkama et al., 2015; Muhammad et al., 2019;



236 Shackleford et al., 2019). Our literature search did not identify any meta-analyses on phosphorus or pesticide  
237 losses.

### 238 **3.3 Tillage systems**

239 A large number of meta-analyses have investigated the effects of tillage practices on soil properties and  
240 functions and the provision of various ecosystem services (figure 5). The control treatment in these published  
241 meta-analyses is usually conventional tillage (CT), which involves both inversion ploughing and shallow  
242 secondary tillage operations for seedbed preparation. This control treatment is then contrasted with either  
243 reduced (or minimum) tillage (RT), whereby the soil is no longer ploughed, or no-till (NT) systems in which the  
244 soil is left completely undisturbed, or both. One meta-analysis has investigated the effects of deep tillage and  
245 contour tillage on surface runoff (Xiong et al., 2018), while another analyzed the effects of occasional tillage  
246 with no-till as the control treatment (Peixoto et al., 2020).

247 Reductions in the depth and intensity of tillage (i.e. from CT to RT to NT) strongly influence carbon cycling in  
248 the soil-crop system. Several meta-analyses show that soil organic carbon concentrations are larger under RT  
249 and NT systems in the uppermost soil layers (e.g. Bai et al., 2018; Lee et al., 2019) especially in fine-textured  
250 soils (Bai et al., 2019). The reasons for this are the lack of soil disturbance that promotes a stable aggregated  
251 structure, which affords a greater physical protection of C against microbial mineralization (Kan et al., 2021) and  
252 the elimination of physical mixing and re-distribution of C within the topsoil due to the absence of soil inversion  
253 by ploughing (Meurer et al., 2020b). Meta-analyses have shown that the accumulation of SOM typically found  
254 in surface soil layers under RT and NT systems, which reflects the deposition and accumulation of plant  
255 residues, is paralleled by a greater microbial biomass (e.g. Spurgeon et al., 2013; Zuber and Villamil, 2016; Li et  
256 al. 2018; Lee et al. 2019; Li et al. 2020b,c; Chen et al. 2020) and increases in enzyme activities (Zuber and  
257 Villamil, 2016; Lee et al., 2019). The diversity of bacterial and sometimes also fungal communities tends to be  
258 greater in RT or NT (Spurgeon et al., 2013; de Graaff et al., 2019; Li et al 2020b), especially where these  
259 systems are combined with the retention of crop residues (Li et al., 2020c).

260 In addition to focusing on organic carbon concentrations in topsoil, differences in SOC stocks under  
261 conservation tillage systems in complete crop root zones and soil profiles are also of interest, not least from the  
262 point of view of climate change mitigation. Based on a meta-analysis of studies with measurements made to at  
263 least 40 cm depth, Luo et al. (2010) concluded that NT did not increase soil carbon stocks. This is because  
264 although SOC contents are usually larger under NT than CT systems in the uppermost soil layers, they can be



265 significantly smaller both at plough depth and in the upper subsoil (Angers and Eriksen-Hamel, 2008). Thus, for  
266 boreo-temperate climates, Haddaway et al. (2017) and Meurer et al. (2018) found increases in soil carbon stocks  
267 under NT compared to CT only in the topsoil, while no overall significant effect on carbon stocks was detected  
268 for soil profiles to 60 cm depth. In a more recent global meta-analysis, Mondal et al. (2020) found no significant  
269 differences in stocks of soil organic carbon between NT and CT systems, while variations in response could not  
270 be attributed to either climate or soil type. In apparent contrast, Mangalassery et al. (2015) concluded that NT  
271 systems result in a net sequestration of carbon, regardless of the depth of soil considered. Sun et al. (2020)  
272 demonstrated significant effects of climate on the changes in organic carbon stocks observed under NT systems.  
273 In their global analysis, they found that soil C sequestration was enhanced in warmer and drier regions, while  
274 soils under no-till in colder and wetter climates were just as likely to lose soil C as gain C. These findings are  
275 supported by the regional-scale studies of Meurer et al. (2018) for boreo-temperate climates and Gonzalez-  
276 Sanchez et al. (2012) and Aguilera et al. (2013) for Mediterranean climates, although for vineyards, Payen et al.  
277 (2021) found larger topsoil C sequestration in temperate climates than hot and dry climates. A loss of organic  
278 carbon following adoption of NT systems can be explained by a decrease in carbon inputs to soil resulting from  
279 poorer crop growth (Mangalassery et al., 2015; Pittelkow et al., 2015), which compensates for reductions in  
280 carbon mineralization rates (Ogle et al., 2012; Virto et al., 2012). Such yield penalties under no-till are especially  
281 prevalent in colder and wetter climates (Sun et al., 2020).

