

**Soil and crop management practices and the water regulation  
functions of soils: a qualitative synthesis of meta-analyses  
relevant to European agriculture**

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1 **Abstract.** Adopting soil and crop management practices that conserve or enhance soil structure is critical for  
2 supporting the sustainable adaptation of agriculture to climate change, as it should help maintain agricultural  
3 production in the face of increasing drought or water excess without impairing environmental quality. In this paper,  
4 we evaluate the evidence for this assertion by synthesizing the results of 34 published meta-analyses of the effects  
5 of such practices on soil physical and hydraulic properties relevant for climate change adaptation in European  
6 agriculture. We also review an additional 127 meta-analyses that investigated synergies and trade-offs or help to  
7 explain the effects of soil and crop management in terms of the underlying processes and mechanisms. Finally, we  
8 identify how responses to alternative soil-crop management systems vary under contrasting agro-environmental  
9 conditions across Europe. This information may help practitioners and policymakers to draw context-specific  
10 conclusions concerning the efficacy of management practices as climate adaptation tools.

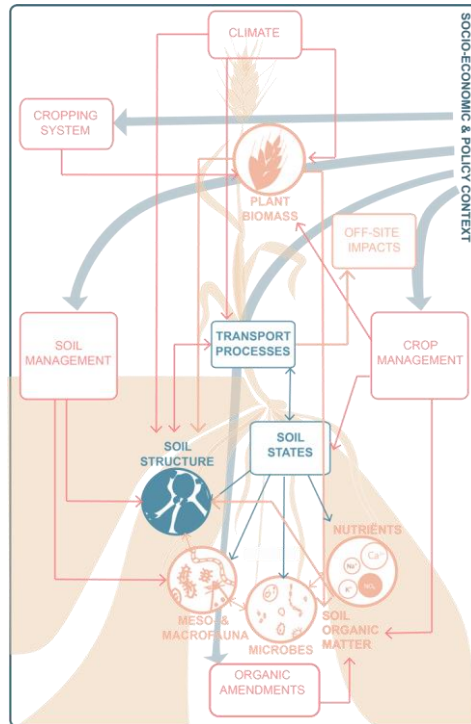
11 Our synthesis demonstrates that organic soil amendments and the adoption of practices that maintain “continuous  
12 living cover” result in significant benefits for the water regulation function of soils, mostly arising from the  
13 additional carbon inputs to soil and the stimulation of biological processes. These effects are clearly related to  
14 improved soil aggregation and enhanced bio-porosity, both of which reduce surface runoff and increase infiltration.  
15 One potentially negative consequence of these systems is a reduction in soil water storage and groundwater  
16 recharge, which may be problematic in dry climates. Some important synergies are reductions in nitrate leaching  
17 to groundwater and greenhouse gas emissions for non-leguminous cover crop systems. The benefits of reducing  
18 tillage intensity appear much less clear-cut. Increases in soil bulk density due to traffic compaction are commonly  
19 reported. However, biological activity is enhanced under reduced tillage intensity, which should improve soil  
20 structure, infiltration capacity, and reduce surface runoff and the losses of agro-chemicals to surface water.  
21 However, the evidence for these beneficial effects is inconclusive, while significant trade-offs include yield  
22 penalties and increases in greenhouse gas emissions and the risks of leaching of pesticides and nitrate.

23 [Our synthesis also highlights important knowledge gaps on the effects of management practices on root growth](#)  
24 [and transpiration. Thus, conclusions related to the impacts of management on the crop water supply and other](#)  
25 [water regulation functions are necessarily based on inferences derived from proxy variables. Based on these](#)  
26 [knowledge gaps, we formulated several key avenues for future research on this topic.](#)

## 27 1 Introduction

28 ~~As a consequence of on-going climate change, the~~ occurrence of extreme weather events (~~i.e. heatwavesigh~~  
29 ~~temperatures, summer droughts, waterlogging and flooding); such as those experienced during the recent summers~~  
30 ~~of 2018, 2021 and 2022~~ will almost ~~probab~~certainly increase in ~~allmany~~ parts of Europe (~~AgriAdapt, 2017,~~  
31 ~~Jacobs et al., 2019; IPCC 2021~~). Climate change impacts on agriculture are projected to result in an average 1%  
32 loss of gross domestic product by 2050, but with large differences among regions and farming systems (Jacobs *et*  
33 *al.*, 2019). An urgent task is therefore to develop guidance on soil and crop management practices that would help  
34 farmers in all regions of Europe adapt to these extreme weather situations.

35 The ecosystem services a soil can deliver depend profoundly on its structure, which we define here as the spatial  
36 arrangement of the soil pore space. Mediated by various biological (e.g., faunal and microbial activity) and  
37 physical processes (e.g., traffic compaction, wet-dry and freeze-thaw cycles), soil structure is constantly evolving  
38 at time scales ranging from seconds to centuries, driven by weather patterns as well as changes in climate and land  
39 management practices (figure 1). In turn, soil structure strongly affects all life in soil as well as the balance between  
40 infiltration and surface runoff, as well as drainage and soil water retention and therefore the supply of water and  
41 nutrients to crops. ~~Agricultural practices can affect soil structure directly (e.g., compaction due to use of heavy~~  
42 ~~machinery) or indirectly (e.g. improved soil structure through increased bioturbation by earthworms after addition~~  
43 ~~of organic matter to the soil).~~ Practices commonly adopted in “conservation agriculture” (Palm *et al.*, 2014) are  
44 thought to enhance soil structure and should therefore help to maintain agricultural production in the face of severe  
45 droughts or heavy rain. Conservation agriculture to improve soil structure rests on three fundamental principles  
46 (Palm, *et al.*, 2014): i.) minimizing mechanical soil disturbance, ii.) maintaining soil cover by plants as much as  
47 possible and for as long as possible (i.e. aspects of both spatial and temporal coverage), and iii.) diversifying  
48 cropping. Other more recently coined and partially related terms are “regenerative agriculture”, which  
49 acknowledges past failures to preserve soil health (Schreefel *et al.*, 2020) and “climate-smart agriculture”, which  
50 is defined by FAO (2010) as “... *agriculture that sustainably increases productivity, enhances resilience, reduces*  
51 *greenhouse gases, and enhances achievement of national food security and development goals*”.



52

53 **Figure 1. Schematic diagram of drivers, agents and processes governing the dynamics of soil structure and**  
 54 **its effects on the soil-plant system.**

55 The effects of soil and crop management practices on soil properties, soil hydrological and biological functioning  
 56 and crop performance have been studied in many long-term field trials throughout the world. In addition to  
 57 narrative reviews (e.g. Palm et al., 2014), many quantitative meta-analyses synthesizing the findings of individual  
 58 experiments have also been published. This is especially the case in the last few years (Beillouin et al., 2019a,b),  
 59 probably because the number of field experiments that have been running for a sufficient length of time has only  
 60 recently reached the critical mass required to enable these kinds of quantitative analyses. Indeed, the increase in  
 61 the number of meta-analyses published on topics related to conservation agriculture has been so dramatic that four  
 62 over-arching syntheses of these meta-analyses have also recently been published. Bolinder et al. (2020) evaluated  
 63 the effects of organic amendments and cover crops on soil organic matter (SOM) storage, while Schmidt et al.  
 64 (2021) focused on the effects of biochar on crop performance. Beillouin et al. (2019b) and Tamburini et al. (2020)  
 65 carried out even more ambitious and comprehensive reviews of meta-analyses of the effects of conservation  
 66 agriculture and crop diversification strategies on a wide range of ecosystem services. Tamburini et al. (2020)

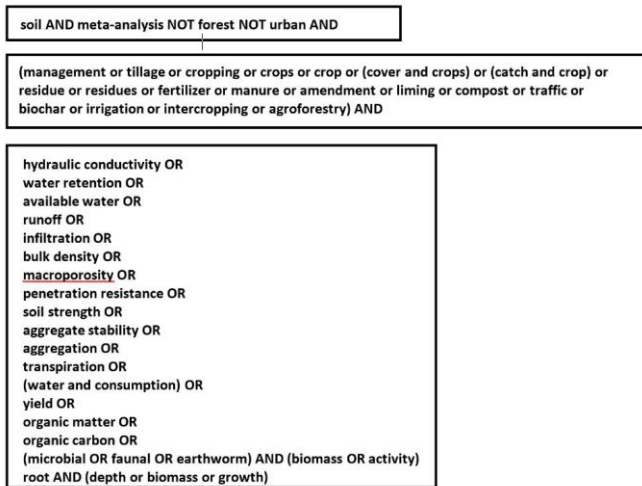
67 concluded that diversification practices most often resulted in a ‘win-win’ situation for ecosystem services  
68 including crop yields, but that the ~~often~~-large variability in responses and the occurrence of trade-offs highlighted  
69 the need to analyze the context-dependency of outcomes, something which was only possible to do to a limited  
70 extent with their broad-brush treatment. These previous syntheses of meta-analyses on the benefits of conservation  
71 agriculture have placed very little emphasis (Tamburini et al., 2020) or none at all (Beillouin et al., 2019a,b;  
72 Bolinder et al., 2020) on soil hydrological functioning even though this is key for climate change adaptation. In  
73 their synthesis, Tamburini et al. (2020) included 17 meta-analyses (involving 31 effects-size comparisons) relevant  
74 to water regulation, but most of these concerned water quality issues rather than hydrological functioning *per se*.  
75 Beillouin et al. (2019b) concluded that ... “*our review reveals that a significant knowledge gap remains, in*  
76 *particular regarding water use*”.

77 In this study, we focus on the implications of agricultural management practices for soil hydrological functioning  
78 for climate change adaptation under European agro-environmental conditions. We do this by identifying and  
79 synthesizing existing meta-analyses of the response of soil physical/hydraulic properties and hydrological  
80 processes relevant for climate change adaptation to soil and crop management practices. ~~This evaluation highlights~~  
81 ~~where consensus has been established and identifies remaining knowledge gaps.~~ In those cases where the  
82 information is available, we summarize knowledge of context-specific effects of relevance for the range of agro-  
83 environmental conditions found in Europe, and as far as possible, explain these variations in terms of individual  
84 driving processes and mechanisms. This kind of information may explain local praxis in agricultural management  
85 (i.e. farmer behavior) and will enable practitioners and policymakers to draw context-specific conclusions  
86 concerning the efficacy of management practices as climate adaptation tools. ~~This evaluation~~ ~~This study highlights~~  
87 ~~where consensus has been established on practices improving the water regulation function of soil that are~~  
88 ~~meaningful for in the light of climate change adaptation and.~~ ~~We also, identifyies~~ remaining knowledge gaps and  
89 ~~sets out~~ key avenues for future research.

## 90 **2 Materials and Methods**

### 91 **2.1 Literature search and data extraction**

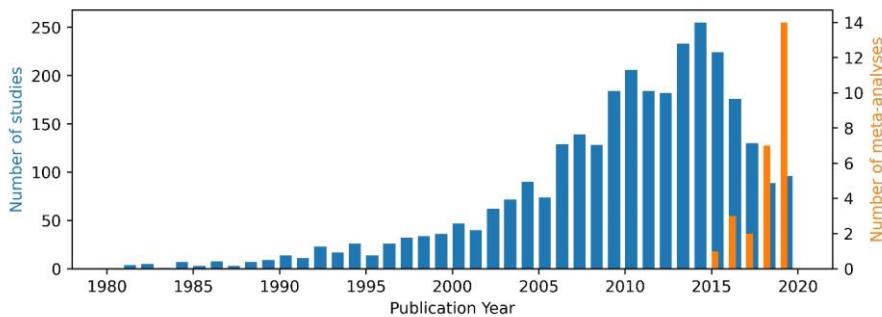
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93

94 **Figure 2. Search string used to identify relevant meta-analyses.**

95 The text string shown in Figure 2 was used to search the published literature using Web of Knowledge in May  
 96 2021. This search returned 663 results. All search results were manually assessed for their relevance to the  
 97 objectives of our study. Meta-analyses that only included studies carried out outside Europe were not retained.  
 98 Our search identified 34 relevant meta-analyses focusing on the effects of soil and crop management on soil  
 99 physical properties and hydrological processes using effects ratios (Appendix 2). Figure 3 shows the number of  
 100 primary studies per publication year included in the 34 meta-analyses. A peak is clearly visible in 2014, which is  
 101 explained by the fact that all the selected meta-analyses were published after 2015. Our search string was also  
 102 designed to identify meta-analyses of management effects on soil organic matter and biological variables (e.g.  
 103 microbial biomass), since these help to explain the observed effects on physical/hydraulic properties and  
 104 hydrological processes, as well as other studies that analyzed target variables representing potential “trade-offs”  
 105 or synergies. Among these, we focused primarily on the impacts of management practices on crop yields,  
 106 greenhouse gas emissions and water quality. An additional 127 published meta-analyses of this kind were  
 107 identified by our literature search. These studies are listed in the supplementary file (“*Supporting studies.xlsx*”).



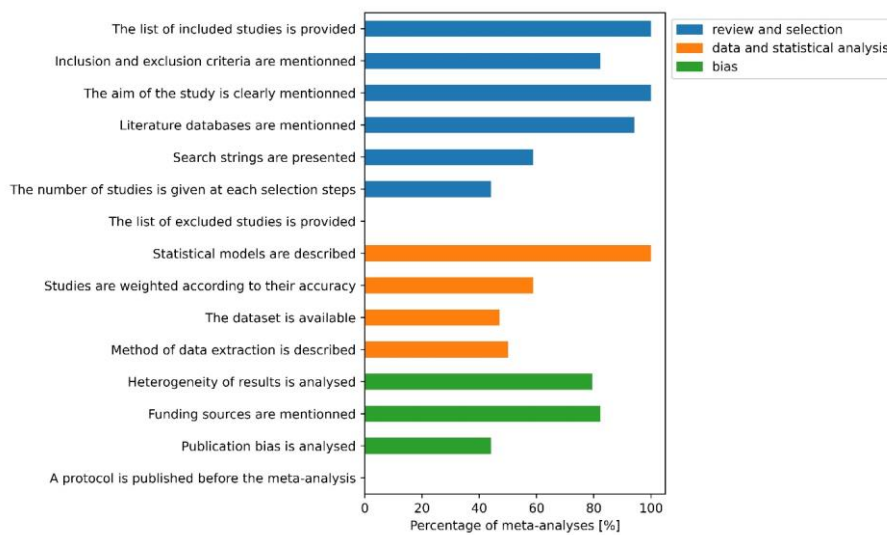
108  
 109 **Figure 3. Number of primary studies included in the 36 selected meta-analyses published per year and the**  
 110 **publication year of these meta-analyses.**

111 The target variables (e.g. soil physical and hydraulic properties) and drivers (i.e. soil and crop management  
 112 practices) included in the 34 meta-analyses were then classified into a limited number of groups. The target  
 113 variables were grouped into five classes: pore space properties (e.g. porosity, bulk density), hydraulic properties  
 114 (e.g. saturated hydraulic conductivity, field capacity), mechanical properties (e.g. soil aggregate stability,  
 115 penetration resistance), water flows (e.g. infiltration, surface runoff, drainage) and plant properties (e.g. root length  
 116 density, water use efficiency). Likewise, the management practices were also grouped into five classes: soil  
 117 amendments (e.g. manure, biochar, organic farming systems), cropping practices and systems (e.g. cover crops,  
 118 crop rotations), tillage systems (e.g. no-till), grazing management and irrigation. ~~In total, the 34 meta-analyses~~  
 119 ~~reported 104 effects ratios comparing the impacts of a management practice to a control treatment for a particular~~  
 120 ~~response variable. The effects of these treatments on the target variables (either positive, negative or neutral i.e.~~  
 121 ~~non significant) were read from tables and figures in each of the 34 meta-analyses. The directions of the effect~~  
 122 ~~sizes are purely statistical and have no connotation of value. We report the effects in a statistical sense because in~~  
 123 ~~some instances it not clear whether effect would be beneficial or detrimental.~~

124  
 125 **2.2 Quality assessment**

126 ~~We~~ We performed a quality assessment of the selected 34 meta-analyses using 15 of the criteria proposed by Beillouin  
 127 et al. (2019a). Figure 4 presents a summary of the quality of the selected meta-analyses according to these criteria.  
 128 Nearly half of the meta-analyses included datasets in the paper, while only ca. 44% investigated ~~the important~~  
 129 ~~issue of~~ publication bias (Philibert et al., 2012). The authors of these studies used simple statistical techniques such  
 130 as frequency distributions of effects sizes or “funnel plots” of sample sizes against effect sizes to investigate

131 whether experiments with non-significant effects are under-represented in the literature. For both of these methods,  
 132 symmetry of the distributions is taken to indicate a lack of bias. Two studies detected evidence of publication bias  
 133 (e.g. Basche and deLonge, 2019; Shackleford et al., 2019) using this method, although in both cases the effects on  
 134 the overall conclusions of the studies were considered marginal. Basche and deLonge (2019) also investigated the  
 135 sensitivity of the outcome to the exclusion of individual studies, which is another important aspect of publication  
 136 bias. They found ~~mostly~~ robust results for the impacts of ~~management practices on infiltration, especially for no-~~  
 137 ~~tillage and cover crops~~ on infiltration.



