

Disentangling Scatter in Long-Term Concentration-Discharge Relationships: the Role of Event Types

Felipe A. Saavedra¹, Andreas Musolff², Jana von Freyberg^{3,4}, Ralf Merz¹, Stefano Basso¹, Larisa Tarasova¹

5 ¹Department Catchment Hydrology, Helmholtz Centre for Environmental Research-UFZ, Halle (Saale), 06120 , Germany

²Department of Hydrogeology, Helmholtz Centre for Environmental Research-UFZ, Leipzig, 04318,Germany

³School of Architecture, Civil and Environmental Engineering, EPFL, 1015 Lausanne, Switzerland

⁴Mountain Hydrology and Mass Movements, Swiss Federal Institute for Forest, Snow and Landscape Research (WSL), 8903 Birmensdorf, Switzerland

10 *Correspondence to:* Felipe A. Saavedra (felipe.saavedra@ufz.de)

Abstract. Relationships between nitrate concentrations and discharge rates (C-Q) at the catchment outlet can provide insights into sources, mobilization and biogeochemical transformations of nitrate within the catchment. Nitrate C-Q relationships often exhibit considerable scatter that might be related to variable hydrologic conditions during runoff events at sampling time, corresponding to variable sources and flow paths despite similar discharge rates. Although previous studies 15 investigated the origins of this scatter in individual or in a few catchments, the role of different runoff event types across a large set of catchments is not yet fully understood.

This study combines a hydrological runoff event classification framework with low-frequency nitrate samples in 184 catchments to explore the role of different runoff events in shaping long-term C-Q relationships and their variability across contrasting catchments. In most of the catchments, snow-impacted events produce positive deviations of concentrations, 20 indicating an increased nitrate mobilization compared to the long-term pattern. In contrast, negative deviations occur mostly for rainfall-induced events with dry antecedent conditions, indicating the occurrence of lower nitrate concentrations in river flows than their long-term pattern values during this type of events. Pronounced differences in event runoff coefficients among different event types indicate their contrasting levels of hydrologic connectivity that in turn might play a key role in controlling nitrate transport due to the activation of faster flow paths between sources and streams. Using long-term, low- 25 frequency nitrate data we demonstrate that runoff event types shape observed scatter in long-term C-Q relationships according to their level of hydrologic connectivity. In addition, we hypothesize that the level of biogeochemical attenuation of catchments can partially explain the spatial variability of the scatter during different event types.

1 Introduction

Diffuse nutrient inputs in catchments are a challenge portend a challenge for water quality management (Paerl 1997; Stumpf 30 et al., 2016). Excess of nutrients as, particularly of nitrate, harms ecosystems by creating favorable conditions for

eutrophication in water bodies and leading to biodiversity loss (GEA, 2017; EEA 2019; Weitere et al., 2021). Fertilizer application on agricultural land remains the main source of nitrate contamination in human-impacted catchments, despite regulations of the past decades that stimulated a reduction of fertilizer application in Europe (Grinsven et al., 2012). Moreover, due to long-lasting legacy effects a delay in the reduction of riverine nitrate concentration was reported in many catchments (Tesoriero et al., 2013; Meter and Basu, 2017; Bieroza et al., 2018; Chang et al., 2021).

Long-term concentration-discharge (C-Q) relationships are a valuable tool for analyzing water quality gradients and trends, and for developing water management strategies (Bowes et al., 2014). The shape of C-Q relationships encodes export patterns and reflects the temporally varying quantities of critical substances such as nutrients delivered to streams (Godsey et al., 2009; Meybeck and Moatar, 2012; Rose et al., 2018). Depending on the slope of the log-log linear dependency of concentrations from discharge, three different export patterns (Godsey et al., 2009) can be defined: dilution (negative slope), enrichment (positive slope) and neutral (no relationship between C and Q or slope close to 0). Differences in long-term C-Q-relationships among catchments can be associated with differences in availability and spatial distribution of solute sources (Musolff et al., 2017; Dupas et al., 2019; Zhi et al., 2019; Casquin et al., 2021), their hydrologic connectivity (Seibert et al., 2009; Dupas et al., 2016; Covino, 2017) and biogeochemical processes within the soil and stream that can retain or permanently remove nitrate from streamwater (Mulholland et al., 2008; Dupas et al., 2016; Moatar et al., 2017; Benettin et al., 2020).

Biogeochemical processes that affect nutrient cycles in soil and water might add variability to long-term C-Q relationships. The effectiveness of the denitrification process, which removes nitrate from the soil, depends on periodic environmental factors such as temperature and soil moisture and the availability of electron donors (Korom et al., 2012; Ortmeyer et al., 2021). Instream removal processes are also more efficient during low flows and higher temperatures, adding more variability to the low-flow portion of the long-term C-Q relationships (Dehaspe et al., 2021; Moatar et al., 2017). Moreover, the availability of nitrate sources is balanced by fertilizer application and mineralization of organic nitrogen compounds and hence varies in time adding temporal variability to C-Q relationships. Timing The time of fertilizer application is often unknown, and the mineralization processes depend on chemical soil conditions and environmental factors (e.g., soil moisture and temperature) that mediate communities of microorganisms (Curtin et al., 2012; Guntiñas et al., 2012). On the other hand, average residence times of nitrate in agricultural catchments can last for decades, producing a legacy in soil (Meter et al., 2016; Puckett et al., 2011; Tesoriero et al., 2013; Vervloet et al., 2018) that can buffer the periodic effect of biogeochemical processes reducing the variability in the concentration of nitrate (Basu et al., 2011; Bieroza et al., 2018; Thompson et al., 2011).

The scatter of C-Q relationships might also be related to hydrologic conditions at the time of sampling (Knapp et al., 2020, Musolff et al., 2021), which are investigated for a large number of catchments only by a few recent studies (Minaudo et al.,

65 2019; Pohle et al., 2021). Minaudo et al. (2019) showed that in most of the 219 French catchments nitrate samples taken during baseflow conditions exhibit an enrichment export pattern, while during runoff events a neutral or opposite pattern (dilution) prevail, generating scatter in the combined long-term C-Q relationships. The cause of this scatter can be also traced to a variety of responses observed at the event-scale in several studies with high-frequency data in single or a few catchments (e.g., Bowes et al., 2015; Lloyd et al., 2016; Koenig et al., 2017; Gorski and Zimmer, 2021).

70

Our study relies on low-frequency nitrate data, which is often used to build long-term C-Q relationships (e.g. Cartwright et al., 2020, Diamond and Cohen 2018). However, studies with high-frequency data found large variability in the C-Q patterns during events (e.g. Knapp et al., 2020; Dupas et al., 2016; Vaughan et al., 2017) that might add scatter to the long-term C-Q relationship. Disparate patterns of the event C-Q relationships in a catchment over time are mainly attributed to varying 75 dominant flow sources (e.g., groundwater, shallow subsurface flow), antecedent wetness conditions (Inamdar et al., 2006; Knapp et al., 2020; Vaughan et al., 2017), time of fertilizer application (Bowes et al., 2015; Dupas et al., 2016; Outram et al., 2016), biogeochemical cycling (Heathwaite and Bieroza, 2021) and runoff event characteristics or types (Butturini et al., 2006; Bauwe et al., 2015; Chen et al., 2020; Knapp et al., 2020). For example, Winter et al. (2022) showed in a few 80 catchments located in Central Germany that runoff events generated by rainfall with dry antecedent conditions export lower nitrate concentrations due to lower hydrologic connectivity but exhibit a high variability of event C-Q slopes. In contrast, Knapp et al. (2020) showed using high-frequency concentration and discharge observations from one small forested catchment located in Switzerland that during larger runoff events with dry antecedent conditions the slopes of the event-scale C-Q relationships are more positive due to the accumulation of nitrate in the in-soil during dry periods by atmof nitrate from atmospheric deposition and the subsequent mobilization by event water. Moreover, in several catchments in US and 85 Europe, snow-induced events were found to export high nitrate concentration (Koenig et al., 2017; Inamdar et al., 2006; Casson et al., 2014). Similarly, in the previously mentioned Central German catchments Winter et al. (2022) found high nitrate concentrations and flat event C-Q slopes during snow-impacted events indicating that sufficient nitrate sources are available and most of the relevant flow paths are activated and connected to the stream during such events.

90 It was shown that hydrologic connectivity as a portion of the catchment connected to the stream via surface or subsurface pathways, increases according to the wetness state of the catchment (Blume & Van Meerveld, 2015, Jencso et al., 2009) and modulates export of nutrients at different scales. At seasonal scale nutrient transport to streams can be increased with higher hydrologic connectivity in catchments with abundant sources (Martin et al., 2004; Veith et al., 2020; Guillemot et al., 2021).. At event scale the activation of different flow paths during different levels of hydrologic connectivity evaluated using 95 shallow wells or models can partially explain changes in nitrate concentration during events (von Freyberg et al., 2014, Ocampo et al., 2006; Stieglitz et al., 2003). However, at the larger scale such observations are not available.

At catchment scale soil moisture or discharge rates are often used as proxy of hydrologic connectivity (e.g., Bracken et al., 2013; Jencso et al., 2009). Event runoff coefficient (i.e., a volumetric ratio of quick flow and input precipitation or 100 snowmelt), which represents how efficiently streamflow responds to catchment water inputs, can be also considered as its proxy (e.g., Blume et al., 2007; Outram et al., 2016; von Freyberg et al., 2014). Higher runoff coefficients are associated with wetter antecedent catchment states, indicating that such conditions favor a more efficient rainfall-runoff response (Tarasova et al., 2018, Outram et al., 2016) and possibly activation of more surface and subsurface hydrologic flow 105 pathways that facilitate fast transport of water and nutrients from the landscape to the stream (Blume & Meerveld, 2015; Hardie et al., 2011; Stieglitz et al., 2003).

New approaches to characterize and classify runoff events according to hydrologic conditions offer a possibility to 110 efficiently aggregate information about antecedent wetness state of catchments and characteristics of inducing events (e.g., rainfall, snowmelt) and to distinguish events with contrasting hydrological responses for large number of catchments (Tarasova et al., 2020). Such classification of event types combined with concentration of nitrate in stream water might 115 unravel scatter in long-term C-Q relationships as exemplified in Fig. 1. In Fig. 1a biweekly nitrate data are associated with the event type at the time of stream water sample collection. When these data are plotted in the log-log C-Q space (Fig. 1b) some event types exhibit positive (higher concentration) or negative (lower concentration) deviations from the long-term C-Q relationship. Our study aims for the first time to investigate the presence of systematic deviations in long-term C-Q relationships produced by different runoff event types for a large dataset of catchments.