282 Soil tillage directly affects soil macro-fauna by mechanically harming or killing them. In addition to these direct  
283 effects of soil disturbance, disruption of the soil also exposes soil macro-fauna to increased risks of desiccation  
284 and predation. Consequently, meta-analyses show that total earthworm biomass and abundance increase as  
285 tillage intensity is reduced (Spurgeon et al., 2013; Briones and Schmidt 2017; Bai et al. 2018), with a negative  
286 relationship between tillage depth and earthworm abundance (Briones and Schmidt 2017). Deep burrowing and  
287 surface-feeding (anecic) earthworm species are particularly favored by NT systems, as their permanent burrows  
288 are no longer destroyed by ploughing and they have a better access to food resources. Thus, a lack of disturbance  
289 of the soil by tillage has also been shown to increase the diversity of earthworm populations in particular (Chan,  
290 2001; Spurgeon et al., 2013; Briones and Schmidt 2017) and soil fauna in general (de Graaff et al., 2019).

291 Changes in tillage systems directly affect the physical properties of soil. For example, bulk density and  
292 penetration resistance often increase after the adoption of RT and NT systems (Lee et al. 2019; Li et al., 2019; Li  
293 et al., 2020a) due to traffic compaction and the lack of loosening by cultivation (Hamza and Anderson 2005).  
294 Peixoto et al. (2020) showed that these negative effects can be alleviated with occasional tillage. The impacts of



295 conservation tillage practices on soil biological agents and processes also give rise to significant indirect effects  
296 on physical properties, hydrological processes and ecosystem services related to water regulation. Thus, meta-  
297 analyses have shown that saturated hydraulic conductivity and surface infiltration rates often increase under  
298 conservation tillage compared with CT, especially for NT systems (Li et al., 2019; Basche and deLonge, 2019;  
299 Li et al., 2020a; Mondal et al., 2020). This suggests that the effects of the enhanced bioporosity in NT systems  
300 created by soil fauna, and especially anecic earthworms, on saturated and near-saturated hydraulic conductivity  
301 (Lee and Foster, 1991) generally outweigh the negative effects of increased bulk density. Thus, Spurgeon et al.  
302 (2013) showed in their meta-analysis that increased earthworm abundances and diversity found under NT  
303 systems were positively correlated with infiltration rates. Comparing ecological groups, they found that the  
304 density of anecic earthworms was positively associated with increased infiltration rates, whereas no effect was  
305 apparent for endogeic earthworms. Aggregate stability is also largest under NT systems, is intermediate when  
306 occasional tillage is practiced (Peixoto et al. 2020) and smallest in CT systems (Bai et al. 2018). In their meta-  
307 analysis, Spurgeon et al. (2013) showed that improved aggregate stability under NT systems was positively  
308 correlated with increases in fungal biomass. A lack of soil disturbance in NT systems also increases the mean  
309 size of aggregates produced in stability tests (Li et al., 2020a; Mondal et al., 2020). Several meta-analyses have  
310 demonstrated increases in field capacity and available water capacity under reduced and no-till systems (Li et al.,  
311 2019; Mondal et al., 2020; Li et al., 2020a), presumably due to enhanced soil biological activity and increases in  
312 organic carbon content. This would improve water supply to crops under drought, although the effects would  
313 appear to be relatively small.

314 In principle, better-developed soil macropore systems and improvements in aggregate stability should promote a  
315 more favorable crop water balance, with increases in infiltration and reductions in surface runoff. Figure 5 shows  
316 that the effect on runoff is one of the most studied hydrological processes related to tillage. The meta-analysis  
317 performed by Sun et al. (2015) found that RT and NT systems decreased surface runoff. However, these results  
318 do not appear to be conclusive as two later meta-analyses (Mhazo et al., 2016; Xiong et al., 2018) failed to detect  
319 significant effects of conservation tillage practices on surface runoff. However, Xiong et al. (2018) found that  
320 contour tillage and deep tillage both reduced surface runoff.

321 Adoption of no-till and reduced tillage systems involve several trade-offs, particularly concerning water quality,  
322 GHG emissions and crop yields. As noted earlier, NT systems tend to give smaller yields for many crops  
323 compared with conventional tillage (Mangalassery et al., 2015, Pittelkow et al., 2015, Sun et al., 2020). This may  
324 explain why no-till systems are still seldom adopted in Europe (Mangalassery et al., 2015; Bai et al. 2018),



325 although reduced tillage (RT) is being increasingly adopted worldwide. In their comprehensive meta-analysis,  
326 Pittlekow et al. (2015) identified several reasons for variations in the yield response to no-till. Crop type was the  
327 most important, with no significant yield losses found under NT for oilseed, cotton and legume crops, while the  
328 yields of cereals and root crops were on average ca. 5% and 20% smaller respectively. In accordance with the  
329 results of the meta-analyses on stocks of soil organic carbon discussed earlier, Pittlekow et al. (2015) and Sun et  
330 al. (2020) also found climate to be a significant factor, with no significant yield losses for no-till systems under  
331 rain-fed conditions in dry climates. In contrast, Peixoto et al. (2020) showed that occasional tillage increased  
332 crop yields compared with NT in dry regions and in soils with limited water retention capacity and availability,  
333 presumably by alleviating soil compaction and improving rooting.