138

139 **Figure 4. Proportion of the quality criteria defined by Bellouin et al. (2019a) that are met by the selected**  
 140 **meta-analyses in this study.**

Met opmaak: Uitvullen

141 **2.3 Redundancy analysis**

142 We performed a redundancy analysis to identify the proportion of common primary or source studies among the  
 143 meta-analyses following the methodology of Beillouin et al. (2019a). For each of the 34 selected meta-analyses,  
 144 the references to the studies used were extracted from the supplementary materials. Each reference contained at  
 145 least the name of the first author, the year of publication, the title, the journal and – if available – the DOI. Of the  
 146 3142 unique primary studies, 437 had no DOI. Old publications or publications not written in English were usually  
 147 found to have no DOI. In some cases, the title and DOI were not available so we had to manually check these  
 148 references based on contextual information supplied in the supplementary material. In most cases, however, the  
 149 title was provided in the meta-analysis and the DOI could be extracted automatically from the Cross-Ref database.



150 We then manually checked if the title of the paper matched the one found on Cross-ref, to confirm the DOI  
151 assignment. The results of the redundancy analysis are presented in the Appendix 1 (Figures A1-A3) as well as in  
152 the notebook at <https://github.com/climasoma/review-of-meta-analyses/blob/main/notebooks/redundancy.ipynb>.

153 The main outcome of this analysis is that redundancy is only ~~an problematic~~important factor for a few meta-  
154 analyses on biochar that were published almost simultaneously (e.g. Edeh, et al., 2020; Rabbi, et al., 2021).

#### 155 **2.4 Qualitative analysis of effect sizes**

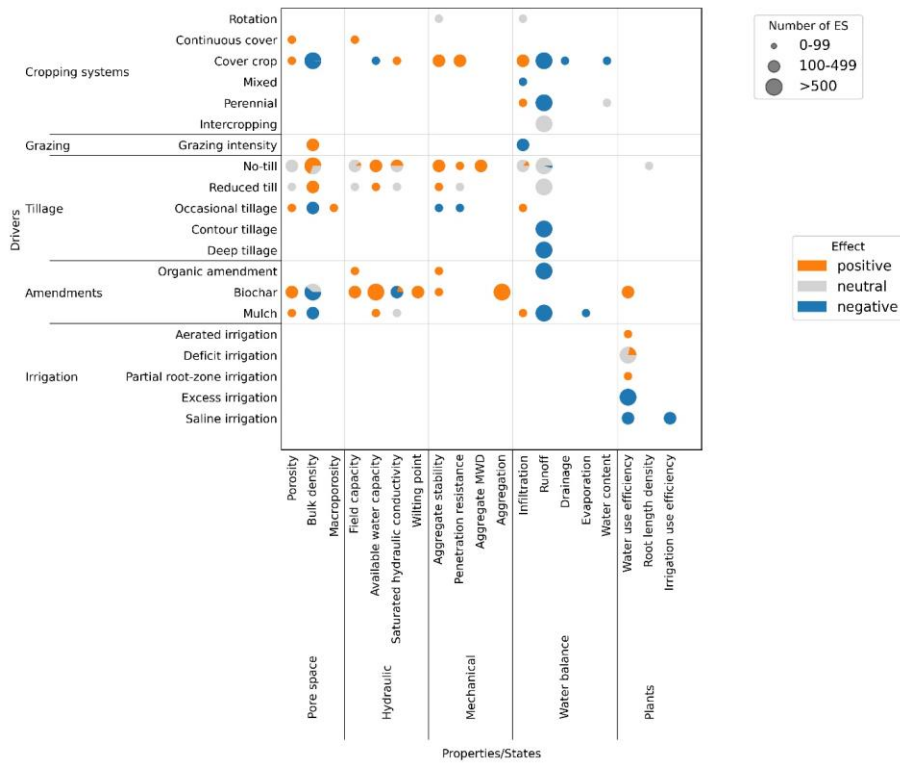
156 In total, the 34 meta-analyses reported 104 effects ratios comparing the impacts of a management practice to a  
157 control treatment for a particular response variable. The effects of these treatments on the target variables (either  
158 positive, negative or neutral i.e. non-significant) were read from tables and figures in each of the 34 meta-analyses  
159 and analyzed in a qualitative way. This is because we do not have access to the effects sizes in all the primary  
160 studies included in the meta-analyses (see section 2.2 “Quality assessment”). We considered the effect as  
161 “positive” if the average log response ratio and the entire 95% confidence interval reported in the meta-analysis  
162 was larger than zero (equivalent to a response ratio of 1 prior to taking logarithms). If part of the 95% confidence  
163 interval for the log response ratio overlapped zero, the effect was considered “neutral”. If the entire confidence  
164 interval was smaller than zero, the overall effect was considered as “negative”. The directions of the effect sizes  
165 are therefore purely statistical and have no connotation of value. We report the effects in a statistical sense because  
166 in some instances it would not be clear whether effects would be beneficial or detrimental. It should also be noted  
167 that positive or negative overall effects derived from the results presented in a given meta-analysis do not imply  
168 that all the individual effects in the primary studies included in this meta-analysis necessarily pointed in the same  
169 direction. For all overall effects retrieved, we also noted the number of individual effects sizes from the primary  
170 studies used to compute the overall effect reported in the meta-analysis.

### 171 **3 Results and discussion**

#### 172 **4.3 Knowledge gaps**

173 Figure 5 summarizes the statistical relationships found between the drivers and target variables in the selected  
174 meta-analyses. It shows that the effects of cropping systems, tillage, organic amendments and, to a lesser extent,  
175 irrigation management have been studied extensively. These topics are discussed in the following sections. It is  
176 equally interesting to consider the empty zones in Figure 5, which represent topics for which existing experimental  
177 data has not yet been summarized or which have been the focus of only a few studies in the past. We discuss these  
178 knowledge gaps in section 3.5. Finally, we use the outcome of our analysis to formulate some key avenues for  
179 future research on the extent to which management practices can reinforce the water regulation function of soils.

Gewijzigde veldcode



180

181 **Figure 5. Effects of drivers (vertical axis) on target variables (horizontal axis) in the 36 selected meta-**  
 182 **analyses. The coloured pie charts represent the directions of the statistical effects in the different meta-**  
 183 **analyses, while the size of the circle indicates the total number of effects sizes (ES) reported. Note that this**  
 184 **number has not been corrected for redundancy. Blank cells denote that no data was available for this target**  
 185 **variable in any of the selected meta-analyses.**

186

187 Figure 5 shows which practices have been studied extensively and for some, this results in scientific consensus on  
 188 the effects of those practices on our target soil variables of interest. We identified three clear consensus areas  
 189 which we discuss in the following sections: (1) cropping systems, (2) tillage, (3) amendment. It is equally  
 190 interesting to investigate the empty zones in Figure 5. These zones represent the areas in which existing knowledge  
 191 has not yet been summarized or which have been in the focus of only few studies in which simply little studies  
 192 have been conducted in the past. We discuss these knowledge gaps in a separate subsection 3.5. Finally, we use  
 193 our analysis to formulate some key avenues for future research on how management practices can reinforce the  
 194 water regulation function of soils.

Met opmaak: Uitvullen

195 **4.13.1 Cropping systems and practices**

196 Broadly speaking, published meta-analyses that have investigated the effects of cropping systems and practices  
197 (figure 5) fall into two categories: i.) studies analyzing the effects of maintaining a more continuous soil surface  
198 cover, either in a temporal (e.g. cover crops in arable rotations) or in a spatial sense (e.g. inter-row cover in widely-  
199 spaced row crops such as vineyards and orchards), and ii.) studies comparing farming systems (e.g. continuous  
200 arable contrasted with either perennial crops or rotations or mixed farming systems with livestock). In the  
201 following, we combine these two aspects, referring to both of them as cropping systems that as far as possible  
202 maintain a “continuous living cover” (Basche and deLonge, 2017).

203 Figure 5 shows that meta-analyses have identified several beneficial effects of such agronomic practices on  
204 important physical and hydraulic properties in soil, such as porosity or bulk density, saturated hydraulic  
205 conductivity and aggregate stability (Basche and deLonge, 2017; Jian et al., 2020). These positive effects are  
206 almost certainly due to a combination of the protective effects of surface cover against the degradation of soil  
207 structure by raindrop impact as well as the enhancement of various biological processes that occurs as a  
208 consequence of plant growth, root production and the additional carbon supply to the soil. In this respect, meta-  
209 analyses have demonstrated that practices that maintain a continuous living cover (e.g. rotations with leys, cover  
210 crops) promote increases in microbial biomass, activity and diversity (Venter et al., 2016; Shackleford et al., 2019;  
211 Jian et al., 2020; Kim et al., 2020; Muhammad et al., 2021) and increase soil organic matter contents in the long-  
212 term (Aguilera et al., 2013; McDaniel et al., 2014; Poeplau and Don, 2015; King and Blesh, 2018; Bai et al., 2019;  
213 Shackleford et al., 2019; Bolinder et al., 2020; Jian et al., 2020; McClelland et al., 2021). This will both promote  
214 stable soil aggregation and reduce soil bulk density (Chenu et al., 2000; Meurer et al., 2020a,b). The abundance of  
215 soil meso- and macro-fauna also increases under long-term cover cropping (Reeleder et al., 2006; Roarty et al.,  
216 2017) and perennial crops such as grass/clover leys (Fraser et al., 1994; Bertrand et al., 2015; Jarvis et al., 2017).  
217 Through their burrowing activity, these “ecosystem engineers” (Jones et al., 1994) create networks of large  
218 biopores in soil (Jarvis, 2007) that greatly increase saturated and near-saturated hydraulic conductivity and thus  
219 infiltration capacity (e.g. Bertrand et al., 2015; Capowiez et al., 2021).

220 The changes in soil physical and hydraulic properties brought about by the introduction of continuous living cover  
221 have significant beneficial consequences for the water regulation function of soils. Thus, cover crops enhance  
222 infiltration capacity and reduce surface runoff (Xiong et al., 2018; Basche and deLonge, 2019; Lee et al., 2019;  
223 Jian et al., 2020; Liu et al., 2021). An increased proportion of perennial crops in the rotation and the presence of  
224 ground cover between the rows of perennial crops (e.g. in vineyards) increase soil infiltration and reduce surface

225 runoff (Xiong et al., 2018; Basche and deLonge, 2019; Liu et al., 2021). These positive effects seem broadly  
226 similar regardless of climate (Xiong et al., 2018; Liu et al., 2021). As noted above, “continuous living cover”  
227 increases soil organic matter contents and both long-term field experiments and meta-analyses suggest that soil  
228 organic matter generally tends to increase the plant available water capacity. However, although the magnitude of  
229 this effect is still a matter of debate (Lal, 2020), in most cases it seems relatively small compared with the crop  
230 water demand (Minasny and McBratney, 2018a,b; Libohova et al., 2018). One potential negative effect of cropping  
231 systems employing “continuous living cover” is that increased transpiration may reduce soil water contents  
232 (Shackelford et al., 2019) and decrease recharge to groundwater. Thus, for a combined dataset of 36 studies  
233 comprising both experimental and modelling studies, Meyer et al. (2019) found that cover crops reduced recharge  
234 by 27 mm/year on average with no apparent effects of climate, soil type or cropping system. For their meta-analysis  
235 based on a more limited dataset of six studies, Winter et al. (2018) found no significant effects of inter-row  
236 vegetation in vineyards on the soil water balance as compared to a bare inter-row strips.

237 Some negative consequences of mixedMixed farming systems with grazing livestock for soil physical properties  
238 have been noted. Meta-analyses have shown that high grazing intensities can result in significantly poorer soil  
239 physical quality, in terms of larger bulk densities (Byrnes et al., 2018) and reduced infiltration rates (deLonge  
240 and Basche, 2018; Basche and deLonge, 2019) as a result of compaction by animal trampling. These impacts of  
241 intensive grazing are similar irrespective of soil texture or climate, although they appear to be are slightly larger  
242 in wetter climates (Byrnes et al., 2018; deLonge and Basche, 2018).

243 Another significant negative effect is that by increasing transpiration, systems employing “continuous living  
244 cover” may reduce soil water content (Shackelford et al., 2019) and decrease recharge to groundwater. Thus, for  
245 a combined dataset of 36 studies comprising both experimental and modelling studies, Meyer et al. (2019) found  
246 that cover crops reduced recharge by 27 mm/year on average with no apparent effects of climate, soil type or  
247 cropping system. For their meta-analysis based on a more limited dataset of six studies, Winter et al. (2018) found  
248 no significant effects of inter-row vegetation in vineyards on the soil water balance as compared to a bare interrow  
249 strip. Other impacts of cover crops on physical properties with adverse consequences for crop water supply have  
250 been reported, for example an increase in soil penetration resistance and a reduced available water capacity (Jian  
251 et al 2020), although it seems quite difficult to identify plausible mechanisms for such effects. As noted earlier,  
252 “continuous living cover” increases soil organic matter contents and both long-term field experiments and meta-  
253 analyses suggest that soil organic matter generally tends to increase the plant available water capacity. However,

254 although the magnitude of this effect is still a matter of debate (Lal, 2020), in most cases it seems relatively small  
255 compared with the crop water demand (Minasny and McBratney, 2018a,b; Libohova et al., 2018).

#### 256 *Synergies and trade-offs*

257 With respect to potential synergies and trade-offs, studies Meta-analyses have shown that cover crops mostly have  
258 either neutral or positive effects on main crop yields (Tonitto et al., 2006; Quemada et al., 2013; Valkama et al.,  
259 2015; Angus et al., 2015; Marcillo and Miguez, 2017). However, Shackleford et al. (2019) reported an average  
260 7% reduction in cash crop yields for systems employing non-legume cover crops in dry Mediterranean climate  
261 conditions. Similarly, in a recent meta-analysis on cover crops grown in climates with less than 500 mm annual  
262 rainfall, Blanco-Canqui et al. (2022) found that cover crops decreased main crop yields in 38% of cases, with no  
263 effects found in 56% of cases and increased yields in 6% of cases. Non-leguminous cover crops significantly  
264 reduce nitrate leaching and, to a lesser extent, N<sub>2</sub>O emissions, although this is clearly not the case for legumes  
265 (Tonitto et al., 2006; Quemada et al., 2013; Basche et al., 2014; Valkama et al., 2015; Muhammad et al., 2019;  
266 Shackleford et al., 2019). Our literature search did not identify any meta-analyses on phosphorus or pesticide  
267 losses.

#### 268 **4.23.2 Tillage systems**

269 ~~A large number of meta-analyses have investigated the~~The effects of tillage practices on soil properties and  
270 ~~functions and the provision of various ecosystem services have been widely investigated~~ (figure 5). The control  
271 treatment in ~~these published~~ meta-analyses is usually conventional tillage (CT), which involves both inversion  
272 ploughing and shallow secondary tillage operations for seedbed preparation. This control treatment is then  
273 contrasted with either reduced (or minimum) tillage (RT), whereby the soil is no longer ploughed, or no-till (NT)  
274 systems in which the soil is left completely undisturbed, or both. ~~One meta-analysis has investigated the effects of~~  
275 ~~deep tillage and contour tillage on surface runoff~~ (Xiong et al., 2018), while another analyzed the effects of  
276 ~~occasional tillage with no-till as the control treatment~~ (Peixoto et al., 2020). ~~Changes in tillage systems directly~~  
277 ~~affect the physical properties of soil. For example, bulk density and penetration resistance often increase after the~~  
278 ~~adoption of RT and NT systems~~ (Lee et al. 2019; Li et al., 2019; Li et al., 2020a) due to continued traffic  
279 ~~compaction of other field operations and the lack of mechanical loosening by cultivation~~ (Hamza and Anderson  
280 ~~2005). Peixoto et al. (2020) showed that these negative effects can be alleviated with occasional tillage.~~

281 Soil tillage indirectly affects soil structure through effects on soil macro-fauna. In addition to the direct impacts of  
282 tillage implements on mortality, disruption of the soil also exposes soil macro-fauna to increased risks of

283 desiccation and predation. Consequently, meta-analyses show that total earthworm biomass and abundance  
284 increase as tillage intensity is reduced (Spurgeon et al., 2013; Briones and Schmidt 2017; Bai et al. 2018), with a  
285 negative relationship between tillage depth and earthworm abundance (Briones and Schmidt 2017). Deep  
286 burrowing and surface-feeding (anecic) earthworm species are particularly favored by NT systems, as their  
287 permanent burrows are no longer destroyed by ploughing and they have a better access to food resources. Thus, a  
288 lack of disturbance of the soil by tillage has also been shown to increase the diversity of earthworm populations  
289 (Chan, 2001; Spurgeon et al., 2013; Briones and Schmidt 2017) and soil fauna in general (de Graaff et al., 2019).