We hypothesize that these deviations are related to the differences in nitrate transport during these event types and we 120 investigate such deviations from the long-term C-Q relationships in 184 German catchments. Specifically, our goal is to examine the effect of runoff event types on the observed scatter in C-Q relationships by addressing the following research questions:

- 1 Do samples collected during different event types deviate differently from the long-term C-Q relationships observed at the catchment outlets?
- 2 Which climatic and landscape characteristics explain differences in the observed C-Q deviations among German catchments?
- 125 3 Which are the potential mechanisms that explain the direction and magnitude of C-Q deviations for different event types?

Understanding the nature of nitrate deviations from the long-term C-Q relationship might provide useful information for water quality managers to reduce the risk of extreme nitrate loads to water bodies, as well as improve sampling campaigns to better capture nitrate C-Q scatter.

2.1 Study catchments and data

In this study we analyzed low frequency (biweekly to monthly) nitrate concentration data from 184 mesoscale catchments in Germany for the period from 2000 to 2015. The data were obtained from the water quality and quantity database of Germany (Musolff, 2020; Ebeling et al., 2021) in combination with a recently developed classification framework of runoff events 135 (Tarasova et al., 2020). Similar to Ebeling et al. (2021) we exclude the data prior to the 2000s to avoid impacts of improved wastewater treatment technologies in Germany. In total, we consider 33,713 nitrate samples.

Sizes of study catchment range from 95 to 23,615 km² (with a median size of 704 km²) and cover all four main German natural regions: the North German Plain, Central Uplands, South German Scarplands and Alpine Forelands (Fig. 2a). The 140 climate varies from temperate oceanic to temperate continental from West to East. Mean annual precipitation ranges from 567 mm in the Lowland northeastern catchments up to 1379 mm in the alpine catchments in the South. The predominant land use in the study catchments is agriculture, with a median coverage among catchments of 50% and a range from 13% to 84%. Median portion of catchment area covered by forest is 41% of catchment area (Fig. 2b).

145 The runoff event classification framework of Tarasova et al. (2020) considers runoff events identified from daily discharge data in catchments with no major flow regulations. The location of the discharge stations does not always coincide with water quality stations in the dataset of Ebeling et al. (2021). Both data sets are linked by pairing stations that are located on the same stream and differ less than 20% in their drainage areas. These were considered as identical outlets similarly to 150 Guillemot et al. (2021). The mean overlap between drainage areas of the corresponding outlets from the two datasets is 95% with a standard deviation of 5%.

2.2 Identification and classification of hydrological events

Runoff events and corresponding precipitation events were separated using an automated time series approach developed by Tarasova et al. (2018). The method was applied to daily discharge and precipitation data obtained from the REGNIE data set (Rauth et al., 2013). The method includes baseflow separation, precipitation attribution (i.e., corresponding inducing 155 rainfall and/or snowmelt events are linked to runoff events) and an iterative procedure to adjust site-specific thresholds for the refinement of multi-peak events. The median event duration is 12 days with a standard deviation of 7.7 days. The shortest event duration is one day, however 95% of the identified events exhibit a duration of 3 or more days. Each identified runoff event was then classified considering in the first place the nature of inducing events (rainfall, mixture of rainfall and snowmelt or rain-on-snow) (Fig. 3a) using the proportions of rainfall and snowmelt in the total volume of precipitation 160 events (Table S1). In the second step, we considered the antecedent wetness state (wet or dry) by accounting for catchment-

averaged soil moisture state prior to the event. Catchment-average snow water equivalent and soil moisture were simulated by the mHM model (Samaniego et al., 2010; Kumar et al., 2013) and provided in Zink et al. (2017). Additionally, the classification considers spatial organization of soil moisture within the catchment using spatial coefficient of variation of soil moisture, classifying events as uniform or patchy, with the latter corresponding to highly variable soil moisture within the catchment. A more detailed description of the classification framework is provided in Tarasova et al. (2020).

Runoff events and corresponding precipitation events were separated using an automated time series approach developed by Tarasova et al. (2018). The method was applied to daily discharge and precipitation data obtained from the REGNIE data set (Rauth et al., 2013). The method includes baseflow separation, precipitation attribution (i.e., corresponding inducing rainfall and/or snowmelt events are linked to runoff events) and an iterative procedure to adjust site-specific thresholds for the refinement of multi-peak events. We use daily streamflow data to identify events. This implies that only events longer than 1 day are captured. Each identified runoff event was then classified considering in the first place the nature of inducing events (rainfall, mixture of rainfall and snowmelt or rain-on-snow) (Fig. 3a) using the proportions of rainfall and snowmelt in the total volume of precipitation events (Table S1). In the second step, we considered the antecedent wetness state (wet or dry) by accounting for catchment-averaged soil moisture state prior to the event. Catchment-average snow water equivalent and soil moisture were simulated by the mHM model (Samaniego et al., 2010; Kumar et al., 2013) and provided in Zink et al. (2017). Additionally, the classification considers spatial organization of soil moisture within the catchment using spatial coefficient of variation of soil moisture, classifying events as uniform or patchy, with the latter corresponding to highly variable soil moisture within the catchment. A more detailed description of the classification framework is provided in Tarasova et al. (2020).

Each nitrate sample was linked to either no event (No.event), or to one of the five event types (Fig. 3a): rain-on-snow (Rain.on.snow), mixture of rainfall and snowmelt (Mix), rainfall during wet antecedent conditions (Rain.wet), rainfall during dry antecedent conditions with spatial uniform distribution of soil moisture (Rain.dry.uniform) and rainfall with dry antecedent condition with heterogeneous spatial distribution of soil moisture (Rain.dry.patchy). Note that we simplified the event types to increase the number of nitrate samples of each event type.

2.3 Long-term C-Q export patterns

For each catchment, the long-term C-Q relationship was derived as a linear regression between nitrate concentration (C) and discharge (Q) in the log-log space (Fig. 3b). Based on the slope of the long-term C-Q relationships (b), we grouped all study catchments according to three different long-term C-Q export patterns: dilution ($b < 0.1$) refers to a limitation of sources during high flows, enrichment ($b > 0.1$) is related to a transport limitation with abundant sources or solute uptake during low flows (Moatar et al., 2017) and neutral ($b \sim 0$) indicates no monotonic relationship between C and Q indicates a weak dependency of C and Q. As stated by Ebeling et al. (2021) this latter group exhibits largely invariable concentration with low

ratios of coefficients of variation (CVc/CVq). Three different catchments are shown as an example of each export pattern in Fig. 3c.

195 **2.4 Quantifying the deviations from long-term C-Q relationship**

For each catchment we want to quantify if samples taken at a specific event type show systematic deviations from the long-term C-Q regression compared to all samples. We quantified the deviation of each grab sample from the long-term C-Q relationship for each catchment by computing the corresponding residual concentration from the long-term C-Q linear regression line (Fig. 3b). Resulting residuals were subsequently grouped according to the hydrological event type at the time 200 of sampling.

205 Due to the variable number of grab samples attributed to different event types (Fig. S1), for each catchment, we performed a bootstrapping procedure that can explicitly handle unbalanced data by iteratively comparing two random subgroups of samples with the same size (undersampling method, e.g., Branco et al., 2015). The procedure is implemented in the following way for each catchment: n nitrate samples of a certain event type and the same number of nitrate samples from all samples (general pattern) are chosen randomly with replacement (i.e., each data point can be chosen more than once, following bootstrapping procedure). The difference of median residuals of an event type and residuals of the general pattern is then the measure of deviation of a corresponding event type from the long-term C-Q relationship (Δ_{res}). We obtained this measure 10,000 times to robustly compute its distribution (Fig. S2) and median value (Δ_{res50}). The number of samples n 210 was chosen for each catchment and event type according to the number of nitrate samples available for the corresponding event type. For each catchment, event types with less than 10 nitrate samples are excluded from the analysis. The median number of nitrate samples among all study catchments and event types is 27.

215 Due to the variable number of grab samples attributed to different event types (Fig. S1), for each catchment, we performed a bootstrapping procedure that can explicitly handle unbalanced data by iteratively comparing two random subgroups of samples with the same size (undersampling method, e.g., Branco et al., 2015). The procedure is implemented in the following way for each catchment: n nitrate samples of a certain event type and the same number of nitrate samples from all samples (general pattern) are chosen randomly with replacement (i.e., each data point can be chosen more than once, following bootstrapping procedure). The difference of median residuals of an event type and residuals of the general pattern is then the measure of deviation of a corresponding event type from the long-term C-Q relationship (Δ_{res}). We obtained this measure 10,000 times to robustly compute its distribution and median value (Δ_{res50}). The number of samples n was chosen 220 for each catchment and event type according to the number of nitrate samples available for the corresponding event type. For each catchment, event types with less than 10 nitrate samples are excluded from the analysis. The median number of nitrate samples among all study catchments and event types is 27.

225 For each catchment and from all the iterations, we obtain the median deviations between event types and the general pattern (Δ_{res50}). In order to evaluate the persistence of C-Q deviations across catchments, we tested the significance of Δ_{res50} across catchments for each event type using the non-parametric Kruskal-Wallis test (Kruskal & Wallis, 1952) at the significance level $\alpha=0.05$.

230 Low frequency datasets such as the one used in our study might contain samples collected during different phases of the event hydrograph (e.g., falling or rising limb). This might hamper the interpretability of the results due to possible bias in observed nitrate concentration linked to the time of sampling and the hysteresis effect revealed in high-frequency observations (e.g., Lloyd et al., 2016; Vaughan et al., 2017). In fact, Pohle et al. (2021) showed systematic differences in nitrate concentration between samples collected during rising and falling limbs for numerous catchments in Scotland. To 235 understand the potential effect of the hysteresis on the deviations from long-term C-Q (Δ_{res50}) we repeat the bootstrapping procedure described above considering samples collected during the rising limb, falling limb and near the event peak (near-to-peak). The rising limb of a runoff event starts at the beginning of the event and finishes one day before the day of the peak discharge. The falling limb starts one day after the day of the peak discharge and finishes at the end of the runoff event. In addition, we defined near-to-peak as samples collected from one day before to one day after the day of the peak discharge.

240 Of the total samples taken during runoff event types 34% correspond to the rising limb, 55% to the falling limb and 30% to near-to-peak. Notice that definition of near-to-peak samples allows some overlap with the other two groups of samples to use a more balanced number of samples than considering samples collected on the day of the peak of discharge only (11% of the samples were collected during the day of the peak discharge).