334 With respect to water quality, Daryanto et al. (2017a) found an overall 40% reduction in phosphorus loads in  
335 surface runoff for NT systems in comparison with CT. This was attributed to significant decreases in losses of  
336 particulate phosphorus, as concentrations of dissolved P actually increased in runoff under NT. For pesticides,  
337 Elias et al. (2018) found no significant differences in concentrations in surface runoff for 14 of the 18  
338 compounds included in their meta-analysis. Pesticide concentrations were actually larger under NT for the  
339 remaining 4 compounds. For loads, no significant difference was detected between CT and NT systems for 15 of  
340 the 18 pesticide compounds. For the three remaining pesticides, losses in surface runoff were larger under NT for  
341 metribuzin and dicamba and smaller for alachlor. As also noted by Elias et al. (2018), these results seem quite  
342 surprising given the documented effects of conservation tillage on soil structure and hydraulic properties in the  
343 uppermost soil layers discussed earlier, which should increase soil infiltration capacity and reduce surface  
344 runoff. For nitrate losses in surface runoff in conventional and no-till systems, Daryanto et al. (2017b) showed  
345 that a change to NT resulted in an increase in nitrate concentrations in surface runoff, but similar loads, implying  
346 that surface runoff was, as expected, less prevalent under NT.

347 Daryanto et al. (2017b) also performed a meta-analysis on nitrate leaching. They found larger leachate losses of  
348 nitrate under NT systems than CT, whereas the concentrations in leachate were similar under both tillage  
349 systems, indicating that the effect of NT on nitrate leaching was largely determined by increases in water  
350 percolation. We did not find any meta-analyses on the effects of tillage systems on pesticide leaching in our  
351 literature search. Leaching is the outcome of several interacting processes involving many complex and poorly  
352 understood processes (Alletto et al., 2010). In practice, with no mechanical disturbance, larger quantities of  
353 pesticides are often used to control weeds and diseases in NT systems. However, pesticide leaching will also be  
354 highly sensitive to changes induced by tillage in soil structure, microbial biomass and activity and SOC, since





355 these will affect water flow velocities, degradation rates and the strength of adsorption in soil. Several studies  
356 suggest that the better-preserved macropore networks established under RT and NT systems may enhance  
357 leaching by preferential flow (Jarvis, 2007; Larsbo et al., 2009; Alletto et al., 2010). Although it is difficult to  
358 draw firm conclusions about the effects of conservation tillage practices on pesticide leaching without the help of  
359 quantitative meta-analyses, we may tentatively conclude that the greater risk of macropore flow under RT and  
360 NT systems appears to outweigh any beneficial impacts of increases in SOC and microbial activity on pesticide  
361 adsorption and degradation.

362 Significant trade-offs have also been reported with respect to greenhouse gases. In an early meta-analysis, van  
363 Kessel et al. (2013) found no overall impact of reduced tillage or no-till on N<sub>2</sub>O emissions, with observed  
364 increases in humid climates compensated by reductions in emissions in drier climates, although neither trend was  
365 significant. However, in a later meta-analysis, Mei et al. (2018) reported a significant overall increase of 18% in  
366 N<sub>2</sub>O emissions under conservation tillage, with the largest effects in warmer and wetter climates and in finer-  
367 textured soils. In a recent meta-analysis, Shakoor et al. (2021) found significant increases of emissions of CO<sub>2</sub>,  
368 N<sub>2</sub>O and CH<sub>4</sub> of 7, 12 and 21% respectively under NT compared with CT. From the perspective of climate  
369 change mitigation, Guenet et al. (2020) concluded that increased greenhouse gas emissions under NT  
370 outweighed any minor gains in soil C stocks.

### 371 **3.4 Amendments**

#### 372 3.4.1 Biochar

373 Biochar is charcoal made for the purpose of soil amendment. It is a type of black carbon, resulting from  
374 incomplete combustion of organic matter through a process known as pyrolysis. Apart from its potential for  
375 long-term soil carbon sequestration, it also has beneficial effects on nutrient availability and soil physical  
376 properties (Joseph et al., 2021).

377 The quantitative analysis of the effects of biochar on physical and hydraulic properties shown in figure 5 is based  
378 on effects ratios presented in five meta-analyses (Omondi et al., 2016; Edeh et al., 2020; Gao et al., 2020; Rabbi  
379 et al., 2021; UI Islam et al., 2021). Our review also draws on findings presented in two additional reviews that  
380 employed different statistical methodologies (Kroeger et al., 2020; Razzaghi et al., 2020). Rabbi et al. (2021)  
381 only presented data for various sub-categories (e.g. for different types of biochar) and not for the overall effects  
382 of biochar addition. Taken altogether, these seven studies present results of analyses for different soil types,



383 textural classes and experimental conditions (i.e. field or laboratory/greenhouse study) as well as for biochars of  
384 different properties and applied at different rates. These analyses show that biochar has several positive effects  
385 on soil hydraulic properties, but that these effects are dependent on all of the above-mentioned variables.