290 Reductions in the depth and intensity of tillage (i.e. from CT to RT to NT) strongly influence carbon cycling in  
291 the soil-crop system. Several meta-analyses show that soil organic carbon concentrations are larger under RT and  
292 NT systems in the uppermost soil layers (e.g. Bai et al., 2018; Lee et al., 2019) especially in fine-textured soils (Bai  
293 et al., 2019). The reasons for this are the lack of soil disturbance that promotes a stable aggregated structure, which  
294 affords a greater physical protection of C against microbial mineralization (Kan et al., 2021) and the elimination  
295 of physical mixing and re-distribution of C within the topsoil due to the absence of soil inversion by ploughing  
296 (Meurer et al., 2020b). Meta-analyses have shown that the accumulation of SOM typically found in surface soil  
297 layers under RT and NT systems, which reflects the deposition and accumulation of plant residues, is paralleled  
298 by a greater microbial biomass (e.g. Spurgeon et al., 2013; Zuber and Villamil, 2016; Li et al. 2018; Lee et al.  
299 2019; Li et al. 2020b,c; Chen et al. 2020) and increases in enzyme activities (Zuber and Villamil, 2016; Lee et al.,  
300 2019). The diversity of bacterial and sometimes also fungal communities tends to be greater in RT or NT (Spurgeon  
301 et al., 2013; de Graaff et al., 2019; Li et al 2020b), especially where these systems are combined with the retention  
302 of crop residues (Li et al., 2020c). Meta-analyses also show that aggregate stability is largest under NT systems,  
303 is intermediate when occasional tillage is practiced (Peixoto et al. 2020) and smallest in CT systems (Bai et al.  
304 2018). In their meta-analysis, Spurgeon et al. (2013) showed that improved aggregate stability under NT systems  
305 was positively correlated with increases in fungal biomass. NT also increases the mean size of aggregates produced  
306 in stability tests (Li et al., 2020a; Mondal et al., 2020). Several meta-analyses have demonstrated increases in field  
307 capacity and available water capacity under reduced and no-till systems (Li et al., 2019; Mondal et al., 2020; Li et  
308 al., 2020a) presumably due to enhanced soil biological activity and increases in organic carbon content.

309 The impacts of conservation tillage practices on soil biological agents and processes give rise to significant indirect  
310 effects on physical properties and hydrological processes. Saturated hydraulic conductivity and surface infiltration  
311 rates often increase under conservation tillage compared with CT, especially for NT systems (Li et al., 2019;  
312 Basche and deLonge, 2019; Li et al., 2020a; Mondal et al., 2020). This suggests that the effects of the enhanced

313 [bioporosity in NT systems created by soil fauna, and especially anecic earthworms, on saturated and near-saturated](#)  
314 [hydraulic conductivity \(Lee and Foster, 1991\) generally outweigh the negative effects of increased bulk density.](#)  
315 [Thus, Spurgeon et al. \(2013\) showed that increased earthworm abundances and diversity found under NT systems](#)  
316 [were positively correlated with infiltration rates. Comparing ecological groups, they found that the density of](#)  
317 [anecic earthworms was positively associated with increased infiltration rates, whereas no effect was apparent for](#)  
318 [endogeic earthworms. In principle, better-developed soil macropore systems and improvements in aggregate](#)  
319 [stability and infiltration capacity should promote a more favorable crop water balance, with reductions in surface](#)  
320 [runoff. Figure 5 shows that the effect on runoff is one of the most studied hydrological processes related to tillage.](#)  
321 [The meta-analysis performed by Sun et al. \(2015\) found that RT and NT systems decreased surface runoff.](#)  
322 [However, these results do not appear to be conclusive as two later meta-analyses \(Mhazo et al., 2016; Xiong et al.,](#)  
323 [2018\) failed to detect significant effects of conservation tillage practices on surface runoff. However, Xiong et al.](#)  
324 [\(2018\) found that contour tillage and deep tillage both reduced surface runoff.](#)

325 [Reductions in the depth and intensity of tillage \(i.e. from CT to RT to NT\) strongly influence carbon cycling in](#)  
326 [the soil-crop system. Several meta-analyses show that soil organic carbon concentrations are larger under RT and](#)  
327 [NT systems in the uppermost soil layers \(e.g. Bai et al., 2018; Lee et al., 2019\) especially in fine-textured soils](#)  
328 [\(Bai et al., 2019\). The reasons for this are the lack of soil disturbance that promotes a stable aggregated structure,](#)  
329 [which affords a greater physical protection of C against microbial mineralization \(Kan et al., 2021\) and the](#)  
330 [elimination of physical mixing and re-distribution of C within the topsoil due to the absence of soil inversion by](#)  
331 [ploughing \(Meurer et al., 2020b\). Meta-analyses have shown that the accumulation of SOM typically found in](#)  
332 [surface soil layers under RT and NT systems, which reflects the deposition and accumulation of plant residues, is](#)  
333 [paralleled by a greater microbial biomass \(e.g. Spurgeon et al., 2013; Zuber and Villamil, 2016; Li et al. 2018; Lee](#)  
334 [et al. 2019; Li et al. 2020b,e; Chen et al. 2020\) and increases in enzyme activities \(Zuber and Villamil, 2016; Lee](#)  
335 [et al., 2019\). The diversity of bacterial and sometimes also fungal communities tends to be greater in RT or NT](#)  
336 [\(Spurgeon et al., 2013; de Graaff et al., 2019; Li et al. 2020b\), especially where these systems are combined with](#)  
337 [the retention of crop residues \(Li et al., 2020e\).](#)

338 [In addition to focusing on organic carbon concentrations in topsoil, differences in SOC stocks under conservation](#)  
339 [tillage systems in complete crop root zones and soil profiles are also of interest, not least from the point of view](#)  
340 [of climate change mitigation. Based on a meta-analysis of studies with measurements made to at least 40 cm depth,](#)  
341 [Luo et al. \(2010\) concluded that NT did not increase soil carbon stocks. This is because although SOC contents](#)  
342 [are usually larger under NT than CT systems in the uppermost soil layers, they can be significantly smaller both](#)

343 at plough depth and in the upper subsoil (Angers and Eriksen-Hamel, 2008). Thus, for boreo-temperate climates,  
344 Haddaway et al. (2017) and Meurer et al. (2018) found increases in soil carbon stocks under NT compared to CT  
345 only in the topsoil, while no overall significant effect on carbon stocks was detected for soil profiles to 60 cm  
346 depth. In a more recent global meta-analysis, Mondal et al. (2020) found no significant differences in stocks of  
347 soil organic carbon between NT and CT systems, while variations in response could not be attributed to either  
348 climate or soil type. In apparent contrast, Mangalassery et al. (2015) concluded that NT systems result in a net  
349 sequestration of carbon, regardless of the depth of soil considered. Sun et al. (2020) demonstrated significant  
350 effects of climate on the changes in organic carbon stocks observed under NT systems. In their global analysis,  
351 they found that soil C sequestration was enhanced in warmer and drier regions, while soils under no-till in colder  
352 and wetter climates were just as likely to lose soil C as gain C. These findings are supported by the regional-scale  
353 studies of Meurer et al. (2018) for boreo-temperate climates and Gonzalez-Sanchez et al. (2012) and Aguilera et  
354 al. (2013) for Mediterranean climates, although for vineyards, Payen et al. (2021) found larger topsoil C  
355 sequestration in temperate climates than hot and dry climates. A loss of organic carbon following adoption of NT  
356 systems can be explained by a decrease in carbon inputs to soil resulting from poorer crop growth (Mangalassery  
357 et al., 2015; Pittelkow et al., 2015), which compensates for reductions in carbon mineralization rates (Ogle et al.,  
358 2012; Virto et al., 2012). Such yield penalties under no-till are especially prevalent in colder and wetter climates  
359 (Sun et al., 2020).

360 *Soil tillage directly affects soil macro-fauna by mechanically harming or killing them. In addition to these direct*  
361 *effects of soil disturbance, disruption of the soil also exposes soil macro-fauna to increased risks of desiccation*  
362 *and predation. Consequently, meta-analyses show that total earthworm biomass and abundance increase as tillage*  
363 *intensity is reduced (Spurgeon et al., 2013; Briones and Schmidt 2017; Bai et al. 2018), with a negative relationship*  
364 *between tillage depth and earthworm abundance (Briones and Schmidt 2017). Deep burrowing and surface-feeding*  
365 *(aneecic) earthworm species are particularly favored by NT systems, as their permanent burrows are no longer*  
366 *destroyed by ploughing and they have a better access to food resources. Thus, a lack of disturbance of the soil by*  
367 *tillage has also been shown to increase the diversity of earthworm populations in particular (Chan, 2001; Spurgeon*  
368 *et al., 2013; Briones and Schmidt 2017) and soil fauna in general (de Graaff et al., 2019).*

369 *Changes in tillage systems directly affect the physical properties of soil. For example, bulk density and penetration*  
370 *resistance often increase after the adoption of RT and NT systems (Lee et al. 2019; Li et al., 2019; Li et al., 2020a)*  
371 *due to traffic compaction and the lack of loosening by cultivation (Hamza and Anderson 2005). Peixoto et al.*  
372 *(2020) showed that these negative effects can be alleviated with occasional tillage. The impacts of conservation*



373 tillage practices on soil biological agents and processes also give rise to significant indirect effects on physical  
374 properties, hydrological processes and ecosystem services related to water regulation. Thus, meta-analyses have  
375 shown that saturated hydraulic conductivity and surface infiltration rates often increase under  
376 conservation tillage compared with CT, especially for NT systems (Li et al., 2019; Basche and deLonge, 2019; Li  
377 et al., 2020a; Mondal et al., 2020). This suggests that the effects of the enhanced bioporosity in NT systems created  
378 by soil fauna, and especially anecic earthworms, on saturated and near-saturated hydraulic conductivity (Lee and  
379 Foster, 1991) generally outweigh the negative effects of increased bulk density. Thus, Spurgeon et al. (2013)  
380 showed in their meta-analysis that increased earthworm abundances and diversity found under NT systems were  
381 positively correlated with infiltration rates. Comparing ecological groups, they found that the density of anecic  
382 earthworms was positively associated with increased infiltration rates, whereas no effect was apparent for endogeic  
383 earthworms. Aggregate stability is also largest under NT systems, is intermediate when occasional tillage is  
384 practiced (Peixoto et al. 2020) and smallest in CT systems (Bai et al. 2018). In their meta-analysis, Spurgeon et al.  
385 (2013) showed that improved aggregate stability under NT systems was positively correlated with increases in  
386 fungal biomass. A lack of soil disturbance in NT systems also increases the mean size of aggregates produced in  
387 stability tests (Li et al., 2020a; Mondal et al., 2020). Several meta-analyses have demonstrated increases in field  
388 capacity and available water capacity under reduced and no-till systems (Li et al., 2019; Mondal et al., 2020; Li et  
389 al., 2020a), presumably due to enhanced soil biological activity and increases in organic carbon content. This  
390 would improve water supply to crops under drought, although the effects would appear to be relatively small.

391 In principle, better developed soil macropore systems and improvements in aggregate stability should promote a  
392 more favorable crop water balance, with increases in infiltration and reductions in surface runoff. Figure 5 shows  
393 that the effect on runoff is one of the most studied hydrological processes related to tillage. The meta-analysis  
394 performed by Sun et al. (2015) found that RT and NT systems decreased surface runoff. However, these results  
395 do not appear to be conclusive as two later meta-analyses (Mhazo et al., 2016; Xiong et al., 2018) failed to detect  
396 significant effects of conservation tillage practices on surface runoff. However, Xiong et al. (2018) found that  
397 contour tillage and deep tillage both reduced surface runoff.

#### 398 *Synergies and trade-offs*

399 Adoption of no-till and reduced tillage systems involve several trade-offs, particularly concerning water quality,  
400 GHG emissions and crop yields. As noted earlier, NT systems tend to give smaller yields for many crops compared  
401 with conventional tillage (Mangalassery et al., 2015; Pittelkow et al., 2015; Sun et al., 2020). This may explain  
402 why no-till systems are still seldom adopted in Europe (Mangalassery et al., 2015; Bai et al. 2018), although

403 reduced tillage (RT) is being increasingly adopted worldwide. ~~In their comprehensive meta-analysis,~~ Pitte~~l~~ekow  
404 et al. (2015) identified several reasons for variations in the yield response to no-till. Crop type was the most  
405 important, with no significant yield losses found under NT for oilseed, cotton and legume crops, while the yields  
406 of cereals and root crops were on average ca. 5% and 20% smaller ~~respectively.~~ ~~In accordance with the results of~~  
407 ~~the meta-analyses on stocks of soil organic carbon discussed earlier,~~ Pitte~~l~~ekow et al. (2015) and Sun et al. (2020)  
408 also ~~found~~ show that climate ~~to be~~ is a significant factor, with no significant yield losses for no-till systems under  
409 rain-fed conditions in dry climates. In contrast, Peixoto et al. (2020) showed that occasional tillage increased crop  
410 yields compared with NT in dry regions and in soils with limited water retention capacity and availability,  
411 presumably by alleviating soil compaction and improving rooting.

412 With respect to water quality, Daryanto et al. (2017a) found an overall 40% reduction in phosphorus loads in  
413 surface runoff for NT systems in comparison with CT. This was attributed to significant decreases in losses of  
414 particulate phosphorus, as concentrations of dissolved P actually increased in runoff under NT. For pesticides,  
415 Elias et al. (2018) found no significant differences in concentrations in surface runoff for 14 of the 18 compounds  
416 included in their meta-analysis. Pesticide concentrations were actually larger under NT for the remaining 4  
417 compounds. For loads, no significant difference was detected between CT and NT systems for 15 of the 18  
418 pesticide compounds. For the three remaining pesticides, losses in surface runoff were larger under NT for  
419 metribuzin and dicamba and smaller for alachlor. As also noted by Elias et al. (2018), these results seem quite  
420 surprising given the documented effects of conservation tillage on soil structure and hydraulic properties in the  
421 uppermost soil layers discussed earlier, which should increase soil infiltration capacity and reduce surface runoff.  
422 For nitrate losses in surface runoff in conventional and no-till systems, Daryanto et al. (2017b) showed that a  
423 change to NT resulted in an increase in nitrate concentrations in surface runoff, but similar loads, implying that  
424 surface runoff was, as expected, less prevalent under NT.

425 Daryanto et al. (2017b) also performed a meta-analysis on nitrate leaching. They found larger leachate losses of  
426 nitrate under NT systems than CT, whereas the concentrations in leachate were similar under both tillage systems,  
427 indicating that the effect of NT on nitrate leaching was largely determined by increases in water percolation. We  
428 did not find any meta-analyses on the effects of tillage systems on pesticide leaching in our literature search.  
429 Leaching is the outcome of several interacting processes involving many complex and poorly understood processes  
430 (Alletto et al., 2010). In practice, with no mechanical disturbance, larger quantities of pesticides are often used to  
431 control weeds and diseases in NT systems. However, pesticide leaching will also be highly sensitive to changes  
432 induced by tillage in soil structure, microbial biomass and activity and SOC, since these will affect water flow

433 velocities, degradation rates and the strength of adsorption in soil. Several studies suggest that the better-preserved  
434 macropore networks established under RT and NT systems may enhance leaching by preferential flow (Jarvis,  
435 2007; Larsbo et al., 2009; Alletto et al., 2010). Although it is difficult to draw firm conclusions about the effects  
436 of conservation tillage practices on pesticide leaching without the help of quantitative meta-analyses, we may  
437 tentatively conclude that the greater risk of macropore flow under RT and NT systems appears to outweigh any  
438 beneficial impacts of increases in SOC and microbial activity on pesticide adsorption and degradation.