2.5 Catchment descriptors and relationships to C-Q deviations

245 In order to explore the differences of deviations from the long-term C-Q relationship across the catchments, we examined the Spearman rank correlation of median residuals for each catchment with various catchment descriptors. Here, we only examine catchment descriptors that were previously identified as primary controls of the nitrate C-Q export patterns in Germany (Ebeling et al., 2021). This includes topographic descriptors: median topographic wetness index, median slope and area; land cover descriptors: fraction of agriculture, forest and artificial surface; soil and aquifer descriptors: median soil 250 depth and fraction of sedimentary aquifer; nitrate sources descriptors: nitrate surplus, agricultural horizontal heterogeneity, nitrate vertical ratio; and hydrometeorological descriptors: aridity index, mean annual potential evapotranspiration, precipitation and temperature (Table S2). Detailed derivation of the above-mentioned catchment descriptors is provided in Ebeling et al. (2021).

3. Results

255 3.1 Frequency of runoff event types

Stream water samples taken during runoff event conditions account for 58% of all samples. These samples are classified to one of the five event types: 18% to Rain.dry.patchy type, 11% to Rain.dry.uniform, 15% to Rain.wet, 7% to Rain.on.snow and 7% to Mix.

260 On average across catchments, the fraction of samples taken during each event type vary at different discharge rates. Above median discharge rate, 74% of all samples correspond to an event and event types Rain.wet, Rain.on.snow and Mix occur more frequently (Fig. 4a). In contrast, only 49% of samples below median discharge rate were taken during an event and most of these grab samples correspond to Rain.dry.patchy and Rain.dry.uniform types.

265 The frequency of event types also varies seasonally (Fig. 4b). In winter most of the grab samples were taken during Rain.on.snow, Mix and Rain.wet event types. In spring months Rain.dry events become more frequent than Rain.on.snow, Mix and Rain.wet event types. During summer, most of the samples were taken either under No.event conditions or during Rain.dry.uniform and Rain.dry.patchy events. In autumn, the frequency of grab samples taken during Rain.wet, Rain.on.snow and Mix event types increases.

270 3.2 Long-term C-Q relationships and deviations during event types

We computed long-term nitrate C-Q relationships for the 184 catchments, obtaining slopes (b) from -0.6 to 1.48, with a mean of 0.13. In total, 88 study catchments exhibit neutral patterns, 80 catchments are characterized by enrichment patterns and only 16 catchments show dilution patterns. Across all catchments, the median R^2 value of the long-term C-Q relationships was low (0.14), indicating presence of a considerable scatter in [the](#) regressions.

275

We explored the residuals (res) of all nitrate data from all catchments, finding that 65% and 68% of the samples taken during Rain.on.snow and Mix event types have positive residual values respectively, indicating that concentrations were higher than the long-term log-log linear C-Q regression. In contrast, during Rain.dry.patchy and Rain.dry.uniform 69% and 60% of the samples have negative residuals. We found a less clear picture for samples taken during Rain.wet events and No.event 280 conditions with 53% and 56% of residuals positive respectively (Fig. S32).

We found strong differences in median deviations from the long-term C-Q relationships (Δ_{res50}) among different event types (Fig. 5a). Rain.on.snow and Mix event types have more often positive Δ_{res50} values (79% and 93% of the study catchments correspondingly) (Fig. 5b) when comparing across catchments. Instead, Rain.dry.patchy and Rain.dry.uniform

285 event types show negative values of Δ_{res50} more often (96% and 61% of the study catchments), with Rain.dry.patchy events showing stronger deviations. Contrasting behavior between snow-impacted events (i.e., Mix and Rain.on.snow) and rainfall events with dry antecedent wetness conditions (Rain.dry.patchy and Rain.dry.uniform) occurs across most of the study catchments independently from their long-term export pattern (Fig. 5b). For Rain.wet events deviations can be negative as well as positive (52% and 48% of study catchments respectively) with a median of Δ_{res50} across catchments close to zero
290 (Fig. 5a). For samples that were taken during No.event conditions, Δ_{res50} value is slightly positive in 85% of all catchments.

295 The sign of C-Q deviations are in line with observed nitrate concentration during different event types (Fig. S43). Negative residuals during Rain.dry.patchy and Rain.dry.uniform coincide with lower nitrate concentrations for most of the catchments independent of the long-term C-Q pattern. Similarly, during Rain.on.snow and Mix events positive C-Q deviations correspond to nitrate concentrations higher than median for most of the catchments with neutral or enrichment C-Q pattern.
300 The sign of C-Q deviations are in line with observed nitrate concentration during different event types (Fig. S3). Negative residuals during Rain.dry.patchy and Rain.dry.uniform coincide with lower nitrate concentrations for most of the catchments independent of the long-term C-Q pattern. Similarly, during Rain.on.snow and Mix events positive C-Q deviations correspond to nitrate concentrations higher than median for most of the catchments with neutral or enrichment C-Q pattern.
305 For catchment with dilution export pattern, nitrate concentration for Rain.on.snow and Mix events is similar to the average, however higher discharge in this case generate positive residuals.

305 We analyzed the influence of the sampling time within runoff events separating samples taken during the rising limb, near to the peak and falling limb. Although there are certain data limitations for a few groups of samples (gray tiles in Fig. S5b), we are able to reproduce the analysis for most of the cases. Similarly to the case when using all samples (Fig. 5b), the values of Δ_{res50} for samples taken during the rising limb, near to the peak and falling limb are mostly positive for Rain.on.snow and Mix events, negative for Rain.dry.patchy and Rain.dry.uniform. Our results confirm that the time of sampling during runoff events does not affect our findings regarding median C-Q deviations for different types of runoff events.

310 The time of sampling in runoff events did not interfere with our main results (Fig. S6b). Although data limitations for a few groups of samples (gray tiles in Fig. S6b), we could reproduce the analysis for most of the cases. We found that similarly to our results using all the samples (Fig. 5b) values of Δ_{res50} for samples taken during the rising limb, near to the peak and falling limb, are positive for Rain.on.snow and Mix events, negative for Rain.dry.patchy and Rain.dry.uniform, and intermediate for Rain.wet events.

315 Although the sign of C-Q deviation is consistent across catchments for most of the event types, the magnitude of deviation varies across catchments (Fig. 5a). The variability of Δ_{res50} expressed as interquartile ranges across catchments (boxplots in Fig. 5a) is the lowest for the samples taken during no event conditions (0.03) and Rain.wet events (0.06). The largest variability was detected for Rain.dry.patchy events (0.1), followed by Mix (0.09) and Rain.dry.uniform events (0.09).

320 Although the sign of C-Q deviation is consistent across catchments for most of the event types, the magnitude of deviation varies across catchments (Fig. 5a). The variability of Δ_{res50} expressed as interquartile ranges across catchments (boxplots in Fig. 5a) is the lowest for the samples taken during no event conditions and Rain.wet events. The largest variability was detected for Rain.dry.patchy events, followed by Mix and Rain.dry.uniform events.

3.3 Variability of C-Q deviations across German catchments

325 We analyzed the spatial variability of C-Q deviations for different event types (Fig. 5a) computing Spearman rank correlations between deviations and catchment descriptors. We found significant correlations between Δ_{res50} for each event type and catchment descriptors. Topographic properties (i.e., median slope and topographic wetness index) have the strongest correlation to the Δ_{res50} values of almost all event types (Fig. 6). Specifically, flatter catchments (low median topographic slope) with greater soil depths that are mostly located in the Northern Germany and Alpine Foreland tend to exhibit more positive residuals for Rain.wet, Rain.on.snow and Mix events, and more negative residuals for Rain.dry.patchy events and samples taken during no event conditions (Fig. 5a). Catchments with these characteristics often show high agricultural land cover (Fig. S6), however the fraction of agriculture show less significant correlations with Δ_{res50} than topographic descriptors. Moreover, in catchments with larger fractions of water-impacted soils (e.g., stagnosols, semi-terrestrial, semi-subhydric, subhydric and moor soils) we found more positive residuals for snow-impacted events (Rain.on.snow, Mix) and more negative residuals for Rain.dry.patchy events. These catchments are often located in Central 330 East or North Weast Germany.

335 We analyzed the spatial variability of C-Q deviations for different event types (Fig. 5a) computing Spearman rank correlations between deviations and catchment descriptors. We found significant correlations between Δ_{res50} for each event type and catchment descriptors. Topographic properties (i.e., median slope and topographic wetness index) have the strongest correlation to the Δ_{res50} values of almost all event types (Fig. 6). Specifically, flatter catchments (low median topographic slope) with greater soil depths that are mostly located in the Northern Germany and Alpine Foreland tend to exhibit more positive residuals for Rain.wet, Rain.on.snow and Mix events, and more negative residuals for Rain.dry.patchy events and samples taken during no event conditions (Fig. 5a). Catchments with these characteristics often also have a higher fraction of agricultural land cover (Fig. S4), however the last feature shows less significant correlations. Moreover, in catchments with larger fractions of water-impacted soils (e.g., stagnosols, semi-terrestrial, semi-subhydric, subhydric and moor soils) we found more positive residuals for snow-impacted events (Rain.on.snow, Mix) and more negative residuals for Rain.dry.patchy events. These catchments are often located in Central East or North Weast Germany.

340 Correlations between Δ_{res50} and fraction of agriculture in the catchments are less significant than those with topographic descriptors (Fig. 6). Instead, we observed strong correlations between Δ_{res50} and the fraction of forest ($p<0.01$). Forested catchments show less positive Δ_{res50} values for Rain.on.snow events and less negative values for Rain.dry.patchy.

However, we also noticed that the fraction of forest is positively correlated with topographic slope and negatively correlated with soil depth and the fraction of agriculture (Fig. S64).

Nutrient source descriptors were also significantly correlated with Δres50 . Horizontal heterogeneity of agricultural sources correlates negatively with Rain.dry.patchy residuals and vertical concentration ratio of nitrate correlates negatively with Δres50 values of Rain.dry.patchy and No.event conditions. ~~NA~~ nitrate surplus is ~~significantly~~ strongly related only to Rain.wet residuals.

3.4 Relationship between hydrologic connectivity and event type variations in residuals

We examined event runoff coefficients corresponding to different catchments and event types to link the relation between hydrologic connectivity for these event types and corresponding deviations of their samples from the long-term C-Q relationship (Fig. 7a). ~~Event runoff coefficients exhibit a larger variability across event types than across catchments for most of the catchments.~~ Catchment median event runoff coefficients exhibit a coefficient of variation of 41% across catchments. Nevertheless, variability of median runoff coefficients across of event types for single catchments is larger in most of the cases, with exhibit coefficients of variation frin different catchments from 12% to 118% and a, with a median value of 67% across catchments. We found that event types with significantly higher median runoff coefficients also exhibit significant differences in Δres50 values (Fig. 7b and 7c). Only Mix and Rain.on.snow events have similar runoff coefficients and similar Δres50 values.