386 Decreases in bulk density and increases in porosity are generally reported after biochar addition (Omondi et al.,  
387 2016; Edeh, et al., 2020). The density of biochar is low and the porosity is often high compared to soil, which  
388 may explain the observed effects. However, if biochar mainly fills existing pores, porosity will decrease and bulk  
389 density increase. Biochar will also influence these variables indirectly through its effects on aggregation (Pituello  
390 et al., 2018).

391 Figure 5 suggests that biochar addition generally increases the plant available water content ( $\theta_{paw}$ ). These meta-  
392 analyses show that although the water contents at field capacity ( $\theta_{fc}$ ; pressure potentials in the range between -  
393 0.033 and -0.01 MPa) and wilting point ( $\theta_{pwp}$ ) both tend to increase following biochar amendment, the effects on  
394  $\theta_{fc}$  appear to be larger (figure 5). Pore sizes in biochars range over at least five orders of magnitude, from the  
395 sub-nanometer scale to pore diameters of the order of tens of micrometers originating from partially preserved  
396 cellular structures (Brewer et al., 2014). However, a large fraction of the pore volume in biochar consists of  
397 pores in the nanometer size range (Downie et al., 2009). These pores will retain water at very low pressure  
398 potentials and therefore have the potential to increase the wilting point water content  $\theta_{pwp}$  upon biochar addition.  
399 It has been suggested that increases in  $\theta_{paw}$  may be due to the filling of existing soil macropores with biochar,  
400 which would shift the pore size distribution from large pores that drain quickly to pores that can retain water at  
401 field capacity (Liu et al., 2017). Biochar itself contains pores in the relevant size range (0.2—100  $\mu\text{m}$  in  
402 diameter) to contribute to  $\theta_{paw}$ . Thus, inter-particle pores in biochar will also contribute to  $\theta_{paw}$  depending on the  
403 size distribution and shapes of the biochar particles and their effects on soil aggregation (Burgeon et al., 2021).  
404 Since  $\theta_{fc}$  is the sum of  $\theta_{pwp}$  and  $\theta_{paw}$ , the same processes are the likely causes of the observed increases in  $\theta_{fc}$ .

405 The effects of biochar on water retention were in most cases larger for coarse-textured soils. Biochar with large  
406 microporosity can fill the larger inter-particle soil pores present in sandy soils so that the pore size distribution  
407 shifts towards the smaller pores that can retain water at the pressure potentials corresponding to field capacity  
408 (Omondi et al., 2016; Edeh et al., 2020; Rabbi et al., 2021). Moreover, fine-textured soils retain more water at  $\theta_{fc}$   
409 so that the relative changes induced by biochar may be smaller (Edeh et al., 2020).

410 All the meta-analyses included data on the effects of biochar production parameters (e.g. feedstock, pyrolysis  
411 temperature) and the chemical and physical properties of biochar. Generally, the influence of these parameters



412 on the effects of biochar addition were minor with respect to soil water retention. Due to lack of data, the  
413 influence of the time between biochar application and measurements on the effects on water retention was not  
414 included. It is, however, clear from studies on century-old charcoal kiln sites that the properties of biochar and  
415 associated soil evolve over time (e.g. Cheng et al., 2008; Hardy et al., 2017).

416 One meta-analysis reported an increase in saturated hydraulic conductivity following biochar addition (Omondi  
417 et al., 2016), while two others reported negative effects (figure 5). Saturated hydraulic conductivity is a function  
418 of pore network properties, including connectivity of the macropores and the presence of pore bottlenecks  
419 (Koestel et al., 2018). A few studies have quantified the effects of biochar addition on the connectivity of  
420 macropore networks using X-ray tomography (e.g. Yu and Lu, 2019; Yan et al., 2021). These studies indicate  
421 that the connected macroporosity and the diameter of pore throats decrease in medium- to coarse-textured soils  
422 amended with biochar. However, the influence of soil texture on the effects of biochar on saturated hydraulic  
423 conductivity reported in the meta-analyses is not consistent.

424 Two meta-analyses reported measures of aggregate stability. Omondi et al. (2016) included only studies that  
425 reported mean weight diameter (MWD) using wet sieving while Ul Islam et al. (2021) included studies that  
426 reported soil aggregate stability as a percentage of water-stable aggregates (WSA), as well as MWD or  
427 gravimetric mean diameter (GMD) using either wet sieving or dry sieving. Both studies showed that aggregate  
428 stability increased with biochar addition. The reason proposed for this effect was the influence of the added  
429 biochar on aggregation processes. The effects on aggregate stability increased with the time between biochar  
430 application and measurements (Ul Islam et al., 2021).

431 One meta-analysis focused on water use efficiency (Gao et al., 2020). They showed that both plant water use  
432 efficiency defined as the ratio of plant or fruit biomass to water supply, and leaf water use efficiency defined as  
433 the ratio of CO<sub>2</sub> uptake by leaves to the loss of water through transpiration, increased with biochar addition.

434 Most biochars are alkaline and may increase soil pH, at least for acidic soils, which leads to improved conditions  
435 for plant growth. Biochars also contain nutrients that may become available for plant uptake. Indeed, previous  
436 studies have shown that the addition of biochar to nutrient poor acidic soils improves yields (Jeffery et al., 2017).