439 Significant trade-offs have also been reported with respect to greenhouse gases. In an early meta-analysis, van  
440 Kessel et al. (2013) found no overall impact of reduced tillage or no-till on N<sub>2</sub>O emissions, with observed increases  
441 in humid climates compensated by reductions in emissions in drier climates, although neither trend was significant.  
442 However, in a later meta-analysis, Mei et al. (2018) reported a significant overall increase of 18% in N<sub>2</sub>O emissions  
443 under conservation tillage, with the largest effects in warmer and wetter climates and in finer-textured soils. ~~In a~~  
444 ~~recent meta-analysis,~~ Shakoob et al. (2021) found significant increases of emissions of CO<sub>2</sub>, N<sub>2</sub>O and CH<sub>4</sub> of 7, 12  
445 and 21% respectively under NT compared with CT. From the perspective of climate change mitigation, Guenet et  
446 al. (2020) concluded that increased greenhouse gas emissions under NT outweighed any minor gains in soil C  
447 stocks.

#### 448 **4.3.3.3 Amendments**

##### 449 **4.3.3.3.1 Biochar**

450 Biochar is charcoal made for the purpose of soil amendment. It is a type of black carbon, resulting from incomplete  
451 combustion of organic matter through ~~a process known as~~ pyrolysis. Apart from its potential for long-term soil  
452 carbon sequestration, it ~~also can also have has~~ beneficial effects on nutrient availability and soil physical properties  
453 (Joseph et al., 2021).

454 The quantitative analysis of the effects of biochar on physical and hydraulic properties shown in figure 5 is based  
455 on effects ratios presented in five meta-analyses (Omondi et al., 2016; Edeh et al., 2020; Gao et al., 2020; Rabbi  
456 et al., 2021; Ul Islam et al., 2021). Our review ~~also~~ draws on findings presented in two additional reviews that  
457 employed different statistical methodologies (Kroeger et al., 2020; Razzaghi et al., 2020). Rabbi et al. (2021) only  
458 presented data for various sub-categories (e.g. for different types of biochar) and not for the overall effects of  
459 biochar addition. Taken altogether, these seven studies present results of analyses for different soil types, textural  
460 classes and experimental conditions (i.e. field or laboratory/greenhouse study) as well as for biochars of different

461 properties and applied at different rates. These analyses show that biochar has several positive effects on soil  
462 hydraulic properties, but that these effects are dependent on all of the above-mentioned variables.

463 Decreases in bulk density and increases in porosity are generally reported after biochar addition (Omondi et al.,  
464 2016; Edeh, et al., 2020). The density of biochar is low and the porosity is often high compared to soil, which may  
465 explain the observed effects. However, if biochar mainly fills existing pores, porosity will decrease and bulk  
466 density increase. Biochar will also influence these variables indirectly through its effects on aggregation (Pituello  
467 et al., 2018). Two meta-analyses reported measures of aggregate stability. Omondi et al. (2016) included only  
468 studies that reported mean weight diameter (MWD) using wet sieving while UI Islam et al. (2021) included studies  
469 that reported soil aggregate stability as a percentage of water-stable aggregates (WSA), as well as MWD or  
470 gravimetric mean diameter (GMD) using either wet sieving or dry sieving. Both studies showed that aggregate  
471 stability increased with biochar addition and that these effects increased with the time between biochar application  
472 and measurements (UI Islam et al., 2021).

473  
474 Figure 5 suggests that biochar addition generally increases the plant available water content ( $\theta_{paw}$ ). These meta-  
475 analyses show that although the water contents at field capacity ( $\theta_{fc}$ ; pressure potentials in the range between -  
476 0.033 and -0.01 MPa) and wilting point ( $\theta_{pwp}$ ) both tend to increase following biochar amendment, the effects on  
477  $\theta_{fc}$  appear to be larger (figure 5). Pore sizes in biochars range over at least five orders of magnitude, from the sub-  
478 nanometer scale to pore diameters of the order of tens of micrometers originating from partially preserved cellular  
479 structures (Brewer et al., 2014). However, a large fraction of the pore volume in biochar consists of pores in the  
480 nanometer size range (Downie et al., 2009). These pores will retain water at very low pressure potentials and  
481 therefore ~~have the potential to~~ increase the wilting point water content  $\theta_{pwp}$  upon biochar addition. It has been  
482 suggested that increases in  $\theta_{paw}$  may be due to the filling of existing soil macropores with biochar, which would  
483 shift the pore size distribution from large pores that drain quickly to pores that can retain water at field capacity  
484 (Liu et al., 2017). Biochar itself contains pores in the relevant size range (0.2—100  $\mu\text{m}$  in diameter) to contribute  
485 to  $\theta_{paw}$ . Thus, inter-Inter-particle pores in biochar will also contribute to  $\theta_{paw}$  depending on the size distribution and  
486 shapes of the biochar particles and their effects on soil aggregation (Burgeon et al., 2021). Since  $\theta_{fc}$  is the sum of  
487  $\theta_{pwp}$  and  $\theta_{paw}$ , the same processes are the likely causes of the observed increases in  $\theta_{fc}$ .

488 The effects of biochar on water retention were in most cases larger for coarse-textured soils. Biochar with large  
489 microporosity can fill the larger inter-particle soil pores present in sandy soils so that the pore size distribution

490 shifts towards the smaller pores that can retain water at the pressure potentials corresponding to field capacity  
491 (Omondi et al., 2016; Edeh et al., 2020; Rabbi et al., 2021). Moreover, fine-textured soils retain more water at  $\theta_r$   
492 so that the relative changes induced by biochar may be smaller (Edeh et al., 2020).

493 All the meta-analyses included data on the effects of biochar production parameters (e.g. feedstock, pyrolysis  
494 temperature) and the chemical and physical properties of biochar. Generally, the influence of these parameters on  
495 the effects of biochar addition were minor with respect to soil water retention. Due to lack of data, the influence  
496 of the time between biochar application and measurements on the effects on water retention was not included. It  
497 is, however, clear from studies on century-old charcoal kiln sites that the properties of biochar and associated soil  
498 evolve over time (e.g. Cheng et al., 2008; Hardy et al., 2017).

499 One meta-analysis reported an increase in saturated hydraulic conductivity following biochar addition (Omondi et  
500 al., 2016), while two others reported negative effects (figure 5). Saturated hydraulic conductivity is a function of  
501 pore network properties, including connectivity of the macropores and the presence of pore bottlenecks (Koestel  
502 et al., 2018). A few studies have quantified the effects of biochar addition on the connectivity of macropore  
503 networks using X-ray tomography (e.g. Yu and Lu, 2019; Yan et al., 2021). These studies indicate that the  
504 connected macroporosity and the diameter of pore throats decrease in medium- to coarse-textured soils amended  
505 with biochar. However, the influence of soil texture on the effects of biochar on saturated hydraulic conductivity  
506 reported in the meta-analyses is not consistent.

507 ~~Two meta-analyses reported measures of aggregate stability. Omondi et al. (2016) included only studies that~~  
508 ~~reported mean weight diameter (MWD) using wet sieving while Ul-Islam et al. (2021) included studies that~~  
509 ~~reported soil aggregate stability as a percentage of water stable aggregates (WSA), as well as MWD or gravimetric~~  
510 ~~mean diameter (GMD) using either wet sieving or dry sieving. Both studies showed that aggregate stability~~  
511 ~~increased with biochar addition. The reason proposed for this effect was the influence of the added biochar on~~  
512 ~~aggregation processes. The effects on aggregate stability increased with the time between biochar application and~~  
513 ~~measurements (Ul-Islam et al., 2021).~~

514 ~~One meta-analysis focused on water use efficiency (Gao et al., 2020). They showed that both plant water use~~  
515 ~~efficiency defined as the ratio of plant or fruit biomass to water supply, and leaf water use efficiency defined as~~  
516 ~~the ratio of CO<sub>2</sub> uptake by leaves to the loss of water through transpiration, increased with biochar addition. Most~~  
517 ~~biochars are alkaline and may increase soil pH, at least for acidic soils, which leads to improved conditions for~~  
518 ~~plant growth. Biochars also contain nutrients that may become available for plant uptake. Indeed, previous studies~~

519 ~~have shown that the addition of biochar to nutrient poor acidic soils improves yields (Jeffery et al., 2017).~~  
520 ~~Interestingly, in the meta-analysis of Gao et al. (2020), soil pH had different effects on plant and leaf water use~~  
521 ~~efficiency. Leaf water use efficiency increased most for soils with pH less than 7, while plant water use efficiency~~  
522 ~~increased most for soils with pH above 8. The reasons for these results could not be determined from their study.~~  
523 ~~Medrano et al. (2015) showed that leaf water use efficiency may not be well correlated with plant water use~~  
524 ~~efficiency. The large variability in reported effects (-36-313%) can be mainly attributed to differences in soil pH,~~  
525 ~~biochar properties and amounts of added biochar.~~

526 The effects of biochar on the studied variables were in many cases larger for higher application rates. Often,  
527 laboratory studies used much larger application rates (>50 t ha<sup>-1</sup>) than ~~the~~ field studies, ~~probably for economic~~  
528 ~~reasons. This may explain why effects usually were larger in laboratory or greenhouse experiments compared to~~  
529 ~~field trials and also reported larger effects.~~ Additionally, as pointed out by Rabbi et al. (2021), mixing of biochar  
530 after field applications is challenging and may be another reason why effects were sometimes small or insignificant  
531 for field experiments. The majority of the studies included in the meta-analyses were short-term experiments (i.e.  
532 duration < 1 year). Future work should therefore focus on longer-term effects of biochar applications under realistic  
533 field conditions. This requires either long-term field experimentation, which is expensive, or the study of historical  
534 ~~biochar sites. It is also impractical to study these effects for the almost infinite number of combinations of soils,~~  
535 ~~biochars and climates that exist.~~ The meta-analyses ~~included in this study, which~~ show large variations in effects  
536 for all the included variables, which suggests that future work should also be directed towards finding biochars  
537 with specific properties (e.g. surface area, particle size) designed to improve soil physical properties under specific  
538 soil and climate conditions while maintaining or improving nutrient availability.

#### 539 4.3.23.3.2 Other organic amendments, residue retention and mulching

540 Figure 5 shows that only a few meta-analyses have focused specifically on the effects of organic soil amendments  
541 or residue retention and mulching on soil properties relevant for water regulation functions. Instead, these practices  
542 are often included in meta-analyses on conservation agriculture or tillage systems. In these studies, the effects of  
543 the treatments are combined. Furthermore, the influence of contrasting soils or climates has not been assessed.

544 Bai et al. (2018) studied the effects of different organic amendments applied in long-term field experiments on soil  
545 physical and hydraulic properties. They found that aggregate stability increased with organic amendments and that  
546 this effect was largest for compost. However, this beneficial effect decreased with time. Not surprisingly, Bai et  
547 al. (2018) also reported greater aggregate stability under organic farming systems compared with conventional

548 ~~agriculture. Xiong et al. (2018) also included soil amendments applied to agricultural land in their global meta-~~  
549 ~~analysis on soil conservation practices. They found that application of soil amendments reduced both surface~~  
550 ~~runoff and soil erosion. However, they provided no information on the type of amendments and how they were~~  
551 ~~incorporated into soil, which is crucial, since also tillage practices affect these variables (see previous section).~~  
552 Gravuer et al. (2019) analysed effects of organic amendments (manure, biosolids and compost) applied to arid,  
553 semi-arid and Mediterranean rangelands. They found increased water contents at field capacity and reduced  
554 surface runoff. Additional benefits were increased soil organic carbon contents and above-ground net primary  
555 productivity, while trade-offs were increased CO<sub>2</sub> emissions, increased soil lead concentrations and increased  
556 losses of N and P in surface runoff.

557 Mulching means to add (or retain) material on the soil surface without incorporation (Kader et al., 2017). In this  
558 review, we focus on organic mulches, but synthetic materials are also used. The most extreme example of mulching  
559 with artificial materials is plastic mulching, which has been shown to increase crop water efficiency under drought  
560 (Yu et al., 2021). The use of organic amendments may have several beneficial effects on soil quality and the  
561 environment and is therefore one important practice in conservation agriculture. Mulching is typically carried out  
562 to limit soil evaporation, reduce soil runoff and erosion but it also affects, among other things, nutrient cycling,  
563 weed infestations and soil carbon storage (Ranaivoson et al., 2017). Mulching was included as one driver in four  
564 meta-analyses that studied effects on soil hydraulic functions. These meta-analyses showed positive effects on the  
565 rather limited number of hydraulic properties included (figure 5). Three meta-analyses analysed the effects of  
566 mulching on surface runoff, (one for agricultural land (Xiong et al., 2018), one for ~~annualnon-perennial~~ crops  
567 (Ranaivoson et al., 2017) and one focusing only on tree crops (Liu et al., 2021) ~~included effects of mulching on~~  
568 ~~surface runoff~~. They all showed reduced surface runoff. The study for ~~annualnon-perennial~~ crops also showed  
569 reduced soil evaporation and increased infiltration. These effects are already well-established in the scientific  
570 literature and in line with the intentions of mulching (Kader et al., 2017). The meta-analysis by Li et al. (2019)  
571 focused on effects of different tillage practices ~~which also includes the comparison between~~ ~~Here we include the~~  
572 ~~comparison between~~ residue retention and ~~removal no residue retention~~ in no-till systems. ~~Li et al. (2019) showed~~  
573 ~~that R~~ residue retention led to a decrease in bulk density, an increase in total porosity and an increase in plant  
574 available water ~~whereas~~ it did not have significant effects on saturated hydraulic conductivity. They attributed  
575 this to increased accumulation of organic material on the soil surface, which leads to increased biological activity  
576 and soil aggregation.

577 Remaining challenges for biochar and organic amendments as climate adaptation tool

578 Overall, ~~both biochar and~~ organic amendments have potentially beneficial effects for several soil properties  
579 relevant for water regulation. Although this is not a new observation, techniques such as mulching, or biochar  
580 application are still rather little applied. Residue retention is more common in the EU, but worldwide residues are  
581 often still burnt on the field or collected for other uses. The availability of the organic material at the right place at  
582 the right time and for an acceptable price is most probably one of the major bottlenecks for a widespread  
583 application of these amendments, especially in the case of biochar. Future research in this field should therefore  
584 urgently tackle the socio-economic challenges related to the ~~availability~~ application of organic amendments that  
585 ~~prevent to make way for their widespread use~~ application.

#### 586 4.43.4 Irrigation

587 Several recent meta-analyses have investigated the impacts of so-called deficit irrigation on water use efficiency  
588 and/or yields of a range of agricultural crops (Qin et al., 2016; Adu et al., 2018; Lu et al., 2019; Yu et al., 2020;  
589 Cheng et al., 2021a,b). The objective of this approach to irrigation scheduling is to reduce water use without  
590 significantly impacting yields by limiting the supply of water during periods of the growing season when it is less  
591 critical for crop growth. One of these meta-analyses (Cheng et al., 2021a) also synthesized the results of studies  
592 investigating the effects of partial root zone irrigation on water use efficiency and crop yields. This method also  
593 has the objective of saving water without impacting yields, but in this case by alternately supplying water to only  
594 one part of the root zone at each irrigation. These meta-analyses show that although these irrigation scheduling  
595 methods either have mostly neutral or sometimes positive effects on crop water use efficiency (figure 5), crop  
596 yields are significantly smaller compared to full irrigation for almost all crops and soil types. This implies that  
597 crop yields may in some cases be reduced less than water consumption, although these water savings may not  
598 compensate farmers for their yield losses. Another way to conserve high quality fresh water resources is to make  
599 use of brackish or saline water for irrigation. In their meta-analysis, Cheng et al. (2021c) showed how the decreases  
600 in water productivity, irrigation use efficiency and crop yields as a result of the use of salty irrigation water (figure  
601 5) depends on crop type, irrigation methods, climates and soil type.



602 ——— Based on the results of a meta-analysis, Qin et al. (2016) suggested that eliminating over-optimal  
603 (excess) irrigation by more efficient irrigation scheduling would improve the water use efficiency of  
604 citrus by 30% and yields by 20%. Du et al. (2018) showed that so-called aerated irrigation increases  
605 water use efficiency and yields of cereals and vegetables by ca. 20%, presumably by eliminating the  
606 development of anoxic conditions following irrigation. Knowledge gaps

607 3.5 Gaps in the scatter plots shown in figure 5 indicate particular combinations of drivers and target  
608 variables that have not been the subject of meta-analysis according to our search criteria.  
609 Inspection of these figures suggests that there are several significant “knowledge gaps”. Question  
610 marks on the effect of irrigation on the water regulation functions of soil

611 Figure 5 shows that there isn't a single meta-analysis which summarizes the effect of irrigation  
612 on physical soil properties. Nevertheless, the type of irrigation techniques (e.g.e.g., surface, sprinkler or drip  
613 irrigation) must affect the particles at the soil surface differently, due to their profoundly different ways of water  
614 application. Information on this topic can be inferred though from a totally different field of soil research: erosion.  
615 In their review on splash erosion, Fernandez-Raga et al. (2017) report many relevant processes and effects that  
616 might apply to irrigation techniques: increased soil bulk density, crusting, decreased infiltration rate, decreased  
617 germination and plant growth, etc. As long-term experiments on this topic are largely lacking, very little is known  
618 on the long-term impact of distinct irrigation techniques on the water regulation function of the soil. This is even  
619 more the case in regions where irrigation was only punctually used, but where climate change is driving farmers  
620 towards increased use of irrigation.