4. Discussion

4.1 Direction and magnitude of C-Q deviations for different event types

We found systematic differences in the direction and magnitude of deviations of nitrate concentrations (Δres50) from the long-term C-Q relationship during different types of runoff events despite the large variety of study catchments (Fig. 5). In the following paragraphs, we will discuss potential mechanisms that can explain the variability of C-Q deviations across event types.

Positive deviations for nitrate concentrations during snow-impacted events (i.e., higher nitrate concentration compared to the general C-Q pattern) are in line with previous studies that have shown an increase of nitrate concentration in stream water during snow-impacted events in forested and agricultural catchments (Inamdar et al., 2006; Casson et al., 2014; Koenig et al., 2017). This is in line with Winter et al. (2022), who showed using high-frequency data that snow-induced events export the highest nitrate concentration compared to other event types in six German catchments with mixed land use. Our results clearly show that snowmelt does not generate lower concentration of nitrate compared to the long-term C-Q relationship.

385 although this might be expected due to lower nitrate concentration in snowfall than in stream water from agricultural catchments (Johannsen et al., 2008). Instead, higher concentration indicates enhanced nitrate transport from soil sources with no source limitation during these types of events. We argue that during snow-impacted events hydrologic connectivity of sources is high due to elevated wetness conditions (Stieglitz et al., 2003) that is consistent with previously reported high nitrate concentration during the winter period (Martin et al., 2004; Ocampo et al., 2006; Yang et al., 2018). Due to excessive catchment wetness during snow-impacted events high amount of new water transported by faster and shallower pathways can reach the stream (a so-called inverse storage effect; Fang et al., 2019) mobilizing large amounts of nitrate available in the soil (Yang et al., 2018). In addition, the mobilized water during this events is less affected by biogeochemical processes due to lower microbial activity induced by low temperature during snow-impacted events (Johannsen et al., 2008).

390 Positive deviations for nitrate concentrations during snow-impacted events (i.e., higher nitrate concentration compared to the general C-Q pattern) are in line with previous studies that have shown an increase of nitrate concentration in stream water during snow-impacted events in forested and agricultural catchments (Inamdar et al., 2006; Casson et al., 2014; Koenig et al., 2017). This is in line with Winter et al. (2022), who showed using high-frequency data that snow-induced events export the highest nitrate concentration compared to other event types in six German catchments with mixed land use. Our results 395 clearly show that snowmelt does not generate lower concentration of nitrate compared to the long-term C-Q relationship, although this might be expected due to lower nitrate concentration in snowfall than in stream water from agricultural catchments (Johannsen et al., 2008). Instead, higher concentration indicates enhanced nitrate transport from soil sources with no source limitation during these types of events. We argue that during snow-impacted events hydrologic connectivity of sources is high due to elevated wetness conditions (Stieglitz et al., 2003) that is consistent with previously reported high 400 nitrate concentration during the winter period (Martin et al., 2004; Ocampo et al., 2006; Yang et al., 2018). Due to excessive catchment wetness during snow-impacted events high amount of new water transported by faster and shallower pathways can reach the stream (a so-called inverse storage effect; Fang et al., 2019) mobilizing large amounts of nitrate available in the soil (Yang et al., 2018). In addition, the mobilized water during this events is less affected by biogeochemical processes due to lower microbial activity induced by low temperature during snow-impacted events.

405 Furthermore, our analysis shows that Rain.dry.uniform and Rain.dry.patchy events generate lower nitrate concentrations compared to the other types of events or no event conditions (Fig. S4), producing strong negative C-Q deviations (Fig. 5). Along the same lines, Winter et al. (2022) showed that runoff events with dry antecedent conditions exhibit lower concentration compared to other event types in six German catchments with mixed land use. There are two possible 410 explanations for the occurrence of this phenomenon. On one hand, Rain.dry.uniform and Rain.dry.patchy events occur more often during the dry season, when nitrate concentrations are reported to be lower (House et al., 2001; Guillemot et al., 2021) due to a hydrological disconnection between agricultural sources and streams under dry conditions, as well as higher biogeochemical nitrate removal processes, including biotic uptake and denitrification (Mulholland et al., 2008; Rode et al., 2016, Lutz et al., 2020, Johannsen et al., 2008). On the other hand, during runoff events with dry antecedent conditions

415 nitrate concentrations can be diluted below pre-event concentration level. This is shown by high-frequency observations in agricultural catchments that report more frequent negative event C-Q slopes during the dry season (Winter et al., 2021, 2022; Zhang et al., 2020; Koenig et al., 2017). In such cases nitrate concentration decreases compared to pre-event concentrations due to hydrologic disconnection between streams and agricultural land, and growing importance of runoff generated from riparian zones (Fang et al., 2019; Grayson et al., 1997; McGlynn & Seibert, 2003), which are known to buffer nitrate inputs due to high denitrification potential (Ocampo et al., 2006; Cole et al., 2020; Sabater et al., 2003). Our results show that the combined effect of lower pre-event concentration and further decrease in concentrations due to runoff events magnify the observed negative deviations of nitrate samples from the long-term C-Q relationships. Nevertheless, the data available for this study does not allow to quantify the contribution of individual effects of these two factors on the scatter of long-term C-Q relationships of nitrate. On the contrary, studies in pristine headwaters and forested catchments found that rainfall events with dry antecedent conditions can mobilize large amounts of nitrate increasing the concentration in streams (Knapp et al., 2020; Koenig et al., 2017). Since these findings are based on the observations in a single or only a few catchments with limited agricultural activity, different nitrate sources, such as atmospheric deposition or nitrate fixation and nitrate accumulation in soil between events, might be more relevant. Agriculture is a dominant land use type in the catchments used in this study (median fraction of agricultural land is 50%), therefore a considerable nitrate accumulation in soil as the result of fertilization dominates over any other nitrate source (Häussermann et al., 2020; Lassaletta et al., 2014), explaining the discrepancy between our findings and the results from pristine headwaters and forested catchments on the role of rainfall events with dry antecedent conditions for nitrate mobilization.

435 Differently from runoff events with dry antecedent conditions, we found that nitrate grab samples taken during no event conditions exhibit slightly positive deviations, indicating higher concentrations compared to the long-term C-Q relationships. No event samples also exhibit higher nitrate concentrations (Fig. S4) compared to Rainfall events with dry antecedent conditions (i.e., Rain.dry.patchy and Rain.dry.uniform), with both group of samples being collected during relatively low discharge conditions (Fig. 4a). This suggest that the lack of dilution during no events might produce more positive residuals.

440 Furthermore, our analysis shows that Rain.dry.uniform and Rain.dry.patchy events generate lower nitrate concentrations compared to the other event types or no event conditions (Fig. S3), producing strong negative C-Q deviations (Fig. 5). Along the same lines, Winter et al. (2022) showed that runoff events with dry antecedent conditions exhibit the lowest concentration compared to other event types in six German catchments with mixed land use. Lower nitrate concentration during runoff events with dry antecedent conditions can be explained by low pre-event conditions linked to hydrological and biogeochemical drivers in addition to possible dilution during runoff events.

445 In fact, Rain.dry.uniform and Rain.dry.patchy events occur more often during the dry season, when nitrate concentrations are reported to be lower in several studies (House et al., 2001; Guillemot et al., 2021). Nitrate concentrations in streams decrease due to a lower hydrologic connectivity between sources and streams, as well as higher biogeochemical nitrate removal

450 (Mulholland et al., 2008; Rode et al., 2016; Lutz et al., 2020; Johannsen et al., 2008), producing low pre-event nitrate concentration. Indeed, instream processes that remove nitrate can notably reduce nitrate concentrations during low flows due to a more efficient removal often altering the linearity of nitrate C-Q relationships (Moatar et al., 2018; Dehaspe et al., 2021).

455 Moreover, during runoff events with dry antecedent conditions nitrate concentrations can decrease below pre-event concentration level. This is shown in studies relying on high frequency data in agricultural catchments that reported more frequent negative event C-Q slopes during the dry season (Winter et al., 2021, 2022; Zhang et al., 2020; Koenig et al., 2017). Nitrate concentration decreases from pre-event concentration during these type of events due to hydrologic disconnection between streams and agricultural land, and growing importance of runoff generated from riparian zones closer to the river (Fang et al., 2019; Grayson et al., 1997; McGlynn & Seibert, 2003), which are known to buffer nitrate inputs due to high denitrification potential (Ocampo et al., 2006; Cole et al., 2020; Sabater et al., 2003). The combined effect of lower pre-event concentration and further decrease in concentrations due to runoff events magnify the observed negative deviations of nitrate samples from the long-term C-Q relationship. Nevertheless, our data does not allow for quantifying the individual effect of these two factors on the scatter of long-term C-Q relationships of nitrate.

460 465 Studies in pristine headwaters and forested catchments found that rainfall events with dry antecedent conditions can mobilize large amounts of nitrate increasing the concentration in streams (Knapp et al., 2020; Koenig et al., 2017). Since these findings are based on the observations in a single or only a few catchments with limited agricultural activity, different nitrate sources, such as atmospheric deposition or nitrate fixation and nitrate accumulation in soil between events, might be more relevant. In our study catchments agriculture is a dominant land use type (median fraction of agricultural land is 50%), therefore a considerable nitrate accumulation in soil as the result of fertilization dominates over any other nitrate source (Häussermann et al., 2020; Lassaletta et al., 2014), explaining the discrepancy between our findings and the results from pristine headwaters and forested catchments on the role of rainfall events with dry antecedent conditions for nitrate mobilization.

470 475 Differently from runoff events with dry antecedent conditions, we found that nitrate grab samples taken during no event conditions exhibit slightly positive deviations, indicating higher concentrations compared to the long-term C-Q relationships. Most of nitrate samples during no event conditions coincide with low rates of discharge (Fig. 4a) as well as Rainfall events with dry antecedent conditions (i.e., Rain:dry:patchy and Rain:dry:uniform). Nevertheless, no event samples exhibit higher nitrate concentrations (Fig. S3) compared to these events independently of the long-term C-Q export pattern indicating that 480 the lack of dilution during no events might produce more positive residuals.