437 Interestingly, in the meta-analysis of Gao et al. (2020), soil pH had different effects on plant and leaf water use  
438 efficiency. Leaf water use efficiency increased most for soils with pH less than 7, while plant water use  
439 efficiency increased most for soils with pH above 8. The reasons for these results could not be determined from  
440 their study. Medrano et al. (2015) showed that leaf water use efficiency may not be well correlated with plant



441 water use efficiency. The large variability in reported effects (-36–313%) can be mainly attributed to differences  
442 in soil pH, biochar properties and amounts of added biochar.

443 The effects of the studied variables were in many cases larger for higher application rates. Often, laboratory  
444 studies used much larger application rates (>50 t ha<sup>-1</sup>) than the field studies, probably for economic reasons. This  
445 may explain why effects usually were larger in laboratory or greenhouse experiments compared to field trials.  
446 Additionally, as pointed out by Rabbi et al. (2021), mixing of biochar after field applications is challenging and  
447 may be another reason why effects were sometimes small or insignificant for field experiments. The majority of  
448 the studies included in the meta-analyses were short-term experiments (i.e. duration < 1 year). Future work  
449 should therefore focus on longer-term effects of biochar applications under realistic field conditions. This  
450 requires either long-term field experimentation, which is expensive, or the study of historic biochar sites. It is  
451 also impractical to study these effects for the almost infinite number of combinations of soils, biochars and  
452 climates that exist. The meta-analyses included in this study, which show large variations in effects for all the  
453 included variables, suggest that future work should also be directed towards finding biochars with specific  
454 properties (e.g. surface area, particle size) designed to improve soil physical properties under specific soil and  
455 climate conditions while maintaining or improving nutrient availability.

#### 456 3.4.2 Other organic amendments, residue retention and mulching

457 Figure 5 shows that only a few meta-analyses have focused specifically on the effects of organic soil  
458 amendments or residue retention and mulching on soil properties relevant for water regulation functions. Instead,  
459 these practices are often included in meta-analyses on conservation agriculture or tillage systems. In these  
460 studies, the effects of the treatments are combined. Furthermore, the influence of contrasting soils or climates has  
461 not been assessed.

462 Bai et al. (2018) studied the effects of different organic amendments applied in long-term field experiments on  
463 soil physical and hydraulic properties. They found that aggregate stability increased with organic amendments  
464 and that this effect was largest for compost. However, this beneficial effect decreased with time. Not  
465 surprisingly, Bai et al. (2018) also reported greater aggregate stability under organic farming systems compared  
466 with conventional agriculture. Xiong et al. (2018) also included soil amendments applied to agricultural land in  
467 their global meta-analysis on soil conservation practices. They found that application of soil amendments  
468 reduced both surface runoff and soil erosion. However, they provided no information on the type of amendments  
469 and how they were incorporated into soil. Gravuer et al. (2019) analysed effects of organic amendments



470 (manure, biosolids and compost) applied to arid, semi-arid and Mediterranean rangelands. They found increased  
471 water contents at field capacity and reduced surface runoff. Additional benefits were increased soil organic  
472 carbon contents and above-ground net primary productivity, while trade-offs were increased CO<sub>2</sub> emissions,  
473 increased soil lead concentrations and increased losses of N and P in surface runoff.

474 Mulching means to add (or retain) material on the soil surface without incorporation (Kader et al., 2017). In this  
475 review, we focus on organic mulches, but synthetic materials are also used. The most extreme example of  
476 mulching with artificial materials is plastic mulching, which has been shown to increase crop water efficiency  
477 under drought (Yu et al., 2021). The use of organic amendments may have several beneficial effects on soil  
478 quality and the environment and is therefore one important practice in conservation agriculture. Mulching is  
479 typically carried out to limit soil evaporation, reduce soil runoff and erosion but it also affects, among other  
480 things, nutrient cycling, weed infestations and soil carbon storage (Ranaivoson et al., 2017). Mulching was  
481 included as one driver in four meta-analyses that studied effects on soil hydraulic functions. These meta-analyses  
482 showed positive effects on the rather limited number of hydraulic properties included. Three meta-analyses (one  
483 for agricultural land (Xiong et al., 2018), one for non-perennial crops (Ranaivoson et al., 2017) and one focusing  
484 only on tree crops (Liu et al., 2021) included effects of mulching on surface runoff. They all showed reduced  
485 surface runoff. The study for non-perennial crops also showed reduced soil evaporation and increased  
486 infiltration. These effects are already well-established in the scientific literature and in line with the intentions of  
487 mulching (Kader et al., 2017). The meta-analysis by Li et al. (2019) focused on effects of different tillage  
488 practices. Here we include the comparison between residue retention and no residue retention in no-till systems.  
489 Li et al. (2019) showed that residue retention led to a decrease in bulk density, an increase in total porosity and  
490 an increase in plant available water while it did not have significant effects on saturated hydraulic conductivity.  
491 They attributed this to increased accumulation of organic material on the soil surface, which leads to increased  
492 biological activity and soil aggregation.