621 Several recent meta-analyses have investigated the impacts of different irrigation scheduling strategies on crop  
622 performance and crop yield. Many studies report neutral or positive effects of 'so-called deficit irrigation' on water  
623 use efficiency and/or yields of a range of agricultural crops (Qin et al., 2016; Adu et al., 2018; Lu et al., 2019; Yu  
624 et al., 2020; Cheng et al., 2021a,b). The objective of this approach to irrigation scheduling is to reduce water use  
625 without significantly impacting yields by limiting the supply of water during periods of the growing season when  
626 it is less critical for crop growth. Although several meta-analyses have focused on the effects of irrigation  
627 management or organic amendments on water use efficiency, none have been published specifically on the effects  
628 of management practices on water supply to crops. Such information should be critical to support policies and  
629 practices for effective adaptation of farming systems to future climates with more frequent and severe summer  
630 droughts. We can therefore only make inferences about the effects of soil management on crop transpiration from  
631 other terms in the soil water balance. One of these meta-analyses (Cheng et al., 2021a) also synthesized the results

632 ~~of studies investigating the effects of partial root zone irrigation on water use efficiency and crop yields. This~~  
633 ~~method also has the objective of saving water without impacting yields, but in this case by alternately suppling~~  
634 ~~water to only one part of the root zone at each irrigation.~~

635 ~~These meta analyses show that although these irrigation scheduling methods either have mostly neutral or~~  
636 ~~sometimes positive effects on crop water use efficiency (figure 5), crop yields are significantly smaller compared~~  
637 ~~to full irrigation for almost all crops and soil types. This implies that crop yields may in some cases be reduced~~  
638 ~~less than water consumption, although these water savings may not compensate farmers for their yield losses.~~

639 ~~Apart from more efficient scheduling and dosing, another way to conserve high quality fresh water resources is to~~  
640 ~~make use of brackish or saline water for irrigation. In their meta-analysis, Cheng et al. (2021c) showed how the~~  
641 ~~decreases in water productivity, irrigation use efficiency and crop yields as a result of the use of salty irrigation~~  
642 ~~water (figure 5) depends on crop type, irrigation methods, climates and soil type. Based on the results of a meta-~~  
643 ~~analysis, Qin et al. (2016) suggested that eliminating over-optimal (excess) irrigation by more efficient irrigation~~  
644 ~~scheduling would improve the water use efficiency of citrus by 30% and yields by 20%. Du et al. (2018) showed~~  
645 ~~that so called aerated irrigation increases water use efficiency and yields of cereals and vegetables by ca. 20%,~~  
646 ~~presumably by eliminating the development of anoxic conditions following irrigation. This is a very relevant~~  
647 ~~topic, since also reclaimed water is increasingly looked at as alternative water source, but no meta-analysis is yet~~  
648 ~~available on how this specific type of water affects soil physical properties. In addition to salinity, organic~~  
649 ~~compounds in those types of water may result in increased hydrophobicity and clogging of soil pores.~~

650

651 ~~A plea for investing in ‘difficult’ variables in long-term experiments: interactions, soil physical~~  
652 ~~properties and rooting characteristics~~Gaps in the scatter plots shown in figure 5 indicate particular  
653 combinations of drivers and target variables that have not been the subject of meta-analysis according to  
654 our search criteria. Inspection of these figures suggests that there are several significant “knowledge  
655 gaps”. Although several meta-analyses have focused on the effects of irrigation management or organic  
656 amendments on water use efficiency, none have been published specifically on the effects of  
657 management practices on water supply to crops. Such information should be critical to support policies  
658 and practices for effective adaptation of farming systems to future climates with more frequent and  
659 severe summer droughts. We can therefore only make inferences about the effects of soil management  
660 on crop transpiration from other terms in the soil-water balance.

661 ~~Presumably for reasons of cost, many long-term field experiments often only have simple designs, neglecting~~  
662 ~~potentially interesting combinations of treatments, for example, no-till combined with the use of cover crops.~~  
663 ~~Similarly, the interactions between soil and crop management and irrigation or drainage systems and practices do~~  
664 ~~not appear to be a common topic of field experimentation apart from the combination between tillage practices~~  
665 ~~and residue management~~only one meta-analysis has studied the effects of management on crop root system  
666 characteristics. This is most probably because root system characterization is tedious, time-consuming and too  
667 ~~invasive for long-term field trials. Nevertheless, the soil-root interface is a crucial environment mediating the flow~~  
668 ~~of water in the soil-plant-atmosphere system. It can be expected that many soil management practices influence~~  
669 ~~root penetration and rooting depths, thereby strongly influencing potential rates of water uptake by plants during~~  
670 ~~droughts. Such information should be critical to support policies and practices for effective adaptation of farming~~  
671 ~~systems to future climates with more frequent and severe summer droughts.~~At present, wWe can therefore only  
672 make inferences about the effects of soil management on crop transpiration, either from other terms in the soil  
673 water balance or from the use of penetration resistance as a proxy (figure 5).

674 ~~Figure 5 shows that although several recent meta-analyses have investigated the impacts of different irrigation~~  
675 ~~scheduling strategies on water use efficiency and crop yields, none has so far summarized the effects of irrigation~~  
676 ~~on soil physical properties. Nevertheless, the type of irrigation technique (e.g., surface, sprinkler or drip irrigation)~~  
677 ~~and the quality of water used are known to strongly affect soil structure and hydraulic properties (Sun et al., 2018;~~  
678 ~~Leuther et al., 2019; Drewry et al., 2021), which should impact the water regulation functions of soil. A quantitative~~  
679 ~~summary of existing experimental information would provide critical support to policies and practices for effective~~  
680 ~~adaptation of farming systems to future climates with more frequent and severe summer droughts.~~

681 ~~Some additional potential knowledge gaps concerning the effects of soil management on water regulation functions~~  
682 ~~are not revealed by inspection of figure 5. Firstly, with some exceptions (i.e. tillage practices and residue~~  
683 ~~management), most long-term field experiments only have simple designs that neglect some potentially interesting~~  
684 ~~combinations of treatments (e.g. the interactions between soil and crop management and irrigation systems). Other~~  
685 ~~knowledge gaps may be only apparent and therefore less serious: macroporosity is rarely studied in the context of~~  
686 ~~meta-analysis, although infiltration has been much more frequently measured and these two variables should be~~  
687 ~~strongly correlated. Figure 5 also indicates that some management practices have been less often the subject of~~  
688 ~~field experiments including, for example, deep tillage, occasional tillage, and crop rotations. Presumably for~~  
689 ~~reasons of cost, many long-term field experiments often only have simple designs, neglecting potentially~~  
690 ~~interesting combinations of treatments, for example, no-till combined with the use of cover crops. Similarly, the~~  
691 ~~interactions between soil and crop management and irrigation or drainage systems and practices do not appear to~~  
692 ~~be a common topic of field experimentation.~~

693 ~~Some additional potential knowledge gaps concerning the effects of soil management on water regulation functions~~  
694 ~~are not revealed by inspection of figure 5. Secondly, some key targets since they concern variables that are rarely~~  
695 ~~measured and so have not yet been the subject of meta-analysis. For example, m~~  
696 ~~Most long-term field trials on the effects of soil and crop management practices on hydrological and biological functioning have measured proxy~~  
697 ~~variables surrogate variables (or proxies) for soil structure, such as infiltration rates or soil hydraulic properties~~  
698 ~~(water retention, hydraulic conductivity at and near saturation). No meta-analyses have been performed yet for~~  
699 ~~metrics quantifying various different aspects of soil structure per se (Rabot et al., 2018) even though the application~~  
700 ~~of X-ray imaging techniques to quantify soil structure is now becoming increasingly common. As a result, the~~  
701 ~~number of X-ray studies published is rapidly increasing, so it should soon not be too long before it will be possible~~  
702 ~~and worthwhile to carry out such an analysis.~~

703 ~~Similarly, very little meta-analyses contain information on the root system characteristics of the crops involved.~~  
704 ~~This is most probably because root system characterization is tedious, time-consuming and often invasive in field~~  
705 ~~trials. Nevertheless, the soil-root interface is a crucial environment mediating the flow of water in the soil-plant-~~  
706 ~~atmosphere system. It can be expected that several of the studied practices influence rooting depths, thereby~~  
707 ~~strongly influencing the water uptake capacity of plants and ultimately the water balance.~~

708

#### 709 54 Implications for climate change adaptation Conclusions and outlook

710 A large number of meta-analyses have been published in recent years on the impacts of soil and crop management  
711 practices on soil properties and processes and the various ecosystem services and functions delivered by soil. In  
712 this report, we have synthesized these analyses with respect to the water regulation functions that are relevant for  
713 climate change adaptation in Europe. ~~Across Europe, climatic extremes (i.e. droughts and intense rains) will  
714 become more frequent and more severe. Specifically, effective adaptation to climate change requires soils with a  
715 well-developed and stable structure with a large infiltration capacity and an ability to sustain water supply to plants  
716 during extended dry periods.~~ This synthesis has revealed a considerable degree of consensus concerning the effects  
717 of soil and crop management practices ~~on key soil properties relevant for these hydrological functions, despite the  
718 fact that meta-analyses cannot easily account for differences in experimental conditions among individual source  
719 studies, not least because many primary studies do not report all details of the experimental treatments. This  
720 overview has also identified several important knowledge gaps, particularly related to the effects of management  
721 practices on root growth and transpiration. Thus, conclusions related to the impacts of management on the crop  
722 water supply are necessarily based on inferences derived from proxy variables such as available water capacity  
723 and infiltration capacity.~~

724 Meta-analyses have demonstrated that the use of organic amendments and the adoption of cropping systems and  
725 practices that maintain, as far as possible, “continuous living cover” both result in significant beneficial effects for  
726 the water regulation function of soils, arising from the additional carbon inputs to soil and the stimulation of  
727 biological processes. These effects are clearly related to improvements in soil structure, both in terms of stable  
728 aggregation at the micro-scale and enhanced bio-porosity, both of which reduce surface runoff and increase  
729 infiltration. ~~Meta-analyses show that amendment of soils with biochar generally increases aggregate stability,  
730 reduces bulk density, increases porosity and improves the plant available water capacity, particularly for coarse-  
731 textured soils.~~ One potentially negative consequence of management practices that maintain “continuous living  
732 cover” is a reduction in soil water storage and groundwater recharge ~~that, in most cases, will likely outweigh any  
733 increases in soil water storage capacity due to carbon sequestration.~~ This may be problematic in dry climates,  
734 where there is some evidence to suggest that yields of the main crop may be affected. With respect to  
735 environmental quality, no other significant trade-offs are known, while some important synergies have been  
736 identified, in particular reductions in nitrate leaching to groundwater and greenhouse gas emissions.

737 ~~There is little evidence from meta-analyses to support the idea that reductions in tillage intensity improve crop  
738 water supply. The effects of no till on SOM stocks and thus the capacity of the soil to store plant available water  
739 appear to be minimal. In contrast, the amelioration of soil structure that occurs under reduced (RT) and no-till~~

740 (NT) practices may improve infiltration capacity and reduce surface runoff, despite the increases in bulk density  
741 that are commonly reported, although the evidence for this is inconclusive. Furthermore, sSome significant trade-  
742 offs with RT and NT systems have also been identified. For example, yield penalties incurred under NT and  
743 increased weed pressure and/or increased herbicide use and thus leaching risks, especially in wetter and colder  
744 climates, constitute a barrier to adoption by farmers. GFurthermore, greenhouse gas emissions are also generally  
745 larger under NT, while leaching losses to groundwater of both nitrate and pesticides may also increase. Although  
746 we might expect losses of agro-chemicals in surface runoff to generally decrease under RT and NT, thereby  
747 compensating for greater leaching losses, this does not always appear to be the case. Reduced tillage intensity in  
748 the temporal sense (i.e. “occasional” tillage) may help to ameliorate some of the negative effects of no-till systems,  
749 whilst retaining some of the advantages.

750 ~~Our~~This extensive synthesis of ~~the~~ existing literature ~~has on the effect of a diverse panel of management practices~~  
751 ~~on the water regulation function of soil in the light of climate change not only showed us areas of consensus. We~~  
752 ~~have~~ also identified several important knowledge gaps, particularly related to the effects of management practices  
753 ~~on soil structure, root growth and transpiration and on combinations of practices. Thus, conclusions related to the~~  
754 ~~impacts of management on the crop water supply are necessarily based on inferences derived from proxy variables~~  
755 ~~such as available water capacity and infiltration capacity.- To address these limitations,~~

756 To address these limitations, we recommend that future research should focus on the following:

757 1) monitoring transpiration (e.g. by sap flow) and crop root development in existing field trials and the  
758 development of techniques to do this in a minimally invasive way for the entire soil root zone;  
759 monitoring of soil structure and hydraulic properties in field trials over the entire soil profile;  
760 application of soil-crop models making use of measured hydraulic properties and climate model  
761 projections to evaluate and predict the impacts of alternative soil/crop management practices on water  
762 balance and crop yields under climate change;  
763 introduction of irrigation and drought treatments at existing long-term field trials to investigate the  
764 consequences for water regulation functions under climate change.

765 1)  
766  
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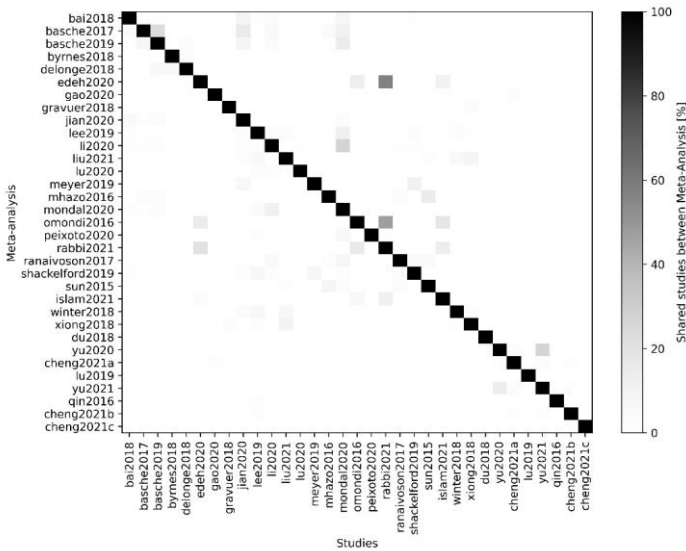
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768 **Appendices**

769 **Appendix 1 Redundancy analysis**

770 Note that for this analysis, the studies of Li et al. (2019) and Li et al. (2020) were considered as one, as they both  
771 rely on the same database, but analyze different variables. We first identified the studies shared between multiple  
772 meta-analyses and computed the percentage of shared studies per meta-analysis. Figure A1 shows the percentage  
773 of shared studies (number of shared studies divided by number of studies in the meta-analysis in the row times  
774 100). A2 shows for each meta-analysis the number of source studies that it shares with at least one other meta-  
775 analysis. Some meta-analyses share nearly 100% of their studies with another meta-analysis (e.g. Omondi et al.,  
776 2016, Edeh et al., 2020). In addition to the extent of redundancy, A2 also shows the number of primary studies  
777 included in each meta-analysis. For example, Jian et al. (2020), Li et al. (2020) and Mondal et al. (2020) considered  
778 more than 200 primary studies in their meta-analyses. Finally, A3 shows for each meta-analysis the percentage of  
779 its primary studies that are shared with another meta-analysis. For example, the studies by Omondi et al. (2016)  
780 and Rabbi et al. (2021) share a large proportion of primary studies. A3 also shows that nearly all the primary  
781 studies included in these two meta-analyses are shared with another meta-analysis.

782  
783 **Redundancy matrix showing the percentage of shared studies among meta-analyses. The percentage refers to the**  
784 **number of shared studies divided by the total number of studies in the meta-analysis in the row. Note that this**  
785 **matrix is not symmetrical, because the percentage is computed for the meta-analysis in the row. If we had shown**  
786 **the number of shared studies as a number and not a percentage, this matrix would have been symmetrical.**



787 **Figure A1: Redundancy matrix showing the percentage of shared studies among meta-analyses. The**  
788 **percentage refers to the number of shared studies divided by the total number of studies in the meta-analysis**  
789 **in the row. Note that this matrix is not symmetrical, because the percentage is computed for the meta-**  
790 **analysis in the row.**

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791 analysis in the row. If we had shown the number of shared studies as a number and not a percentage, this  
792 matrix would have been symmetrical.