4.2 The role of hydrologic connectivity between different event types

The hypothesized role of hydrologic connectivity on shaping nitrate deviations during runoff events is supported by the relation between event runoff coefficient and the deviation of nitrate concentrations from the long-term C-Q relationship for different event types (Fig. 7). Higher runoff coefficients indicate a more efficient rainfall-runoff response, either due to the 485 activation of stored water or the fast runoff of rainwater or snowmelt into the stream. Across all studied catchments the highest runoff coefficients are consistently found for snow-impacted events (Rain.on.snow and Mix) (Fig. 7). High values of runoff coefficients were connected to highly positive residuals indicating that compared to the C-Q relationship more nitrate was mobilized during high levels of hydrologic connectivity. Studies using high-frequency data show that during runoff events in wet seasons, when catchments are hydrologically more connected, shallow flow paths are activated transporting 490 greater amounts of nitrate (Inamdar et al., 2006; Outram et al., 2016; Schwientek et al., 2013). Similarly, von Freyberg et al. (2014), Ocampo et al. (2006) and Stieglitz et al. (2003) showed that upland zones are more efficiently connected to riparian zones by shallow pathways during wet months, permitting effective transport of nutrients to the stream.

There are no significant differences in event runoff coefficient between two types of snow-impacted events (i.e., Mix and 495 Rain.on.snow) are (Fig. 7), despite possible differences in their characteristic snowmelt intensities (Tarasova et al., 2020). While during Mix events the melting of the snowpack is induced by temperature increase only, during rain-on-snow events additional portions of snowpack might be melted by additional energy brought by rainfall (Cohen et al., 2015). However, in both cases event runoff coefficients are similarly high (Fig. 7b) suggesting similar hydrologic connectivity (Blume & Meerveld, 2015). This, in turn, results in similar deviations of nitrate concentrations from the long-term C-Q relationships 500 for these two snow-impacted event types, again highlighting the primary role of hydrologic connectivity on the observed deviations of C-Q relationships.

In contrast to snow-impacted events, lower runoff coefficients typical for rainfall events with dry antecedent conditions (i.e., Rain.dry.patchy and Rain.dry.uniform) indicate that a small portion of event water can reach the stream, which means that 505 distant zones from the stream network associated with longer pathways exhibit lower or lack of connection during these types of runoff events. Moreover, the overall dry antecedent conditions with a heterogeneous spatial distribution of soil moisture indicate a potential disconnection of runoff generation zones, and therefore the hydrologic connectivity might be lower during such events (i.e., Rain.dry.patchy) than during events with uniform spatial distribution of soil moisture (i.e., Rain.dry.uniform) as shown by lower event runoff coefficients of the former (Fig. 7). Differences in the connectivity of these 510 two types of events are also in line with differences in residuals with the strongest negative residuals observed for Rain.dry.patchy events. Similarly, Outram et al. (2016) showed that lower event runoff coefficients during runoff events with dry antecedent conditions mobilize only a small quantity of nitrate due to the inactive subsurface pathways. Studies using high-frequency nitrate and discharge data showed that during dry periods upland and riparian zones are usually

disconnected (Ocampo et al., 2006; Stieglitz et al., 2003; von Freyberg et al., 2014). These evidence from single catchments

515 are in line with our results across a large set of German catchments suggesting a crucial role of hydrologic connectivity for
nutrient transport.

During Rain.wet events runoff coefficients are between those of snow-impacted events and rainfall events with dry
antecedent conditions (Fig. 7) which indicate intermediate level of hydrologic connectivity between sources and streams, and

520 thus both positive and negative C-Q deviations (Fig. 5).

4.3 Climatic and landscape controls of the variability of C-Q deviations across Germany

Despite systematic differences of C-Q deviations for different event types, we found considerable spatial variability in the
magnitude of these deviations across German catchments (Fig. 5a). In the next paragraphs, we discuss how catchment
characteristics may control deviations of concentrations taken during events from the long-term C-Q relationship (Δ_{res50}).

525

The correlation of C-Q deviations during snow-impacted events with topographic and soil properties (Fig. 6) indicates that in flatter catchments with thick soils and a high fraction of sedimentary aquifers these types of events generate higher nitrate concentrations compared to the average long-term behavior. Contrarily, C-Q deviations during Rain.dry.patchy are more negative indicating that nitrate concentrations during these events tend to be smaller than the long-term average in catchments with these characteristics. Previous studies have shown how these characteristics are able to promote nitrate removal in catchments. Deep sedimentary aquifers have a high potential of denitrification due to a great availability of electron donors, longer transit times and more anoxic conditions due to sufficient reduction capacity (Kunkel et al., 2004; Wendland et al., 2008; Knoll et al., 2020) generating a lower nitrate supply in deeper soils compared to shallow soil (Dupas et al., 2016). In addition, flat catchments (low topographic slope, higher topographic wetness index) tend to have a higher portion of riparian wetlands (Musolff et al., 2018) that can reduce nitrate concentrations from stream water. During snow-impacted events fast flow pathways between nitrate sources and the stream are activated and nitrate stored in shallow soils can be mobilized bypassing the denitrification attenuation of the soil and the riparian zone, which is also largely suppressed during low temperatures in winter (Johannsen et al., 2008; Lutz et al., 2020), resulting in higher nitrate C-Q deviations. In contrast, Rain.dry.patchy events might mobilize water from connected near-stream source zones, where riparian wetlands from flatter areas contribute water with low nitrate concentration (Fovet et al., 2018; Lutz et al., 2020) generating more negative C-Q deviations. In addition, deviations during these events might be associated with longer transit times due to thicker soil and less hydrologic connectivity (Yang et al., 2018) which can reduce nitrate concentrations in streams. Instead, steeper catchments with shallow soils during Rain.dry.patchy events show less nitrate attenuation due to shorter flow paths and less favorable conditions for denitrification, generating relatively higher streamwater nitrate concentrations during these events and therefore decreasing the magnitude of C-Q deviations.

We acknowledge that some catchment characteristics are highly correlated (Fig. S6). Flatter catchments often exhibit higher fractions of agriculture, therefore more diffuse source availability. Although the correlation of the fraction of agriculture and C-Q deviations during Rain.on.snow events was less significant than topographic descriptors, a potential increment of diffuse sources in flatter catchments might also enhance the mechanism of nitrate bypassing the buffer capacity of catchments during Rain.on.snow events generating higher C-Q deviations. For most of the event types we found that the fraction of agriculture itself is not sufficient to explain the differences in nitrate deviations from the long-term C-Q relationship between catchments (Fig. 6). However, the vertical ratio of nitrate between topsoil and groundwater and horizontal spatial distribution of agricultural land within the catchment (i.e., horizontal heterogeneity) were strongly correlated with C-Q deviations for Rain.dry.patchy events (Fig. 6). During Rain.dry.patchy events the mobilization of distant nitrate sources (horizontally and vertically) is reduced due to the low hydrologic connectivity resulting in lower nitrate concentration of stream water and more negative deviations in catchments with top loaded nitrate profile as well as with more distant agricultural lands from streams. The high spatial variability of agricultural nitrate sources, expressed as horizontal heterogeneity and vertical ratio of nitrate, and the temporal variability of sources possibly induced by elevated subsurface and riparian zone removal during different levels of hydrologic connectivity, promotes deviations of nitrate concentrations from the long-term C-Q relationships.

The correlation of C-Q deviations during snow-impacted events with topographic and soil properties (Fig. 6) indicates that in flatter catchments with thick soils and a high fraction of sedimentary aquifers these types of events generate higher nitrate concentrations compared to the average long-term behavior. Contrarily, C-Q deviations during Rain.dry.patchy are more negative indicating that nitrate concentrations during these events tend to be smaller than the long-term average in catchments with these characteristics. Previous studies have shown how these characteristics are able to promote nitrate removal in catchments. Deep sedimentary aquifers have a high potential of denitrification due to a great availability of electron donors, longer transit times and more anoxic conditions due to sufficient reduction capacity (Kunkel et al., 2004; Wendland et al., 2008; Knoll et al., 2020) generating a lower nitrate supply in deeper soils compared to shallow soil (Dupas et al., 2016). In addition, flat catchments (low topographic slope, higher topographic wetness index) tend to have a higher portion of riparian wetlands (Musolff et al., 2018) that can reduce nitrate concentrations from stream water. During snow-impacted events fast flow pathways between nitrate sources and the stream are activated and nitrate stored in shallow soils can be mobilized bypassing the denitrification attenuation of the soil and the riparian zone, which is also largely suppressed during low temperatures in winter (Johannsen et al., 2008; Lutz et al., 2020), resulting in higher nitrate C-Q deviations. We acknowledge that catchment characteristics might be highly correlated (Fig. S4). Flatter catchments often exhibit higher fractions of agriculture, therefore more diffuse source availability. Although the correlation of the fraction of agriculture and C-Q deviations during Rain.on.snow events was less significant than topographic descriptors, a potential increment of diffuse sources in flatter catchments might also enhance the mechanism of nitrate bypassing the buffer capacity of catchments during Rain.on.snow events generating higher C-Q deviations.

In contrast, Rain.dry.patchy events might mobilize water from connected near-stream source zones, where riparian wetlands from flatter areas contribute water with low nitrate concentration (Fovet et al., 2018; Lutz et al., 2020) generating more negative C-Q deviations. In addition, deviations during these events might be associated with longer transit times due to thicker soil and less hydrologic connectivity (Yang et al., 2018) which can reduce nitrate concentrations in streams. We

590 acknowledge that catchment characteristics might be highly correlated (Fig. S4). Flatter catchments often exhibit higher fractions of agriculture, therefore more diffuse source availability. Although the correlation of the fraction of agriculture and C-Q deviations during Rain.on.snow events was less significant than topographic descriptors, a potential increment of diffuse sources in flatter catchments might also enhance the mechanism of nitrate bypassing the buffer capacity of catchments during Rain.on.snow events generating higher C-Q deviations. In contrast, steeper catchments with shallow soils during Rain.dry.patchy events show less nitrate attenuation due to less favorable conditions for denitrification and shorter flow paths, generating relatively higher streamwater nitrate concentrations during these events and therefore decreasing the magnitude of C-Q deviations.