### 493 **3.5 Irrigation**

494 Several recent meta-analyses have investigated the impacts of so-called deficit irrigation on water use efficiency  
495 and/or yields of a range of agricultural crops (Qin et al., 2016; Adu et al., 2018; Lu et al., 2019; Yu et al., 2020;  
496 Cheng et al., 2021a,b). The objective of this approach to irrigation scheduling is to reduce water use without  
497 significantly impacting yields by limiting the supply of water during periods of the growing season when it is  
498 less critical for crop growth. One of these meta-analyses (Cheng et al., 2021a) also synthesized the results of  
499 studies investigating the effects of partial root zone irrigation on water use efficiency and crop yields. This



500 method also has the objective of saving water without impacting yields, but in this case by alternately supplying  
501 water to only one part of the root zone at each irrigation. These meta-analyses show that although these irrigation  
502 scheduling methods either have mostly neutral or sometimes positive effects on crop water use efficiency (figure  
503 5), crop yields are significantly smaller compared to full irrigation for almost all crops and soil types. This  
504 implies that crop yields may in some cases be reduced less than water consumption, although these water savings  
505 may not compensate farmers for their yield losses. Another way to conserve high quality fresh water resources is  
506 to make use of brackish or saline water for irrigation. In their meta-analysis, Cheng et al. (2021c) showed how  
507 the decreases in water productivity, irrigation use efficiency and crop yields as a result of the use of salty  
508 irrigation water (figure 5) depends on crop type, irrigation methods, climates and soil type.

509 Based on the results of a meta-analysis, Qin et al. (2016) suggested that eliminating over-optimal (excess)  
510 irrigation by more efficient irrigation scheduling would improve the water use efficiency of citrus by 30% and  
511 yields by 20%. Du et al. (2018) showed that so-called aerated irrigation increases water use efficiency and yields  
512 of cereals and vegetables by ca. 20%, presumably by eliminating the development of anoxic conditions  
513 following irrigation.

#### 514 **4 Conclusions**

515 A large number of meta-analyses have been published in recent years on the impacts of soil and crop  
516 management practices on soil properties and processes and the various ecosystem services and functions  
517 delivered by soil. In this report, we have synthesized these analyses with respect to the water regulation functions  
518 that are relevant for climate change adaptation in Europe. This synthesis has revealed a considerable degree of  
519 consensus concerning the effects of soil and crop management practices, despite the fact that meta-analyses  
520 cannot easily account for differences in experimental conditions among individual source studies, not least  
521 because many primary studies do not report all details of the experimental treatments. This overview has also  
522 identified several important knowledge gaps, particularly related to the effects of management practices on root  
523 growth and transpiration. Thus, conclusions related to the impacts of management on the crop water supply are  
524 necessarily based on inferences derived from proxy variables such as available water capacity and infiltration  
525 capacity.

526 Meta-analyses have demonstrated that the use of organic amendments and the adoption of cropping systems and  
527 practices that maintain, as far as possible, “continuous living cover” both result in significant beneficial effects  
528 for the water regulation function of soils, arising from the additional carbon inputs to soil and the stimulation of



529 biological processes. These effects are clearly related to improvements in soil structure, both in terms of stable  
530 aggregation at the micro-scale and enhanced bio-porosity, both of which reduce surface runoff and increase  
531 infiltration. Meta-analyses show that amendment of soils with biochar generally increases aggregate stability,  
532 reduces bulk density, increases porosity and improves the plant available water capacity, particularly for coarse-  
533 textured soils. One potentially negative consequence of management practices that maintain “continuous living  
534 cover” is a reduction in soil water storage and groundwater recharge that, in most cases, will likely outweigh any  
535 increases in soil water storage capacity due to carbon sequestration. This may be problematic in dry climates,  
536 where there is evidence to suggest that yields of the main crop may be affected. With respect to environmental  
537 quality, no other significant trade-offs are known, while some important synergies have been identified, in  
538 particular reductions in nitrate leaching to groundwater and greenhouse gas emissions.

539 There is little evidence from meta-analyses to support the idea that reductions in tillage intensity improve crop  
540 water supply. The effects of no-till on SOM stocks and thus the capacity of the soil to store plant-available water  
541 appear to be minimal. In contrast, the amelioration of soil structure that occurs under RT and NT practices may  
542 improve infiltration capacity and reduce surface runoff, despite the increases in bulk density that are commonly  
543 reported, although the evidence for this is inconclusive. Some significant trade-offs with RT and NT systems  
544 have also been identified. For example, yield penalties incurred under NT and increased weed pressure and/or  
545 increased herbicide use and thus leaching risks, especially in wetter and colder climates, constitute a barrier to  
546 adoption by farmers. Furthermore, greenhouse gas emissions are generally larger under NT, while leaching  
547 losses to groundwater of both nitrate and pesticides may also increase. Although we might expect losses of agro-  
548 chemicals in surface runoff to generally decrease under RT and NT, thereby compensating for greater leaching  
549 losses, this does not always appear to be the case. Reduced tillage intensity in the temporal sense (i.e.  
550 “occasional” tillage) may help to ameliorate some of the negative effects of no-till systems, whilst retaining  
551 some of the advantages.