793 ~~Redundancy matrix showing the percentage of shared studies among meta-analyses.~~

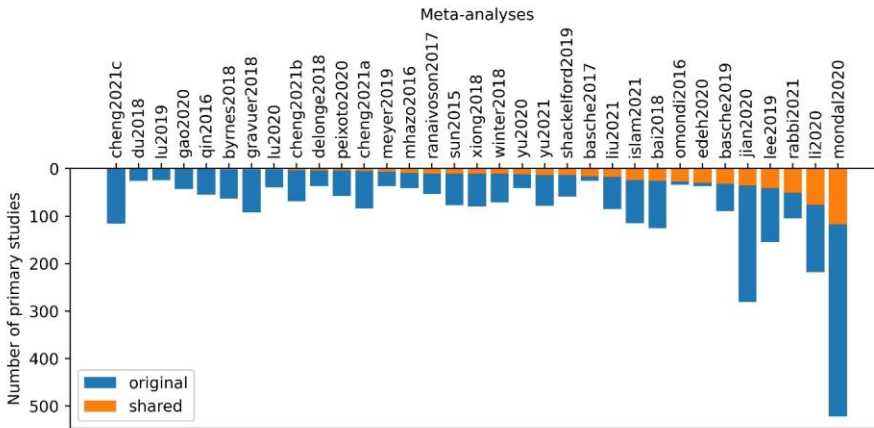
794

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795 Histogram showing the number of studies per meta-analysis. The studies shared by at least one other meta-  
 796 analysis are displayed in light green (shared), while the studies found only in this meta-analysis are shown in  
 797 dark green (original).

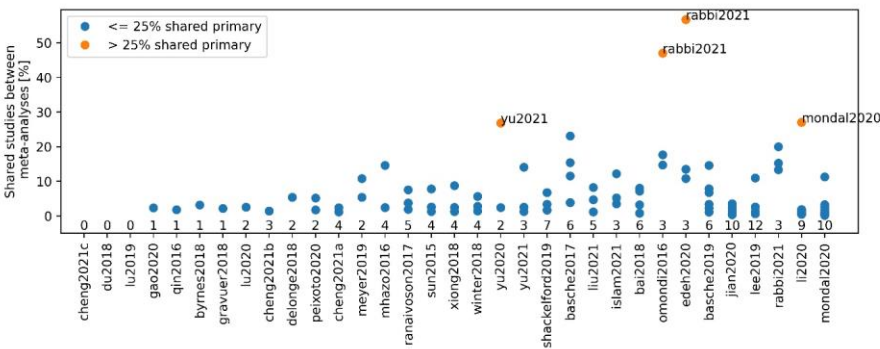
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798  
 799 **Figure A2:** Histogram showing the number of primary studies per meta-analysis. The studies shared by at  
 800 least one other meta-analysis are displayed in light green (shared), while the studies found only in this meta-  
 801 analysis are shown in dark green (original).  
 802

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803  
 804 Redundancy among the selected meta-analyses (horizontal axis). Dots represent the percentage of shared primary  
 805 studies between two meta-analyses. When this percentage is above 25%, the dots are shown in red, and the name  
 806 of the meta-analysis is displayed. For instance, Li et al. (2020) shares more than 25% of its primary studies with  
 807 the meta-analysis of Mondal et al. (2020). The number on the horizontal-axis denotes the number of other meta-  
 808 analyses that share primary studies with the meta-analysis named horizontally. Note that several meta-analysis do  
 809 not share any studies with others. Meta-analysis are sorted according to the amount of shared primary studies they  
 810 have (same order as A2).



811

812 **Figure A3: Redundancy among the selected meta-analyses (horizontal axis). Dots represent the percentage**  
813 **of shared primary studies between two meta-analyses. When this percentage is above 25%, the dots are**  
814 **shown in red, and the name of the meta-analysis is displayed. For instance, Li et al. (2020) shares more than**  
815 **25% of its primary studies with the meta-analysis of Mondal et al (2020). The number on the horizontal**  
816 **axis denotes the number of other meta-analyses that share primary studies with the meta-analysis named**  
817 **horizontally. Note that several meta-analysis do not share any studies with others. Meta-analysis are sorted**  
818 **according to the amount of shared primary studies they have (same order as A2).**  
819

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▲ **Redundancy among the selected meta-analyses (horizontal axis). Dots represent the percentage of shared primary studies between two meta-analyses.**

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**Met opmaak:** Standaard, Uitvullen, Regelaafstand: 1,5 regel

- 824 **Appendix 2 Meta-analyses on soil/crop management and water regulation in the EU**
- 825 Bai, Z., Caspari, T., Gonzalez, M., Batjes, N., Mäder, P., Bünemann, E., de Goede, R., Brussaard, L., Xu, M.,  
826 Ferreira, C., Reintam, E., Fan, H., Mihelič, R., Glavan, M., Tóth, Z. 2018. Effects of agricultural management  
827 practices on soil quality: a review of long-term experiments for Europe and China. *Agriculture, Ecosystems*  
828 *and Environment*, 265, 1-7.
- 829 Basche, A., deLonge, M. 2017. The impact of continuous living cover on soil hydrologic properties: a meta-  
830 analysis. *Soil Science Society of America Journal*, 81, 1179-1190.
- 831 Basche, A., deLonge, M. 2019. Comparing infiltration rates in soils managed with conventional and alternative  
832 farming methods: a meta-analysis. *PLOS One*, 14, e0215702.
- 833 Byrnes, R., Eastburn, D., Tate, K., Roche, L. 2018. A global meta-analysis of grazing impacts on soil health  
834 indicators. *Journal of Environmental Quality*, 47, 758-765.
- 835 Cheng, M., Wang, H., Fan, J., Zhang, S., Liao, Z., Zhang, F., Wang, Y. 2021a. A global meta-analysis of yield  
836 and water use efficiency of crops, vegetables and fruits under full, deficit and alternate partial root-zone  
837 irrigation. *Agricultural Water Management*, 248, 106771.
- 838 Cheng, M., Wang, H., Fan, J., Zhang, S., Wang, Y., Li, Y., Sun, X., Yang, L., Zhang, F. 2021b. Water productivity  
839 and seed cotton yield in response to deficit irrigation: A global meta-analysis. *Agricultural Water*  
840 *Management*, 255, 107027.
- 841 Cheng, M., Wang, H., Fan, J., Wang, X., Sun, X., Yang, L., Zhang, S., Xiang, Y., Zhang, F. 2021c. Crop yield  
842 and water productivity under salty water irrigation: A global meta-analysis. *Agricultural Water Management*,  
843 256, 107105.
- 844 DeLonge, M., Basche, A. 2018. Managing grazing lands to improve soils and promote climate change adaptation  
845 and mitigation: a global synthesis. *Renewable Agriculture and Food Systems*, 33, 267-278.
- 846 Du, Y-D., Niu, W-Q., Gu, X-B., Zhang, Q., Cui, B-J., Zhao, Y. 2018. Crop yield and water use efficiency under  
847 aerated irrigation: a meta-analysis. *Agricultural Water Management*, 210, 158-164.
- 848 Edeh, I., Masek, O., Buss, W. 2020. A meta-analysis on biochar's effects on soil water properties - new insights  
849 and future research challenges. *Science of the Total Environment*, 714, 136857.
- 850 Gao, Y., Shao, G., Lu, J., Zhang, K., Wu, S., Wang, Z. 2020. Effects of biochar application on crop water use  
851 efficiency depend on experimental conditions: a meta-analysis. *Field Crops Research*, 249, 107763.
- 852 Gravuer, K., Gennet, S., Throop, H. 2018. Organic amendment additions to rangelands: a meta-analysis of multiple  
853 ecosystem outcomes. *Global Change Biology*, 25, 1152-1170.
- 854 Jian, J., Lester, B., Du, X., Reiter, M., Stewart, R. 2020. A calculator to quantify cover crop effects on soil health  
855 and productivity. *Soil and Tillage Research*, 199, 104575.
- 856 Lee, H., Lautenbach, S., Nieto, A., Bondeau, A., Cramer, W., Geijzendorffer, I. 2019. The impact of conservation  
857 farming practices on Mediterranean agro-ecosystem services provisioning - a meta-analysis. *Regional*  
858 *Environmental Change*, 19, 2187-2202.
- 859 Li, Y., Li, Z., Cui, S., Zhang, Q. 2020. Trade-off between soil pH, bulk density and other soil physical properties  
860 under global no-tillage agriculture. *Geoderma*, 361, 114099.
- 861 Li, Y., Li, Z., Cui, S., Jagadamma, S., Zhang, Q. 2019. Residue retention and minimum tillage improve physical  
862 environment of the soil in croplands: a global meta-analysis. *Soil and Tillage Research*, 194, 104292.

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863 Liu, R., Thomas, B., Shi, X., Zhang, X., Wang, Z., Zhang, Y. 2021. Effects of ground cover management on  
864 improving water and soil conservation in tree crop systems: a meta-analysis. *Catena*, 199, 105085.

865 Lu, J., Shao, G., Cui, J., Wang, X., Keabetswe, L. 2019. Yield, fruit quality and water use efficiency of tomato for  
866 processing under regulated deficit irrigation: A meta-analysis. *Agricultural Water Management*, 222, 301-312.

867 Lu, X. 2020. A meta-analysis of the effects of crop residue return on crop yields and water use efficiency. *PLOS*  
868 *One*, 15, e0231740.

869 Meyer, N., Bergez, J., Constantin, J., Justes, E. 2019. Cover crops reduce water drainage in temperate climates: a  
870 meta-analysis. *Agronomy for Sustainable Development*, 39, 3.

871 Mhazo, N., Chivenge, P., Chaplot, V. 2016. Tillage impact on soil erosion by water: discrepancies due to climate  
872 and soil characteristics. *Agriculture, Ecosystems and Environment*, 230, 231-241.

873 Mondal, S; Chakraborty, D; Bandyopadhyay, K; Aggarwal, P; Rana, DS. 2020. A global analysis of the impact of  
874 zero-tillage on soil physical condition, organic carbon content, and plant root response. *Land Degradation and*  
875 *Development*, 31, 557-567.

876 Omondi, M., Xia, X., Nahayo, A., Liu, X., Korai, P., Pan, G. 2016. Quantification of biochar effects on soil  
877 hydrological properties using meta-analysis of literature data. *Geoderma*, 274, 28-34.

878 Peixoto, D., da Silva, L., de Melo, L., Azevedo, R., Araujo, B., de Carvalho, T., Moreira, S., Curi, N., Silva, B.  
879 2020. Occasional tillage in no-tillage systems: a global meta-analysis. *Science of the Total Environment*, 745,  
880 140887.

881 Qin, W., Assinck, F., Heinen, M., Oenema, O. 2016. Water and nitrogen use efficiencies in citrus production: a  
882 meta-analysis. *Agriculture, Ecosystems and Environment*, 222, 103-111.

883 Rabbi, S., Minasny, B., Salami, S., McBratney, A., Young, I. 2021. Greater, but not necessarily better: the  
884 influence of biochar on soil hydraulic properties. *European Journal of Soil Science*, DOI: 10.1111/ejss.13105.

885 Ranaivoson, L., Naudin, K., Ripoché, A., Affholder, F., Rabeharisoa, L., Corbeels, M. 2017. Agro-ecological  
886 functions of crop residues under conservation agriculture. A review. *Agronomy for Sustainable Development*,  
887 37, 26.

888 Shackelford, G., Kelsey, R., Dicks, L. 2019. Effects of cover crops on multiple ecosystem services: ten meta-  
889 analyses of data from arable farmland in California and the Mediterranean. *Land Use Policy*, 88, 104204.

890 Sun, Y., Zeng, Y., Shi, Q., Pan, X., Huang, S. 2015. No-tillage controls on runoff: a meta-analysis. *Soil and Tillage*  
891 *Research*, 153, 1-6.

892 Ul Islam, M., Jiang, F., Guo, Z., Peng, X. 2021. Does biochar application improve soil aggregation? A meta-  
893 analysis. *Soil and Tillage Research*, 209, 104926.

894 Winter, S., Bauer, T., Strauss, P., Kratschmer, S., Paredes, D., Popescu, D., Landa, B., Guzman, G., Gomez, J.,  
895 Guernion, M., Zaller, J., Batary, P. 2018. Effects of vegetation management intensity on biodiversity and  
896 ecosystem services in vineyards: a meta-analysis. *Journal of Applied Ecology*, 55, 2484-2495.

897 Xiong, M., Sun, R., Chen, L. 2018. Effects of soil conservation techniques on water erosion control: a global  
898 analysis. *Science of the Total Environment*, 645, 753-760.

899 Yu, L., Zhao, X., Gao, X., Siddique, K. 2020. Improving/maintaining water-use efficiency and yield of wheat by  
900 deficit irrigation: a global meta-analysis. *Agricultural Water Management*, 228, 105906.

901 Yu, L., Zhao, X., Gao, X., Jia, R., Yang, M., Yang, X., Wu, Y., Siddique, K. 2021. Effect of natural factors and  
902 management practices on agricultural water use efficiency under drought: a meta-analysis of global drylands.  
903 Journal of Hydrology, 594, 125977.

904

#### 905 **Code availability**

906 Code is available at <https://github.com/climasoma/review-of-meta-analyses>

Gewijzigde veldcode

#### 907 **Data availability**

908 Data is available at <https://github.com/climasoma/review-of-meta-analyses>

#### 909 **Author contribution**

910 Conceptualization: all co-authors

911 Data collation and analysis: GuB, GiB, NJ

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913 Writing, review and editing: all co-authors

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## References

- Adu, M., Yawson, D., Armah, F., Asare, P., Frimpong, K. 2018. Meta-analysis of crop yields of full, deficit, and partial root-zone drying irrigation. *Agricultural Water Management*, 197, 79-90. <https://doi.org/10.1016/j.agwat.2017.11.019>
- Aguilera, E., Lassaletta, L., Gattinger, A., Gimeno, B. 2013. Managing soil carbon for climate change mitigation and adaptation in Mediterranean cropping systems: a meta-analysis. *Agriculture, Ecosystems and Environment*, 168, 25-36. <https://doi.org/10.1016/j.agee.2013.02.003>.
- Alletto, L., Coquet, Y., Benoit, P., Heddadj, D., Barriuso, E. 2010. Tillage management effects on pesticide fate in soils. A review. *Agronomy for Sustainable Development*, 30, 367-400. [https://doi.org/10.1007/978-94-007-0394-0\\_35](https://doi.org/10.1007/978-94-007-0394-0_35).
- Angers, D., Eriksen-Hamel, N. 2008. Full-inversion tillage and organic carbon distribution in soil profiles: a meta-analysis. *Soil Science Society of America Journal*, 72, 1370-1374. <https://doi.org/10.2136/sssaj2007.0342>.
- Angus, J., Kirkegaard, J., Hunt, J., Ryan, M., Ohlander, L., Peoples, M. 2015. Break crops and rotations for wheat. *Crop and Pasture Science*, 66, 523-552. <https://doi.org/10.1071/CP14252>.
- Bai, Z., Caspari, T., Gonzalez, M., Batjes, N., Mäder, P., Bünemann, E., de Goede, R., Brussaard, L., Xu, M., Ferreira, C., Reintam, E., Fan, H., Mihelič, R., Glavan, M., Tóth, Z. 2018. Effects of agricultural management practices on soil quality: a review of long-term experiments for Europe and China. *Agriculture, Ecosystems and Environment*, 265, 1-7. <https://doi.org/10.1016/j.agee.2018.05.028>.
- Bai, X., Huang, Y., Ren, W., Coyne, M., Jacinthe, P-A., Tao, B., Hui, D., Yang, J., Matocha, C. 2019. Responses of soil carbon sequestration to climate-smart agriculture practices: a meta-analysis. *Global Change Biology*, 25, 2591-2606. <https://doi.org/10.1111/gcb.14658>.
- Basche, A., Miguez, F., Kaspar, T., Castellano, M. 2014. Do cover crops increase or decrease nitrous oxide emissions? A meta-analysis. *Journal of Soil and Water Conservation*, 69, 471-482. <https://doi.org/10.2489/jswc.69.6.471>.
- Basche, A., deLonge, M. 2017. The impact of continuous living cover on soil hydrologic properties: a meta-analysis. *Soil Science Society of America Journal*, 81, 1179-1190. <https://doi.org/10.2136/sssaj2017.03.0077>.
- Basche, A., deLonge, M. 2019. Comparing infiltration rates in soils managed with conventional and alternative farming methods: a meta-analysis. *PLOS One*, 14, e0215702. <https://doi.org/10.1371/journal.pone.0215702>.
- Beillouin, D., Ben-Ari, T., Makowski, D. 2019a. A dataset of meta-analyses on crop diversification at the global scale. *Data in Brief*, 24, 103898. <https://doi.org/10.1016/j.dib.2019.103898>.
- Beillouin, D., Ben-Ari, T., Makowski, D. 2019b. Evidence map of crop diversification strategies at the global scale. *Environmental Research Letters*, 14, 123001. <https://doi.org/10.1088/1748-9326/ab4449>.
- Bertrand, M., Barot, S., Blouin, M., Whalen, J., de Oliveira, T., Roger-Estrade, J. 2015. Earthworm services for cropping systems. A review. *Agronomy and Sustainable Development*, 35, 553-567. <https://doi.org/10.1007/s13593-014-0269-7>
- Blanco-Canqui, H., Ruis, S., Holman, J., Creech, C., Obour, A. 2022. Can cover crops improve soil ecosystem services in water-limited environments? A review. *Soil Science Society of America Journal*, 86, 1-18. <https://doi.org/10.1002/saj2.20335>

Gewijzigde veldcode

Gewijzigde veldcode

Gewijzigde veldcode

Bolinder, M., Crotty, F., Elsen, A., Frac, M., Kismányoky, T., Lipiec, J., Tits, M., Tóth, Z., Kätterer, T. 2020. The effect of crop residues, cover crops, manures and nitrogen fertilization on soil organic carbon changes in agroecosystems: a synthesis of reviews. *Mitigation and Adaptation Strategies for Global Change*, 25, 929-952. <https://doi.org/10.1007/s11027-020-09916-3>.