595 Many studies have highlighted the importance of agricultural sources for nitrate export patterns in several catchments (e.g., Moatar et al., 2017; Minaudo et al., 2019). We found that the fraction of agriculture itself is not sufficient to explain the differences in nitrate deviations from the long-term C-Q relationship between catchments for most of the event types (Fig. 6). However, the vertical ratio of nitrate between topsoil and groundwater and horizontal spatial distribution of agricultural land within the catchment (i.e., horizontal heterogeneity) were strongly correlated with C-Q deviations for Rain.dry.patchy events (Fig. 6). During Rain.dry.patchy events the mobilization of distant nitrate sources (horizontally and vertically) is reduced due to the low hydrologic connectivity resulting in lower nitrate concentration of stream water and more negative deviations in catchments with top loaded nitrate profile as well as with more distant agricultural lands from streams. The high spatial variability of agricultural nitrate sources, expressed as horizontal heterogeneity and vertical ratio of nitrate, and the temporal variability of sources possibly induced by elevated subsurface and riparian zone removal during different levels 600 of hydrologic connectivity, promotes deviations of nitrate concentrations from the long-term C-Q relationships.

605

4.4 Implications of this study

In this study we performed the first large-scale analysis of long-term nitrate C-Q relationships differentiating runoff event types. We show that flatter catchments with soil conditions favorable for denitrification or distant nitrate sources are prone to generate disproportional loads during runoff events with high levels of hydrologic connectivity, presenting an ecological risk 610 for aquatic ecosystems. These findings can be instructive for implementing more effective water quality management strategies to prevent extreme nitrate loads reaching water bodies in such catchments during events associated with high levels of hydrologic connectivity (i.e., snow-impacted events).

The connection between nitrate concentrations and different types of runoff events shown in our study indicates that possible 615 changes in the occurrences of different event types due to the ongoing climate change might in turn affect the dynamics of nutrient exports in the catchments. With advancing climate change air temperature is projected to further increase further leading to a substantial decline in seasonal snowpack accumulation and earlier snowmelt onset in Central Europe (IPCC, 2021). Several studies reported a redA reduction in the frequency of snow accumulatio-impacted events was already shown

in Germany over the last decades (Fontrodona Bach et al., 2018; Chan et al., 2020; Taszarek et al., 2020) with a consistent reduction in the frequency of rain-on-snow events (Cohen et al., 2015), suggesting that the corresponding positive deviations from the long-term nitrate C-Q relationship ~~associated with snow-impacted events~~ are likely to occur less often in the future. Less frequent snow-impacted events would reduce nitrate mobilization from the soil under these critical event conditions. Consequently, more nitrate may remain in the soil sources. A fraction of this soil nitrate is expected to be removed by denitrification whereas another fraction may last longer as soil nitrate legacy (Dupas et al., 2020; Meter et al., 2016), thus generating unknown long-term effects in the nitrate dynamics during future runoff events. On the other hand, higher temperatures lead to a decrease of soil moisture (Dai et al., 2004), propitiating dry conditions and reducing hydrologic connectivity. An increase in frequency of rainfall events with dry antecedent conditions observed in several German catchments (Winter et al., 2022) indicates that negative deviations might become even more frequent during warm seasons in the future.

By using low-frequency, long-term nitrate data we were able to provide information about characteristic nitrate transport during different types of events and to identify hydrologic connectivity associated with these types as a critical control of nitrate dynamics in German catchments. Our findings using low-frequency data are largely supported by the detailed analysis of high-frequency data in individual catchments from the previous studies: but thanks to the large number of analyzed catchments allow for a more comprehensive analysis of systematic deviations of nitrate concentrations during events of different types and provide valuable insights on the origins of the scatter in C-Q relationships. The abundance of low frequency data worldwide and transferable nature of the applied event classification framework provide the means of further applications in contrasting environments to better understand nitrate long-term C-Q relationships across contrasting environments. Moreover, our results suggest that sampling campaigns should be designed specifically to capture runoff events with different levels of hydrologic connectivity in order to better explain the scatter in long-term C-Q relationships and better isolate the role of singular processes (i.e., nitrate uptake, denitrification).

Although the presence of the event-scale hysteresis effect might considerably affect nitrate concentration during rising and falling limbs of the event hydrograph in some catchments (Pohle et al, 2021) we found a similar direction of deviations from the long-term C-Q relationships when we considered samples taken during rising limb, falling limb and near to the peak (Fig S6b). Hence, our results suggest that the variability potentially added by the presence of hysteresis patterns is lower than the deviations observed for different event types from the long-term C-Q relationship. Increasing availability of high-frequency datasets coupled with new statistical modeling approaches might be used in the future to evaluate hysteresis-related effects in the existing long-term C-Q datasets to further disentangle inter- and intra-event variability of nitrate dynamics at larger scales.

5. Conclusions

We analyzed for the first time the effect of different runoff event types on the scatter observed in concentration-discharge (C-Q) relationship across 184 German catchments. Specifically, we examined the deviations of the concentration of nitrate samples collected during different runoff event types from the long-term C-Q relationships. Our results highlight pronounced deviations in most of the catchments regardless of their overall long-term C-Q export patterns (dilution, neutral, or enrichment). Thus, scatter apparent in long-term C-Q relationships can indeed be partially explained by different types of runoff event conditions.

We analyzed for the first time the effect of different runoff event types on the scatter observed in concentration-discharge (C-Q) relationship across 184 German catchments. Specifically, we examined the deviations of the concentration of nitrate samples collected during different runoff event types from the long-term C-Q relationships. Our results highlight pronounced deviations in most of the catchments regardless of their overall long-term C-Q export patterns (dilution, neutral, or enrichment). Thus, scatter apparent in long-term C-Q relationships can indeed be partially explained by different types of runoff event conditions. Moreover, we inferred using catchment descriptors physical mechanisms that explain the spatial variability of this scatter.

We found that nitrate transport is enhanced during snow-impacted events compared to long-term C-Q relationships. On the other hand, nitrate concentrations tend to be lower than the long-term C-Q relationships when rainfall coincides with dry antecedent conditions. The C-Q relationships during rainfall on wet antecedent conditions were not significantly different from the long-term relationship. We argue that hydrologic connectivity to the nitrate sources, here represented by the values of event runoff coefficient, is crucial to explain deviations from the long-term C-Q relationship during different event types.

Finally, we found that flatter catchments with high denitrification potential (i.e., deep soils, presence of sedimentary aquifers), as well as catchments with agricultural areas located farther from the stream or with top-loaded nitrate profile, exhibit an enhanced nitrate transport during snow-impacted events and lower nitrate concentrations during events induced by rainfall with dry antecedent condition compared to the long-term C-Q relationship. Catchments with these characteristics are prone to generate disproportional loads during snow-impacted events, exacerbating ecological risk for receiving water bodies. Findings from this study improve our understanding of the effects of runoff event types on nutrient dynamics and provide valuable insights for optimizing water quality management and monitoring.

References

Basu NB, Thompson SE, Rao PSC. 2011. Hydrologic and biogeochemical functioning of intensively managed catchments: A synthesis of top-down analyses. *Water Resources Research* **47** (10) DOI: 10.1029/2011WR010800

Bauwe A, Tiemeyer B, Kahle P, Lennartz B. 2015. Classifying hydrological events to quantify their impact on nitrate leaching across three spatial scales. *Journal of Hydrology* **531**: 589–601 DOI: 10.1016/j.jhydrol.2015.10.069

Benettin P, Fovet O, Li L. 2020. Nitrate removal and young stream water fractions at the catchment scale. *Hydrological Processes* **34** (12): 2725–2738 DOI: 10.1002/hyp.13781

Bieroza MZ, Heathwaite AL, Bechmann M, Kyllmar K, Jordan P. 2018. The concentration-discharge slope as a tool for water quality management. *Science of The Total Environment* **630**: 738–749 DOI: 10.1016/j.scitotenv.2018.02.256

Blume T, Meerveld HJ (Ilja) van. 2015. From hillslope to stream: methods to investigate subsurface connectivity. *WIREs Water* **2** (3): 177–198 DOI: 10.1002/wat2.1071

Blume T, Zehe E, Bronstert A. 2007. Rainfall—runoff response, event-based runoff coefficients and hydrograph separation. *Hydrological Sciences Journal* **52** (5): 843–862 DOI: 10.1623/hysj.52.5.843

Bowes MJ, Jarvie HP, Halliday SJ, Skeffington RA, Wade AJ, Loewenthal M, Gozzard E, Newman JR, Palmer-Felgate EJ. 2015. Characterising phosphorus and nitrate inputs to a rural river using high-frequency concentration-flow relationships. *The Science of the Total Environment* **511**: 608–620 DOI: 10.1016/j.scitotenv.2014.12.086

Bowes MJ, Jarvie HP, Naden PS, Old GH, Scarlett PM, Roberts C, Armstrong LK, Harman SA, Wickham HD, Collins AL. 2014. Identifying priorities for nutrient mitigation using river concentration–flow relationships: The Thames basin, UK. *Journal of Hydrology* **517**: 1–12 DOI: 10.1016/j.jhydrol.2014.03.063

Bracken LJ, Wainwright J, Ali GA, Tetzlaff D, Smith MW, Reaney SM, Roy AG. 2013. Concepts of hydrological connectivity: Research approaches, pathways and future agendas. *Earth-Science Reviews* **119**: 17–34 DOI: 10.1016/j.earscirev.2013.02.001

Branco P, Torgo L, Ribeiro R. 2015. A Survey of Predictive Modelling under Imbalanced Distributions. *arXiv:1505.01658 [cs]* Available at: <http://arxiv.org/abs/1505.01658> [Accessed 17 November 2021]

Butturini A, Gallart F, Latron J, Vazquez E, Sabater F. 2006. Cross-site Comparison of Variability of DOC and Nitrate c-q Hysteresis during the Autumn–winter Period in Three Mediterranean Headwater Streams: A Synthetic Approach. *Biogeochemistry* **77** (3): 327–349 DOI: 10.1007/s10533-005-0711-7

Casquin A, Dupas R, Gu S, Couic E, Gruau G, Durand P. 2021. The influence of landscape spatial configuration on nitrogen and phosphorus exports in agricultural catchments. *Landscape Ecology* **36** (12): 3383–3399 DOI: 10.1007/s10980-021-01308-5

Casson NJ, Eimers MC, Watmough SA. 2014. Sources of nitrate export during rain-on-snow events at forested catchments. *Biogeochemistry* **120** (1): 23–36 DOI: 10.1007/s10533-013-9850-4

Chan SC, Kendon EJ, Berthou S, Fosser G, Lewis E, Fowler HJ. 2020. Europe-wide precipitation projections at convection permitting scale with the Unified Model. *Climate Dynamics* **55** (3): 409–428 DOI: 10.1007/s00382-020-05192-8