552



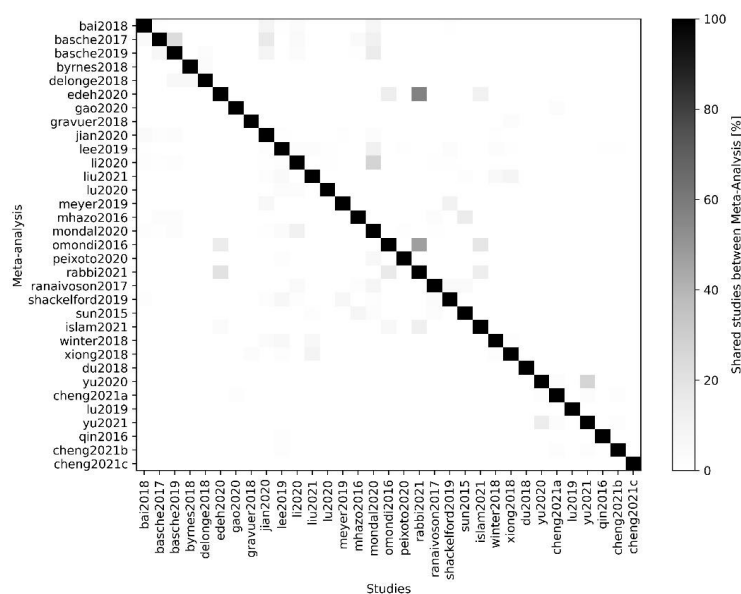
553 **Appendices**

554 **Appendix 1 Redundancy analysis**

555 Note that for this analysis, the studies of Li et al. (2019) and Li et al. (2020) were considered as one, as they both  
556 rely on the same database, but analyze different variables. We first identified the studies shared between multiple  
557 meta-analyses and computed the percentage of shared studies per meta-analysis. Figure A1 shows the percentage  
558 of shared studies (number of shared studies divided by number of studies in the meta-analysis in the row times  
559 100). A2 shows for each meta-analysis the number of source studies that it shares with at least one other meta-  
560 analysis. Some meta-analyses share nearly 100% of their studies with another meta-analysis (e.g., Omondi et al.,  
561 2016, Edeh et al., 2020). In addition to the extent of redundancy, A2 also shows the number of primary studies  
562 included in each meta-analysis. For example, Jian et al. (2020), Li et al. (2020) and Mondal et al. (2020) considered  
563 more than 200 primary studies in their meta-analyses. Finally, A3 shows for each meta-analysis the percentage of  
564 its primary studies that are shared with another meta-analysis. For example, the studies by Omondi et al. (2016)  
565 and Rabbi et al. (2021) share a large proportion of primary studies. A3 also shows that nearly all the primary  
566 studies included in these two meta-analyses are shared with another meta-analysis.

567

568 Redundancy matrix showing the percentage of shared studies among meta-analyses. The percentage refers to the  
569 number of shared studies divided by the total number of studies in the meta-analysis in the row. Note that this  
570 matrix is not symmetrical, because the percentage is computed for the meta-analysis in the row. If we had shown  
571 the number of shared studies as a number and not a percentage, this matrix would have been symmetrical.



572

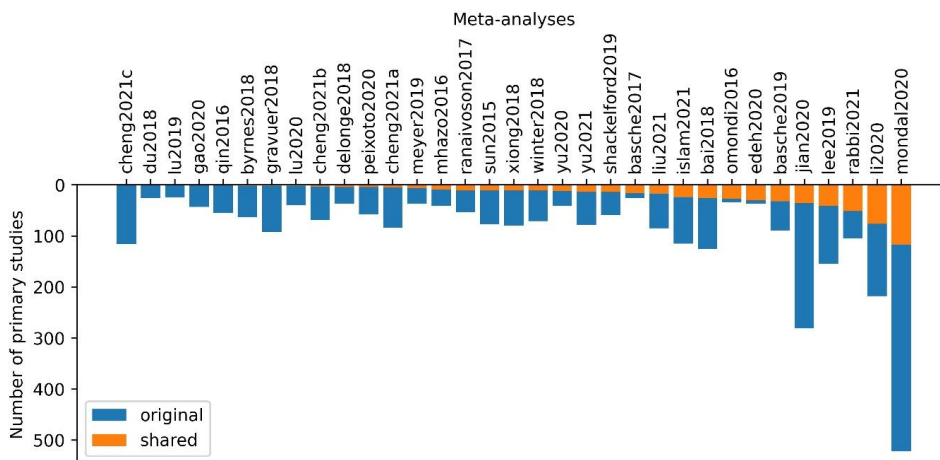
573 **Figure A1: Redundancy matrix showing the percentage of shared studies among meta-analyses.**