Brewer, C., Chuang, V., Masiello, C., Gonnermann, H., Gao, X., Dugan, B., Driver, L., Panzacchi, P., Zygourakis, K., Davies, C. 2014. New approaches to measuring biochar density and porosity. *Biomass and Bioenergy* 66, 176-185. <https://doi.org/10.1016/j.biombioe.2014.03.059>

Briones, M.J.I., Schmidt, O., 2017. Conventional tillage decreases the abundance and biomass of earthworms and alters their community structure in a global meta-analysis. *Global Change Biology*, 23, 4396-4419. <https://doi.org/10.1111/gcb.13744>

Burgeon, V., Fouché, J., Leifeld, J., Chenu, C., Cornélis, J-T. 2021. Organo-mineral associations largely contribute to the stabilization of century-old pyrogenic organic matter in cropland soils. *Geoderma*, 388, 114841. <https://doi.org/10.1016/j.geoderma.2020.114841>

Byrnes, R., Eastburn, D., Tate, K., Roche, L. 2018. A global meta-analysis of grazing impacts on soil health indicators. *Journal of Environmental Quality*, 47, 758-765.

<https://doi.org/10.2134/jeq2017.08.0313>.

Capowiez, Y., T. Lévêque, T., Pelosi, C., Capowiez, L., Mazzia, C., Schreck, E., Dumat, C. 2021. Using the ecosystem engineer concept to test the functional effects of a decrease in earthworm abundance due to an historic metal pollution gradient. *Applied Soil Ecology*, 158, 103816.

<https://doi.org/10.1016/j.apsoil.2020.103816>.

Chan, K.Y. 2001. An overview of some tillage impacts on earthworm population abundance and diversity – implications for functioning in soils. *Soil and Tillage Research*, 57, 179-191.

[https://doi.org/10.1016/S0167-1987\(00\)00173-2](https://doi.org/10.1016/S0167-1987(00)00173-2).

Chen, H., Dai, Z., Veach, A.M., Zheng, J., Xu, J., Schadt, C.W., 2020. Global meta-analyses show that conservation tillage practices promote soil fungal and bacterial biomass. *Agriculture, Ecosystems and Environment*, 293, 106841. <https://doi.org/10.1016/j.agee.2020.106841>.

Cheng, M., Wang, H., Fan, J., Zhang, S., Liao, Z., Zhang, F., Wang, Y. 2021a. A global meta-analysis of yield and water use efficiency of crops, vegetables and fruits under full, deficit and alternate partial root-zone irrigation. *Agricultural Water Management*, 248, 106771.

<https://doi.org/10.1016/j.agwat.2021.106771>.

Cheng, M., Wang, H., Fan, J., Zhang, S., Wang, Y., Li, Y., Sun, X., Yang, L., Zhang, F. 2021b. Water productivity and seed cotton yield in response to deficit irrigation: A global meta-analysis. *Agricultural Water Management*, 255, 107027. <https://doi.org/10.1016/j.agwat.2021.107027>.

Cheng, M., Wang, H., Fan, J., Wang, X., Sun, X., Yang, L., Zhang, S., Xiang, Y., Zhang, F. 2021c. Crop yield and water productivity under salty water irrigation: A global meta-analysis. *Agricultural Water Management*, 256, 107105. <https://doi.org/10.1016/j.agwat.2021.107105>.

Chenu, C., Le Bissonnais, Y., Arrouays, D., 2000. Organic matter influence on clay wettability and soil aggregate stability. *Soil Science Society America Journal*, 64, 1479-1486.

<https://doi.org/10.2136/sssaj2000.6441479x>.

Daryanto, S., Wang, L., Jacinthe, P-A. 2017a. Meta-analysis of phosphorus loss from no-till soils. *Journal of Environmental Quality*, 46, 1028-1037. <https://doi.org/10.2134/jeq2017.03.0121>.

Daryanto, S., Wang, L., Jacinthe, P-A. 2017b. Impacts of no-tillage management on nitrate loss from corn, soybean and wheat cultivation: a meta-analysis. *Scientific Reports* 7: 12117 <https://doi.org/10.1038/s41598-017-12383-7>.

de Graaff, M-A., Hornslein, N., Throop, H.L., Kardol, P., van Diepen, L.T.A., 2019. Effects of agricultural intensification on soil biodiversity and implications for ecosystem functioning: a meta-analysis. *Advances in Agronomy*, 155, 1-44. <https://doi.org/10.1016/bs.agron.2019.01.001>.

Gewijzigde veldcode

Gewijzigde veldcode



deLonge, M., Basche, A. 2018. Managing grazing lands to improve soils and promote climate change adaptation and mitigation: a global synthesis. *Renewable Agriculture and Food Systems*, 33, 267-278. <https://doi.org/10.1017/S1742170517000588>.

Downie, A., Crosky, A., Munroe, P. 2009. Physical properties of biochar. In *Biochar for environmental management - Science, Technology and Implementation*. Lehmann, J. and Joseph, S. (eds), Earthscan, London, UK.

Du, Y-D., Niu, W-Q., Gu, X-B., Zhang, Q., Cui, B-J., Zhao, Y. 2018. Crop yield and water use efficiency under aerated irrigation: a meta-analysis. *Agricultural Water Management*, 210, 158-164.

<https://doi.org/10.1016/j.agwat.2018.07.038>

Edeh, I., Masek, O., Buss, W. 2020. A meta-analysis on biochar's effects on soil water properties - new insights and future research challenges. *Science of the Total Environment*, 714, 136857.

<https://doi.org/10.1016/j.scitotenv.2020.136857>

Elias, D., Wang, L., Jacinthe, P-A. 2018. A meta-analysis of pesticide loss in runoff under conventional tillage and no-till management. *Environmental Monitoring and Assessment*, 190: 79.

<https://doi.org/10.1007/s10661-017-6441-1>

Food and Agriculture Organization of the United Nations (FAO). 2010. "Climate-Smart" Agriculture: Policies, Practices and Financing for Food Security, Adaptation and Mitigation; Food and Agriculture Organization of the United Nations (FAO): Rome, Italy.

Fraser, P., Haynes, R., Williams, P. 1994. Effects of pasture improvement and intensive cultivation on microbial biomass, enzyme activities, and composition and size of earthworm populations. *Biology and Fertility of Soils*, 17, 185-190. <https://doi.org/10.1007/BF00336320>.

Gao, Y., Shao, G., Lu, J., Zhang, K., Wu, S., Wang, Z. 2020. Effects of biochar application on crop water use efficiency depend on experimental conditions: a meta-analysis. *Field Crops Research*, 249, 107763. <https://doi.org/10.1016/j.fcr.2020.107763>.

Gonzalez-Sanchez, E., Ordonez-Fernandez, R., Carbonell-Bojollo, R., Veroz-Gonzalez, O., Gil-Ribes, J. 2012. Meta-analysis on atmospheric carbon capture in Spain through the use of conservation agriculture. *Soil and Tillage Research*, 122, 52-60. <https://doi.org/10.1016/j.still.2012.03.001>.

Gravuer, K., Gennet, S., Throop, H. 2019. Organic amendment additions to rangelands: a meta-analysis of multiple ecosystem outcomes. *Global Change Biology*, 25, 1152-1170.

<https://doi.org/10.1111/gcb.14535>.

Guenet, B., Benoit, G., Chenu, C., Arrouays, D., Balesdent, J., Bernoux, M., Bruni, E., Caliman, J-P., Cardinael, R., Chen, S., Ciais, P., Desbois, D., Fouche, J., Frank, S., Henault, C., Lugato, E., Naipal, V., Nesme, T., Obersteiner, M., Pellerin, S., Powlson, D., Rasse, D., Rees, F., Soussana, J-F., Su, Y., Tian, H., Valin, H., Zhou, F. 2020. Can N<sub>2</sub>O emissions offset the benefits from soil organic carbon storage? *Global Change Biology*, 27, 237-256. <https://doi.org/10.1111/gcb.15342>.

Haddaway, N., Hedlund, K., Jackson, L., Kätterer, T., Lugato, E., Thomsen, I., Jørgensen, H., Isberg, P-E., 2017. How does tillage intensity affect soil organic carbon? A systematic review. *Environmental Evidence*, 6:30. <https://doi.org/10.1186/s13750-017-0108-9>.

Hamza, M., Anderson, W. 2005. Soil compaction in cropping systems. A review of the nature, causes and possible solutions. *Soil and Tillage Research*, 82, 121-145. <https://doi.org/10.1016/j.still.2004.08.009>.

IPCC. 2021. *Climate Change 2021: The Physical Science Basis. Contribution of Working Group I to the Sixth Assessment Report of the Intergovernmental Panel on Climate Change* [Masson-Delmotte, V., P. Zhai, A. Pirani, S.L. Connors, C. Péan, S. Berger, N. Caud, Y. Chen, L. Goldfarb, M.I. Gomis, M. Huang, K. Leitzell, E. Lonnoy, J.B.R. Matthews, T.K. Maycock, T. Waterfield, O. Yelekçi, R. Yu, and B. Zhou (eds.)]. Cambridge University Press, in press.

Jarvis, N. 2007. Review of non-equilibrium water flow and solute transport in soil macropores: principles, controlling factors and consequences for water quality. *European Journal of Soil Science*, 58, 523-546. <https://doi.org/10.1111/j.1365-2389.2007.00915.x>.

Jarvis, N., Forkman, J., Koestel, J., Kätterer, T., Larsbo, M., Taylor, A. 2017. Long-term effects of grass-clover leys on the structure of a silt loam soil in a cold climate. *Agriculture, Ecosystems and Environment*, 247, 319-328. <https://doi.org/10.1016/j.agee.2017.06.042>.

Gewijzigde veldcode

Jeffery, S., Abalos, D., Prodana, M., Bastos, A., van Groenigen, J., Hungate, B., Verheijen, F. 2017. Biochar boosts tropical but not temperate crop yields. *Environmental Research Letters* 12(5), 053001. <https://doi.org/10.1088/1748-9326/aa67bd>.

Gewijzigde veldcode

Jian, J., Lester, B., Du, X., Reiter, M., Stewart, R. 2020. A calculator to quantify cover crop effects on soil health and productivity. *Soil and Tillage Research*, 199, 104575. <https://doi.org/10.1016/j.still.2020.104575>.

Jones C., Lawton, J., Shachak, M. 1994. Organisms as ecosystem engineers. *Oikos*, 69, 373–386. <https://doi.org/10.2307/3545850>

Joseph, S., Cowie, A.L., Van Zwieten, L., Bolan, N., Budai, A., Buss, W., Cayuela, M., Graber, E., Ippolito, J., Kuzyakov, Y., Luo, Y., Ok, Y., Palansooriya, K., Shepherd, J., Stephens, S., Weng, Z., Lehmann, J. 2021. How biochar works, and when it doesn't: a review of mechanisms controlling soil and plant responses to biochar. *Global Change Biology Bioenergy*, 13, 1731-1764. <https://doi.org/10.1111/gcbb.12885>.

Kader, M., Senge, M., Mojid, M., Ito, K. 2017. Recent advances in mulching materials and methods for modifying soil environment. *Soil and Tillage Research*, 168, 155-166.

<https://doi.org/10.1016/j.still.2017.01.001>.

Gewijzigde veldcode

Kan, Z-R., Liu, Q-Y., Virk, A., He, C., Qi, J-Y., Dang, Y., Zhao, X., Zhang, H-L. 2021. Effects of experiment duration on carbon mineralization and accumulation under no-till. *Soil and Tillage Research*, 209, 104939. <https://doi.org/10.1016/j.still.2021.104939>.

Kim, N., Zabaloy, M., Guan, K., Villamil, M. 2020. Do cover crops benefit soil microbiome? A meta-analysis of current research. *Soil Biology and Biochemistry*, 142, 107701.

<https://doi.org/10.1016/j.soilbio.2019.107701>.

King, A., Blesh, J. 2018. Crop rotations for increased soil carbon: perennality as a guiding principle. *Ecological Applications*, 28, 249-261. <https://doi.org/10.1002/eap.1648>.

Gewijzigde veldcode

Koestel, J., Dathe, A., Skaggs, T., Klakegg, O., Ahmad, M., Babko, M., Gimenez, D., Farkas, C., Nemes, A., Jarvis, N. 2018. Estimating the permeability of naturally structured soil from percolation theory and pore space characteristics imaged by X-Ray. *Water Resources Research*, 54, 9255-9263. <https://doi.org/10.1029/2018WR023609>.

Kroeger, J., Pourhashem, G., Medlock, K., Masiello, C. 2020. Water cost savings from soil biochar amendment: a spatial analysis. *Global Change Biology Bioenergy*, 13, 133-142.

<https://doi.org/10.1111/gcbb.12765>

Lal, R. 2020. Soil organic matter and water retention. *Agronomy Journal*, 112, 3265-3277. <https://doi.org/10.1002/agj2.20282>.

Larsbo, M., Stenström, J., Etana, A., Börjesson, E., Jarvis, N. 2009. Herbicide sorption, degradation, and leaching in three Swedish soils under long-term conventional and reduced tillage. *Soil and Tillage Research*, 105, 200-208. <https://doi.org/10.1016/j.still.2009.08.003>.

Lee, K., Foster, R. 1991. Soil fauna and soil structure. *Australian Journal of Soil Research*, 29, 745-775. <https://doi.org/10.1071/SR9910745>.

Lee, H., Lautenbach, S., García Nieto, A., Bondeau, A., Cramer, W., Gejjendorffer, I., 2019. The impact of conservation farming practices on Mediterranean agro-ecosystem services provisioning – a meta-analysis. *Regional Environ. Change*, 19, 2187-2202. <https://doi.org/10.1007/s10113-018-1447-y>

Li, Y., Chang, S.X., Tian, L., Zhang, Q., 2018. Conservation agriculture practices increase soil microbial biomass carbon and nitrogen in agricultural soils: A global meta-analysis. *Soil Biology and Biochemistry*, 121, 50-58. <https://doi.org/10.1016/j.soilbio.2018.02.024>

Li, Y., Li, Z., Cui, S., Jagadamma, S., Zhang, Q. 2019. Residue retention and minimum tillage improve physical environment of the soil in croplands: a global meta-analysis. *Soil and Tillage Research*, 194, 104292. <https://doi.org/10.1016/j.still.2019.06.009>

Li, Y., Li, Z., Cui, S., Zhang, Q. 2020a. Trade-off between soil pH, bulk density and other soil physical properties under global no-tillage agriculture. *Geoderma*, 361, 114099. <https://doi.org/10.1016/j.geoderma.2019.114099>

Li, Y., Song, D., Liang, S., Dang, P., Qin, X., Liao, Y., Siddique, K.H.M., 2020b. Effect of no-tillage on soil bacterial and fungal community diversity: A meta-analysis. *Soil & Tillage Research*, 204, 104721. <https://doi.org/10.1016/j.still.2019.06.009>.