Chang S, Zhang Q, Byrnes D, Basu N, Van Meter K. 2021. Chesapeake legacies: The importance of legacy nitrogen to improving Chesapeake Bay water quality. *Environmental Research Letters* **16** DOI: 10.1088/1748-9326/ac0d7b

Chen X, Parajka J, Széles B, Valent P, Viglione A, Blöschl G. 2020. Impact of Climate and Geology on Event Runoff Characteristics at the Regional Scale. *Water* **12** (12): 3457 DOI: 10.3390/w12123457

Cohen J, Ye H, Jones J. 2015. Trends and variability in rain-on-snow events. *Geophysical Research Letters* **42** (17): 7115–7122 DOI: 10.1002/2015GL065320

Cole LJ, Stockan J, Helliwell R. 2020. Managing riparian buffer strips to optimise ecosystem services: A review. *Agriculture, Ecosystems & Environment* **296**: 106891 DOI: 10.1016/j.agee.2020.106891

Covino T. 2017. Hydrologic connectivity as a framework for understanding biogeochemical flux through watersheds and along fluvial networks. *Geomorphology* **277**: 133–144 DOI: 10.1016/j.geomorph.2016.09.030

Curtin D, Beare MH, Hernandez-Ramirez G. 2012. Temperature and Moisture Effects on Microbial Biomass and Soil Organic Matter Mineralization. *Soil Science Society of America Journal* **76** (6): 2055–2067 DOI: 10.2136/sssaj2012.0011

Dai A, Trenberth KE, Qian T. 2004. A Global Dataset of Palmer Drought Severity Index for 1870–2002: Relationship with Soil Moisture and Effects of Surface Warming. *Journal of Hydrometeorology* **5** (6): 1117–1130 DOI: 10.1175/JHM-386.1

Dehaspe J, Sarrazin F, Kumar R, Fleckenstein J, Musolff A. 2021. Bending of the concentration discharge relationship can inform about in-stream nitrate removal. *Hydrology and Earth System Sciences* **25**: 6437–6463 DOI: 10.5194/hess-25-6437-2021

Dupas R, Abbott BW, Minaudo C, Fovet O. 2019. Distribution of Landscape Units Within Catchments Influences Nutrient Export Dynamics. *Frontiers in Environmental Science* **7** DOI: 10.3389/fenvs.2019.00043

Dupas R, Ehrhardt S, Musolff A, Fovet O, Durand P. 2020. Long-term nitrogen retention and transit time distribution in agricultural catchments in western France. *Environmental Research Letters* **15** (11): 115011 DOI: 10.1088/1748-9326/abbe47

Dupas R, Jomaa S, Musolff A, Borchardt D, Rode M. 2016. Disentangling the influence of hydroclimatic patterns and agricultural management on river nitrate dynamics from sub-hourly to decadal time scales. *Science of The Total Environment* **571**: 791–800 DOI: 10.1016/j.scitotenv.2016.07.053

Ebeling P, Kumar R, Weber M, Knoll L, Fleckenstein JH, Musolff A. 2021. Archetypes and Controls of Riverine Nutrient Export Across German Catchments. *Water Resources Research* **57** (4): e2020WR028134 DOI: 10.1029/2020WR028134

Fang Z, Carroll RWH, Schumer R, Harman C, Wilusz D, Williams KH. 2019. Streamflow partitioning and transit time distribution in snow-dominated basins as a function of climate. *Journal of Hydrology* **570**: 726–738 DOI: 10.1016/j.jhydrol.2019.01.029

Fontrodona Bach A, van der Schrier G, Melsen LA, Klein Tank AMG, Teuling AJ. 2018. Widespread and Accelerated Decrease of Observed Mean and Extreme Snow Depth Over Europe. *Geophysical Research Letters* **45** (22): 12,312–12,319 DOI: 10.1029/2018GL079799

Fovet O, Humbert G, Dupas R, Gascuel-Odoux C, Gruau G, Jaffrezic A, Thelusma G, Faucheux M, Gilliet N, Hamon Y, et al. 2018. Seasonal variability of stream water quality response to storm events captured using high-frequency and multi-parameter data. *Journal of Hydrology* **559**: 282–293 DOI: 10.1016/j.jhydrol.2018.02.040

von Freyberg J, Radny D, Gall HE, Schirmer M. 2014. Implications of hydrologic connectivity between hillslopes and riparian zones on streamflow composition. *Journal of Contaminant Hydrology* **169**: 62–74 DOI: 10.1016/j.jconhyd.2014.07.005

GEA, German Environment Agency (2017): Waters in Germany: Status and assessment. Dessau-Roßlau.

Godsey SE, Kirchner JW, Clow DW. 2009. Concentration–discharge relationships reflect chemostatic characteristics of US catchments. *Hydrological Processes* **23** (13): 1844–1864 DOI: 10.1002/hyp.7315

Gorski G, Zimmer M. 2021. Hydrologic regimes drive nitrate export behavior in human-impacted watersheds. *Hydrology and Earth System Sciences* DOI: 10.5194/HESS-25-1333-2021

Grayson RB, Western AW, Chiew FHS, Blöschl G. 1997. Preferred states in spatial soil moisture patterns: Local and nonlocal controls. *Water Resources Research* **33** (12): 2897–2908 DOI: 10.1029/97WR02174

Grinsven H, Berge H, Dalgaard T, Fraters B, Durand P, Hart A, Hofman G, Jacobsen B, Lalor S, Lesschen JP, et al. 2012. Management, regulation and environmental impacts of nitrogen fertilization in northwestern Europe under the Nitrates Directive; a benchmark study. *Biogeosciences* **9** DOI: 10.5194/bg-9-5143-2012

Guillemot S, Fovet O, Gascuel-Odoux C, Gruau G, Casquin A, Curie F, Minaudo C, Strohmenger L, Moatar F. 2021. Spatio-temporal controls of C–N–P dynamics across headwater catchments of a temperate agricultural region from public data analysis. *Hydrology and Earth System Sciences* **25**: 2491–2511 DOI: 10.5194/hess-25-2491-2021

Guntiñas ME, Leirós MC, Trasar-Cepeda C, Gil-Sotres F. 2012. Effects of moisture and temperature on net soil nitrogen mineralization: A laboratory study. *European Journal of Soil Biology* **48**: 73–80 DOI: 10.1016/j.ejsobi.2011.07.015

Hardie MA, Cotching WE, Doyle RB, Holz G, Lisson S, Mattern K. 2011. Effect of antecedent soil moisture on preferential flow in a texture-contrast soil. *Journal of Hydrology* **398** (3): 191–201 DOI: 10.1016/j.jhydrol.2010.12.008

Häussermann U, Klement L, Breuer L, Ullrich A, Wechsung G, Bach M. 2020. Nitrogen soil surface budgets for districts in Germany 1995 to 2017. *Environmental Sciences Europe* **32** DOI: 10.1186/s12302-020-00382-x

Heathwaite AL, Bieroza M. 2021. Fingerprinting hydrological and biogeochemical drivers of freshwater quality. *Hydrological Processes* **35** (1): e13973 DOI: 10.1002/hyp.13973

House WA, Leach DV, Armitage PD. 2001. Study of dissolved silicon, and nitrate dynamics in a fresh water stream. *Water Research* **35** (11): 2749–2757 DOI: 10.1016/S0043-1354(00)00548-0

Inamdar SP, O’Leary N, Mitchell MJ, Riley JT. 2006. The impact of storm events on solute exports from a glaciated forested watershed in western New York, USA. *Hydrological Processes* **20** (16): 3423–3439 DOI: 10.1002/hyp.6141

IPCC, 2021: Climate Change 2021: The Physical Science Basis. Contribution of Working Group I to the Sixth Assessment Report of the Intergovernmental Panel on Climate Change [Masson-Delmotte, V., P. Zhai, A. Pirani, S.L. Connors, C. Péan, S. Berger, N. Caud, Y. Chen, L. Goldfarb, M.I. Gomis, M. Huang, K. Leitzell, E. Lonnoy, J.B.R. Matthews, T.K. Maycock, T. Waterfield, O. Yelekçi, R. Yu, and B. Zhou (eds.)]. Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, In press, doi:10.1017/9781009157896.

Jencso KG, McGlynn BL, Gooseff MN, Wondzell SM, Bencala KE, Marshall LA. 2009. Hydrologic connectivity between landscapes and streams: Transferring reach- and plot-scale understanding to the catchment scale. *Water Resources Research* **45** (4) DOI: 10.1029/2008WR007225

Johannsen A, Dähnke K, Emeis K. 2008. Isotopic composition of nitrate in five German rivers discharging into the North Sea. *Organic Geochemistry* **39** (12): 1678–1689 DOI: 10.1016/j.orggeochem.2008.03.004

Knapp JLA, von Freyberg J, Studer B, Kiewiet L, Kirchner JW. 2020. Concentration–discharge relationships vary among hydrological events, reflecting differences in event characteristics. *Hydrology and Earth System Sciences* **24** (5): 2561–2576 DOI: 10.5194/hess-24-2561-2020

Knoll L, Breuer L, Bach M. 2020. Nation-wide estimation of groundwater redox conditions and nitrate concentrations through machine learning. *Environmental Research Letters* **15** (6): 064004 DOI: 10.1088/1748-9326/ab7d5c

Koenig LE, Shattuck MD, Snyder LE, Potter JD, McDowell WH. 2017. Deconstructing the Effects of Flow on DOC, Nitrate, and Major Ion Interactions Using a High-Frequency Aquatic Sensor Network. *Water Resources Research* **53** (12): 10655–10673 DOI: 10.1002/2017WR020739

Korom SF, Schuh WM, Tesfay T, Spencer EJ. 2012. Aquifer denitrification and in situ mesocosms: Modeling electron donor contributions and measuring rates. *Journal of Hydrology* **432–433**: 112–126 DOI: 10.1016/j.jhydrol.2012.02.023

Kruskal WH, Wallis WA. 1952. Use of Ranks in One-Criterion Variance Analysis. *Journal of the American Statistical Association* **47** (260): 583–621 DOI: 10.1080/01621459.1952.10483441

Kumar R, Livneh B, Samaniego L. 2013. Toward computationally efficient large-scale hydrologic predictions with a multiscale regionalization scheme. *Water Resources Research* **49** (9): 5700–5714 DOI: 10.1002/wrcr.20431

Kunkel R, Bach M, Behrendt H, Wendland F. 2004. Groundwater-borne nitrate intakes into surface waters in Germany. *Water Science and Technology* **49** (3): 11–19 DOI: 10.2166/wst.2004.0152

Lassaletta L, Billen G, Grizzetti B, Anglade J, Garnier J. 2014. 50 year trends in nitrogen use efficiency of world cropping systems: the relationship between yield and nitrogen input to cropland. *9* (10): 105011 DOI: 10.1088/1748-9326/9/10/105011

Lloyd CEM, Freer JE, Johnes PJ, Collins AL. 2016. Using hysteresis analysis of high-resolution water quality monitoring data, including uncertainty, to infer controls on nutrient and sediment transfer in catchments. *Science of The Total Environment* **543**: 388–404 DOI: 10.1016/j.scitotenv.2015.11.028

Lutz SR, Trauth N, Musolff A, Van Breukelen BM, Knöller K, Fleckenstein JH. 2020. How Important is Denitrification in Riparian Zones? Combining End-Member Mixing and Isotope Modeling to Quantify Nitrate Removal from Riparian Groundwater. *Water Resources Research* **56** (1) DOI: 10.1029/2019WR025528

Martin C, Aquilina L, Gascuel-Odoux C, Molénat J, Faucheu M, Ruiz L. 2004. Seasonal and interannual variations of nitrate and chloride in stream waters related to spatial and temporal patterns of groundwater concentrations in agricultural catchments. *Hydrological Processes* **18** (7): 1237–1254 DOI: 10.1002/hyp.1395

Meter KJV, Basu NB. 2017. Time lags in watershed-scale nutrient transport: an exploration of dominant controls.