574



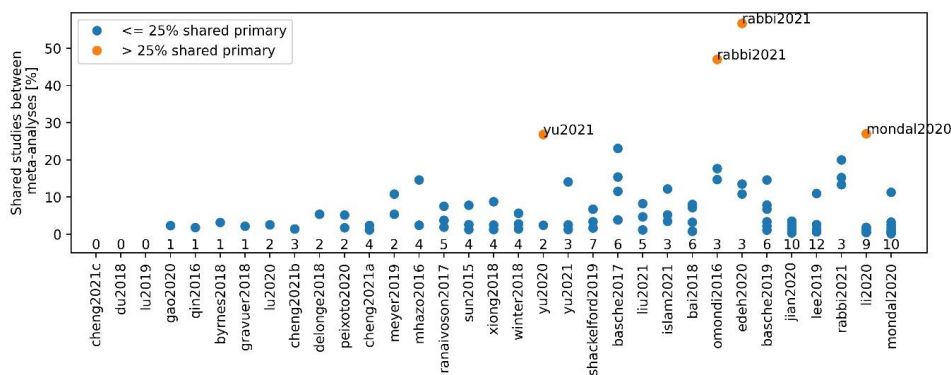


575 Histogram showing the number of studies per meta-analysis. The studies shared by at least one other meta-analysis  
 576 are displayed in light green (shared), while the studies found only in this meta-analysis are shown in dark green  
 577 (original).



578  
 579 **Figure A2: Histogram showing the number of studies per meta-analysis.**

580  
 581 Redundancy among the selected meta-analyses (horizontal axis). Dots represent the percentage of shared primary  
 582 studies between two meta-analyses. When this percentage is above 25%, the dots are shown in red, and the name  
 583 of the meta-analysis is displayed. For instance, Li et al. (2020) shares more than 25% of its primary studies with  
 584 the meta-analysis of Mondal et al (2020). The number on the horizontal axis denotes the number of other meta-  
 585 analyses that share primary studies with the meta-analysis named horizontally. Note that several meta-analysis do  
 586 not share any studies with others. Meta-analysis are sorted according to the amount of shared primary studies they  
 587 have (same order as A2).



588  
 589 **Figure A3: Redundancy among the selected meta-analyses (horizontal axis). Dots represent the percentage**  
 590 **of shared primary studies between two meta-analyses.**

591



- 592           **Appendix 2 Meta-analyses on soil/crop management and water regulation in the EU**
- 593   Bai, Z., Caspari, T., Gonzalez, M., Batjes, N., Mäder, P., Bünemann, E., de Goede, R., Brussaard, L., Xu, M.,  
594    Ferreira, C., Reintam, E., Fan, H., Mihelič, R., Glavan, M., Tóth, Z. 2018. Effects of agricultural management  
595    practices on soil quality: a review of long-term experiments for Europe and China. *Agriculture, Ecosystems*  
596    and Environment, 265, 1-7.
- 597   Basche, A., deLonge, M. 2017. The impact of continuous living cover on soil hydrologic properties: a meta-  
598    analysis. *Soil Science Society of America Journal*, 81, 1179-1190.
- 599   Basche, A., deLonge, M. 2019. Comparing infiltration rates in soils managed with conventional and alternative  
600    farming methods: a meta-analysis. *PLOS One*, 14, e0215702.
- 601   Byrnes, R., Eastburn, D., Tate, K., Roche, L. 2018. A global meta-analysis of grazing impacts on soil health  
602    indicators. *Journal of Environmental Quality*, 47, 758-765.
- 603   Cheng, M., Wang, H., Fan, J., Zhang, S., Liao, Z., Zhang, F., Wang, Y. 2021a. A global meta-analysis of yield  
604    and water use efficiency of crops, vegetables and fruits under full, deficit and alternate partial root-zone  
605    irrigation. *Agricultural Water Management*, 248, 106771.
- 606   Cheng, M., Wang, H., Fan, J., Zhang, S., Wang, Y., Li, Y., Sun, X., Yang, L., Zhang, F. 2021b. Water productivity  
607    and seed cotton yield in response to deficit irrigation: A global meta-analysis. *Agricultural Water*  
608    Management, 255, 107027.
- 609   Cheng, M., Wang, H., Fan, J., Wang, X., Sun, X., Yang, L., Zhang, S., Xiang, Y., Zhang, F. 2021c. Crop yield  
610    and water productivity under salty water irrigation: A global meta-analysis. *Agricultural Water Management*,  
611    256, 107105.
- 612   DeLonge, M., Basche, A. 2018. Managing grazing lands to improve soils and promote climate change adaptation  
613    and mitigation: a global synthesis. *Renewable Agriculture and Food Systems*, 33, 267-278.
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### 673 **Code availability**

674 Code is available at <https://github.com/climasoma/review-of-meta-analyses>

### 675 **Data availability**

676 Data is available at <https://github.com/climasoma/review-of-meta-analyses>

### 677 **Author contribution**

678 Conceptualization: all co-authors

679 Data collation and analysis: GuB, GiB, NJ

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681 Writing, review and editing: all co-authors

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