Li, Y., Zhang, Q., Cai, Y., Yang, Q., Chang, S.X., 2020c. Minimum tillage and residues retention increase soil microbial population size and diversity: Implications for conservation tillage. *Science of the Total Environment*, 716, 137164. <https://doi.org/10.1016/j.scitotenv.2020.137164>

Libohova, Z., Seybold, C., Wysocki, D., Wills, S., Schoeneberger, P., Williams, C., Lindbo, D., Stott, D., Owens, P. (2018). Reevaluating the effects of soil organic matter and other properties on available water-holding capacity using the National Cooperative Soil Survey Characterization Database. *Journal of Soil and Water Conservation*, 73, 411-421. <https://doi.org/10.2489/jswc.73.4.411>

Liu, Z., Dugan, B., Masiello, C., Gonnermann, H. 2017. Biochar particle size, shape, and porosity act together to influence soil water properties. *PLoS One* 12(6). e0179079. <https://doi.org/10.1371/journal.pone.0179079>

Liu, R., Thomas, B., Shi, X., Zhang, X., Wang, Z., Zhang, Y. 2021. Effects of ground cover management on improving water and soil conservation in tree crop systems: a meta-analysis. *Catena*, 199, 105085. <https://doi.org/10.1016/j.catena.2020.105085>

Lu, J., Shao, G., Cui, J., Wang, X., Keabetswe, L. 2019. Yield, fruit quality and water use efficiency of tomato for processing under regulated deficit irrigation: A meta-analysis. *Agricultural Water Management*, 222, 301-312. <https://doi.org/10.1016/j.agwat.2019.06.008>

Lu, X. 2020. A meta-analysis of the effects of crop residue return on crop yields and water use efficiency. *PLOS One*, 15, e0231740. <https://doi.org/10.1371/journal.pone.0231740>

Luo, Z., Wang, E., Sun, O. 2010. Can no-tillage stimulate carbon sequestration in agricultural soils? A meta-analysis of paired experiments. *Agriculture, Ecosystems and Environment*, 139, 224-231. <https://doi.org/10.1016/j.agee.2010.08.006>.

Mangalassery, S., Sjögersten, S., Sparkes, D., Mooney, S. 2015. Examining the potential for climate change mitigation from zero tillage. *Journal of Agricultural Science*, 153, 1151-1173. <https://doi.org/10.1017/S0021859614001002>

Marcillo, G., Miguez, F. 2017. Corn yield response to winter cover crops: an updated meta-analysis. *Journal of Soil and Water Conservation*, 72, 226-239. <https://doi.org/10.2489/jswc.72.3.226>.

McClelland, S., Paustian, K., Schipanski, M. 2021. Management of cover crops in temperate climates influences soil organic carbon stocks: a meta-analysis. *Ecological Applications*, 31, e02278. <https://doi.org/10.1002/eap.2278>

McDaniel, M., Tiemann, L., Grandy, A. 2014. Does agricultural crop diversity enhance soil microbial biomass and organic matter dynamics? A meta-analysis. *Ecological Applications*, 24, 560-570. <https://doi.org/10.1890/13-0616.1>

Medrano, H., Tomás, M., Martorell, S., Flexas, J., Hernández, E., Rosselló, J., Pou, A., Escalona, J-M., Bota, J. 2015. From leaf to whole-plant water use efficiency (WUE) in complex canopies: limitations of leaf WUE as a selection target. *The Crop Journal*, 3, 220-228.

Gewijzigde veldcode

<https://doi.org/10.1016/j.cj.2015.04.002>

Mei, K., Wang, Z., Huang, H., Zhang, C., Shang, X., Dahlgren, R., Zhang, M., Xia, F. 2018. Stimulation of N<sub>2</sub>O emission by conservation tillage management in agricultural lands: a meta-analysis. *Soil and Tillage Research*, 182, 86-93. <https://doi.org/10.1016/j.still.2018.05.006>.

Meurer, K.H.E., Haddaway, N.R., Bolinder, M.A., Kätterer, T., 2018. Tillage intensity affects total SOC stocks in boreo-temperate regions only in the topsoil – A systematic review using an ESM approach. *Earth-Science Reviews*, 177, 613-622.

<https://doi.org/10.1016/j.earscirev.2017.12.015>

Meyer, N., Bergez, J., Constantin, J., Justes, E. 2019. Cover crops reduce water drainage in temperate climates: a meta-analysis. *Agronomy for Sustainable Development*, 39, 3.

<https://doi.org/10.1016/j.agwat.2019.105998>

Mhazo, N., Chivenge, P., Chaplot, V. 2016. Tillage impact on soil erosion by water: discrepancies due to climate and soil characteristics. *Agriculture, Ecosystems and Environment*, 230, 231-241.

<https://doi.org/10.1016/j.agee.2016.04.033>

Minasny, B., McBratney, A. 2018a. Limited effect of organic matter on soil available water capacity. *European Journal of Soil Science*, 69, 39-47. <https://doi.org/10.1111/ejss.12475>.

Minasny, B., McBratney, A. 2018b. Rejoinder to the comment on: B. Minasny and A.B. McBratney. 2018. Limited effect of organic matter on soil available water capacity. *European Journal of Soil Science*, 69, 155-157. <https://doi.org/10.1111/ejss.12526>.

Mondal, S., Chakraborty, D., Bandyopadhyay, K., Aggarwal, P., Rana, D. 2020. A global analysis of the impact of zero-tillage on soil physical condition, organic carbon content, and plant root response. *Land Degradation and Development*, 31, 557-567. <https://doi.org/10.1002/ldr.3470>

Muhammad, I., Sainju, U., Zhao, F., Khan, A., Ghimire, R., Fu, X., Wang, J. 2019. Regulation of soil CO<sub>2</sub> and N<sub>2</sub>O emissions by cover crops: a meta-analysis. *Soil and Tillage Research*, 192, 103-112.

<https://doi.org/10.1016/j.still.2019.04.020>

Muhammad, I., Wang, J., Sainju, U., Zhang, S., Zhao, F., Khan, A. 2021. Cover cropping enhances soil microbial biomass and affects microbial community structure: a meta-analysis. *Geoderma*, 381, 114696.

<https://doi.org/10.1016/j.geoderma.2020.114696>

Ogle, S., Swan, A., Paustian, K. 2012. No-till management impacts on crop productivity, carbon input and soil carbon sequestration. *Agriculture, Ecosystems and Environment*, 149, 37-49.

<https://doi.org/10.1016/j.agee.2011.12.010>

Omondi, M., Xia, X., Nahayo, A., Liu, X., Korai, P., Pan, G. 2016. Quantification of biochar effects on soil hydrological properties using meta-analysis of literature data. *Geoderma*, 274, 28-34.

<https://doi.org/10.1016/j.geoderma.2016.03.029>

Palm, C., Blanco-Canqui, H., DeClerck, F., Gatere, L., Grace, P. 2014. Conservation agriculture and ecosystem services. *Agriculture, Ecosystems and Environment*, 187, 87-105.

<https://doi.org/10.1016/j.agee.2013.10.010>

Payen, F., Sykes, A., Aitkenhead, M., Alexander, P., Moran, D., MacLeod, M. 2021. Soil organic carbon sequestration rates in vineyard agroecosystems under different soil management practices: A meta-analysis. *Journal of Cleaner Production*, 290, 125736.

<https://doi.org/10.1016/j.jclepro.2020.125736>

Peixoto, D., da Silva, L., de Melo, L., Azevedo, R., Araujo, B., de Carvalho, T., Moreira, S., Curi, N., Silva, B. 2020. Occasional tillage in no-tillage systems: a global meta-analysis. *Science of the Total Environment*, 745, 140887. <https://doi.org/10.1016/j.scitotenv.2020.140887>

- Philibert A., Loyce, C., Makowski D. 2012. Assessment of the quality of meta-analysis in agronomy. *Agriculture, Ecosystems and Environment*, 148, 72–82.  
<https://doi.org/10.1016/j.agee.2011.12.003>
- Pittelkow, C., Linquist, B., Lundy, M., Liang, X., van Groenigen, K., Lee, J., van Gestel, N., Six, J., Venterea, R., van Kessel, C. 2015. When does no-till yield more? A global meta-analysis. *Field Crops Research*, 183, 156-168. <https://doi.org/10.1016/j.fcr.2015.07.020>
- Pituello, C., Dal Ferro, N., Francioso, O., Simonetti, G., Berti, A., Piccoli, I., Pisi, A., Morari, F. 2018. Effects of biochar on the dynamics of aggregate stability in clay and sandy loam soils. *European Journal of Soil Science*, 69, 827-842. <https://doi.org/10.1111/ejss.12676>
- Poeplau, C., Don, A. 2015. Carbon sequestration in agricultural soils via cultivation of cover crops – a meta-analysis. *Agriculture, Ecosystems and Environment*, 200, 33-41.  
<https://doi.org/10.1016/j.agee.2014.10.024>
- Qin, W., Assinck, F., Heinen, M., Oenema, O. 2016. Water and nitrogen use efficiencies in citrus production: a meta-analysis. *Agriculture, Ecosystems and Environment*, 222, 103-111.  
<https://doi.org/10.1016/j.agee.2016.01.052>
- Quemada, M., Baranski, M., Nobel-de Lange, M., Vallejo, A., Cooper, J. 2013. Meta-analysis of strategies to control nitrate leaching in irrigated agricultural systems and their effects on crop yield. *Agric., Ecosyst. Environ.*, 174, 1-10. <https://doi.org/10.1016/j.agee.2013.04.018>
- Rabbi, S., Minasny, B., Salami, S., McBratney, A., Young, I. 2021. Greater, but not necessarily better: the influence of biochar on soil hydraulic properties. *European Journal of Soil Science*, 72, 2033-2048.  
<https://doi.org/10.1111/ejss.13105>
- Rabot E., Wiesmeier, M., Schlüter, S., Vogel, H-J. 2018. Soil structure as an indicator of soil functions: a review. *Geoderma*, 314, 122-137. <https://doi.org/10.1016/j.geoderma.2017.11.009>
- Ranaivoson, L., Naudin, K., Ripoche, A., Affholder, F., Rabeharisoa, L., Corbeels, M. 2017. Agro-ecological functions of crop residues under conservation agriculture. A review. *Agronomy for Sustainable Development* 37, 26. <https://doi.org/10.1007/s13593-017-0432-z>
- Razzaghi, F., Obour, P., Arthur, E. 2020. Does biochar improve soil water retention? A systematic review and meta-analysis. *Geoderma*, 361, 114055. <https://doi.org/10.1016/j.geoderma.2019.114055>
- Reeleder, R., Miller, J., Ball Coelho, B., Roy, R. 2006. Impacts of tillage, cover crop, and nitrogen on populations of earthworms, microarthropods, and soil fungi in a cultivated fragile soil. *Applied Soil Ecology*, 33, 243-257. <https://doi.org/10.1016/j.apsoil.2005.10.006>
- Roarty, S., Hackett, R., Schmidt, O. 2017. Earthworm populations in twelve cover crop and weed management combinations. *Applied Soil Ecology*, 114, 142-151.  
<https://doi.org/10.1016/j.apsoil.2017.02.001>
- Schmidt, H., Kammann, C., Hagemann, N., Leifeld, J., Bucheli, T., Monedero, M., Cayuela, M. 2021. Biochar in agriculture - a systematic review of 26 global meta-analyses. *Global Change Biology Bioenergy*, 13, 1708-1730. <https://doi.org/10.1111/gcbb.12889>
- Schreefel, L., Schulte, R., de Boer, I., Pas Schrijver, A., van Zanten, H. 2020. Regenerative agriculture – the soil is the base. *Global Food Security*, 26, 100404. <https://doi.org/10.1016/j.gfs.2020.100404>
- Shackelford, G., Kelsey, R., Dicks, L. 2019. Effects of cover crops on multiple ecosystem services: ten meta-analyses of data from arable farmland in California and the Mediterranean. *Land Use Policy*, 88, 104204. <https://doi.org/10.1016/j.landusepol.2019.104204>
- Shakoor, A., Shahbaz, M., Farooq, T., Sahar, N., Shahzad, S., Altaf, M., Ashraf, M. 2021. A global meta-analysis of greenhouse gases emission and crop yield under no-tillage as compared to conventional tillage. *Science of the Total Environment*, 750, 142299.  
<https://doi.org/10.1016/j.scitotenv.2020.142299>

- Spurgeon, D.J., Keith, A.M., Schmidt, O., Lammertsma, D.R., Faber, J.H., 2013. Land-use and land management change: relationships with earthworm and fungi communities and soil structural properties. *BMC Ecology*, 13, 46. <https://doi.org/10.1186/1472-6785-13-46>
- Sun, W., Canadell, J., Yu, L., Yu, L., Zhang, W., Smith, P., Fischer, T., Huang, Y. 2020. Climate drives global soil carbon sequestration and crop yield changes under conservation agriculture. *Global Change Biology*, 26, 3325-3335. <https://doi.org/10.1111/gcb.15001>
- Tamburini, G., Bommarco, R., Cherico Wanger, T., Kremen, C., van der Heijden, M., Liebman, M., Hallin, S. 2020. Agricultural diversification promotes multiple ecosystem services without compromising yield. *Science Advances*, 6: eaba1715. <https://doi.org/10.1126/sciadv.aba1715>
- Tonitto, C., David, M., Drinkwater, L. 2006. Replacing bare fallows with cover crops in fertilizer-intensive cropping systems: a meta-analysis of crop yield and N dynamics. *Agriculture, Ecosystems and Environment*, 112, 58-72. <https://doi.org/10.1016/j.agee.2005.07.003>
- Valkama, E., Lemola, R., Känkänen, H., Turtola, E. 2015. Meta-analysis of the effects of undersown catch crops on nitrogen leaching loss and grain yields in the Nordic countries. *Agriculture, Ecosystems and Environment*, 203, 93-101. <https://doi.org/10.1016/j.agee.2015.01.023>
- Van Kessel, C., Venterea, R., Six, J., Adviento-Borbe, M., Linnquist, B., van Groenigen, K-J. 2013. Climate, duration, and N placement determine N<sub>2</sub>O emissions in reduced tillage systems: a meta-analysis. *Global Change Biology*, 19, 33-44. <https://doi.org/10.1111/j.1365-2486.2012.02779.x>
- Venter, Z., Jacobs, K., Hawkins, H-J. 2016. The impact of crop rotation on soil microbial diversity: a meta-analysis. *Pedobiologia*, 59, 215-223. <https://doi.org/10.1016/j.pedobi.2016.04.001>
- Virto, I., Barré, P., Burlot, A., Chenu, C. 2012. Carbon input differences as the main factor explaining the variability in soil organic C in no-tilled compared to inversion tilled agrosystems. *Biogeochemistry*, 108, 17-26. <https://doi.org/10.1007/s10533-011-9600-4>
- Winter, S., Bauer, T., Strauss, P., Kratschmer, S., Paredes, D., Popescu, D., Landa, B., Guzmán, G., Gómez, J.A., Guernion, M., Zaller, J.G., Batáry, P., 2018. Effects of vegetation management intensity on biodiversity and ecosystem services in vineyards: a meta-analysis. *Journal of Applied Ecology*, 55, 2484-2495. <https://doi.org/10.1111/1365-2664.13124>
- Xiong, M., Sun, R., Chen, L. 2018. Effects of soil conservation techniques on water erosion control: a global analysis. *Science of the Total Environment*, 645, 753-760. <https://doi.org/10.1016/j.scitotenv.2018.07.124>
- Yan, Y., Nakhli, S., Jin, J., Mills, G., Willson, C., Legates, D., Manahiloh, K., Imhoff, P. 2021. Predicting the impact of biochar on the saturated hydraulic conductivity of natural and engineered media. *J. Environ. Management*, 295, 113143. <https://doi.org/10.1016/j.jenvman.2021.113143>
- Yu, X., Lu, S. 2019. Reconfiguration of macropore networks in a silty loam soil following biochar addition identified by X-ray microtomography and network analyses. *European Journal of Soil Science*, 70, 591-603. <https://doi.org/10.1111/ejss.12773>
- Yu, L., Zhao, X., Gao, X., Siddique, K. 2020. Improving/maintaining water-use efficiency and yield of wheat by deficit irrigation: a global meta-analysis. *Agricultural Water Management*, 228, 105906. <https://doi.org/10.1016/j.agwat.2019.105906>
- Yu, L., Zhao, X., Gao, X., Jia, R., Yang, M., Yang, X., Wu, Y., Siddique, K. 2021. Effect of natural factors and management practices on agricultural water use efficiency under drought: a meta-analysis of global drylands. *Journal of Hydrology*, 594, 125977. <https://doi.org/10.1016/j.jhydrol.2021.125977>
- Zuber, S., Villamil, M. 2016. Meta-analysis approach to assess effect of tillage on microbial biomass and enzyme activities. *Soil Biology & Biochemistry*, 97, 176-187. <https://doi.org/10.1016/j.soilbio.2016.03.011>