Environmental Research Letters **12** (8): 084017 DOI: 10.1088/1748-9326/aa7bf4

Meter KJV, Basu NB, Veenstra JJ, Burras CL. 2016. The nitrogen legacy: emerging evidence of nitrogen accumulation in anthropogenic landscapes. *Environmental Research Letters* **11** (3): 035014 DOI: 10.1088/1748-9326/11/3/035014

Meybeck M, Moatar F. 2012. Daily variability of river concentrations and fluxes: indicators based on the segmentation of the rating curve. *Hydrological Processes* **26** (8): 1188–1207 DOI: 10.1002/hyp.8211

Minaudo C, Dupas R, Gascuel-Odoux C, Roubeix V, Danis P-A, Moatar F. 2019. Seasonal and event-based concentration-discharge relationships to identify catchment controls on nutrient export regimes. *Advances in Water Resources* **131**: 103379 DOI: 10.1016/j.advwatres.2019.103379

Moatar F, Abbott BW, Minaudo C, Curie F, Pinay G. 2017. Elemental properties, hydrology, and biology interact to shape concentration-discharge curves for carbon, nutrients, sediment, and major ions. *Water Resources Research* **53** (2): 1270–1287 DOI: 10.1002/2016WR019635

Mulholland PJ, Helton AM, Poole GC, Hall RO, Hamilton SK, Peterson BJ, Tank JL, Ashkenas LR, Cooper LW, Dahm CN, et al. 2008. Stream denitrification across biomes and its response to anthropogenic nitrate loading. *Nature* **452** (7184): 202–205 DOI: 10.1038/nature06686

Musolff A. 2020. WQQDB - water quality and quantity data base Germany: metadata DOI: 10.4211/hs.a42addcbd59a466a9aa56472dfef8721

Musolff A, Fleckenstein J, Opitz M, Büttner O, Kumar R, Tittel J. 2018. Spatio-temporal controls of dissolved organic carbon stream water concentrations. *Journal of Hydrology* **566** DOI: 10.1016/j.jhydrol.2018.09.011

Musolff A, Fleckenstein JH, Rao PSC, Jawitz JW. 2017. Emergent archetype patterns of coupled hydrologic and biogeochemical responses in catchments: Emergence of Archetype C-Q Patterns. *Geophysical Research Letters* **44** (9): 4143–4151 DOI: 10.1002/2017GL072630

Musolff A, Zhan Q, Dupas R, Minaudo C, Fleckenstein J, Rode M, Dehaspe J, Rinke K. 2021. Spatial and Temporal Variability in Concentration-Discharge Relationships at the Event Scale. *Water Resources Research* **57** DOI: 10.1029/2020WR029442

Ocampo CJ, Sivapalan M, Oldham C. 2006. Hydrological connectivity of upland-riparian zones in agricultural catchments: Implications for runoff generation and nitrate transport. *Journal of Hydrology* **331** (3): 643–658 DOI: 10.1016/j.jhydrol.2006.06.010

Ortmeyer F, Begerow D, Guerreiro MA, Wohnlich S, Banning A. 2021. Comparison of Denitrification Induced by Various Organic Substances—Reaction Rates, Microbiology, and Temperature Effect. *Water Resources Research* **57** (11) DOI: 10.1029/2021WR029793

Outram FN, Cooper RJ, Sünnenberg G, Hiscock KM, Lovett AA. 2016. Antecedent conditions, hydrological connectivity and anthropogenic inputs: Factors affecting nitrate and phosphorus transfers to agricultural headwater streams. *Science of The Total Environment* **545–546**: 184–199 DOI: 10.1016/j.scitotenv.2015.12.025

Paerl HW. 1997. Coastal eutrophication and harmful algal blooms: Importance of atmospheric deposition and groundwater as “new” nitrogen and other nutrient sources. *Limnology and Oceanography* **42** (5part2): 1154–1165 DOI: 10.4319/lo.1997.42.5_part_2.1154

Puckett LJ, Tesoriero AJ, Dubrovsky NM. 2011. Nitrogen Contamination of Surficial Aquifers—A Growing Legacy. *Environmental Science & Technology* **45** (3): 839–844 DOI: 10.1021/es1038358

Rauthe M, Steiner H, Riediger U, Mazurkiewicz A, Gratzki A. 2013. A Central European precipitation climatology – Part I: Generation and validation of a high-resolution gridded daily data set (HYRAS). *Meteorologische Zeitschrift*: 235–256 DOI: 10.1127/0941-2948/2013/0436

Rode M, Halbedel née Angelstein S, Anis MR, Borchardt D, Weitere M. 2016. Continuous In-Stream Assimilatory Nitrate Uptake from High-Frequency Sensor Measurements. *Environmental Science & Technology* **50** (11): 5685–5694 DOI: 10.1021/acs.est.6b00943

Rose LA, Karwan DL, Godsey SE. 2018. Concentration–discharge relationships describe solute and sediment mobilization, reaction, and transport at event and longer timescales. *Hydrological Processes* **32** (18): 2829–2844 DOI: 10.1002/hyp.13235

Sabater S, Butturini A, Clement J-C, Burt T, Dowrick D, Hefting M, Matre V, Pinay G, Postolache C, Rzepecki M, et al. 2003. Nitrogen Removal by Riparian Buffers along a European Climatic Gradient: Patterns and Factors of Variation. *Ecosystems* **6** (1): 0020–0030 DOI: 10.1007/s10021-002-0183-8

Samaniego L, Kumar R, Attinger S. 2010. Multiscale parameter regionalization of a grid-based hydrologic model at the mesoscale. *Water Resources Research* **46** (5) DOI: 10.1029/2008WR007327

Schwientek M, Osenbrück K, Fleischer M. 2013. Investigating hydrological drivers of nitrate export dynamics in two agricultural catchments in Germany using high-frequency data series. *Environmental Earth Sciences* **69** (2): 381–393 DOI: 10.1007/s12665-013-2322-2

Seibert J, Grabs T, Köhler S, Laudon H, Winterdahl M, Bishop K. 2009. Linking soil- and stream-water chemistry based on a Riparian Flow-Concentration Integration Model. *Hydrology and Earth System Sciences* **13** (12): 2287–2297 DOI: 10.5194/hess-13-2287-2009

Stieglitz M, Shaman J, McNamara J, Engel V, Shanley J, Kling GW. 2003. An approach to understanding hydrologic connectivity on the hillslope and the implications for nutrient transport. *Global Biogeochemical Cycles* **17** (4) DOI: 10.1029/2003GB002041

Stumpf RP, Johnson LT, Wynne TT, Baker DB. 2016. Forecasting annual cyanobacterial bloom biomass to inform management decisions in Lake Erie. *Journal of Great Lakes Research* **42** (6): 1174–1183 DOI: 10.1016/j.jglr.2016.08.006

Tarasova L, Basso S, Wendi D, Viglione A, Kumar R, Merz R. 2020. A Process-Based Framework to Characterize and Classify Runoff Events: The Event Typology of Germany. *Water Resources Research* **56** (5): e2019WR026951 DOI: 10.1029/2019WR026951

Tarasova L, Basso S, Zink M, Merz R. 2018. Exploring Controls on Rainfall-Runoff Events: 1. Time Series-Based Event Separation and Temporal Dynamics of Event Runoff Response in Germany. *Water Resources Research* **54** (10): 7711–7732 DOI: 10.1029/2018WR022587

Taszarek M, Kendzierski S, Pilgij N. 2020. Hazardous weather affecting European airports: Climatological estimates of situations with limited visibility, thunderstorm, low-level wind shear and snowfall from ERA5. *Weather and Climate Extremes* **28**: 100243 DOI: 10.1016/j.wace.2020.100243

Tesoriero AJ, Duff JH, Saad DA, Spahr NE, Wolock DM. 2013. Vulnerability of Streams to Legacy Nitrate Sources. *Environmental Science & Technology* **47** (8): 3623–3629 DOI: 10.1021/es305026x

Thompson SE, Basu NB, Lascurain Jr. J, Aubeneau A, Rao PSC. 2011. Relative dominance of hydrologic versus biogeochemical factors on solute export across impact gradients. *Water Resources Research* **47** (10) DOI: 10.1029/2010WR009605

Vaughan MCH, Bowden WB, Shanley JB, Vermilyea A, Sleeper R, Gold AJ, Pradhanang SM, Inamdar SP, Levia DF, Andres AS, et al. 2017. High-frequency dissolved organic carbon and nitrate measurements reveal differences in storm hysteresis and loading in relation to land cover and seasonality. *Water Resources Research* **53** (7): 5345–5363 DOI: 10.1002/2017WR020491

Veith TL, Preisendanz HE, Elkin KR. 2020. Characterizing transport of natural and anthropogenic constituents in a long-term agricultural watershed in the northeastern United States. *Journal of Soil and Water Conservation* **75** (3): 319–329 DOI: 10.2489/jswc.75.3.319

Vervloet LSC, Binning PJ, Børgesen CD, Højberg AL. 2018. Delay in catchment nitrogen load to streams following restrictions on fertilizer application. *Science of The Total Environment* **627**: 1154–1166 DOI: 10.1016/j.scitotenv.2018.01.255

Weber G, Honecker U, Kubiniok J. 2020. Nitrate dynamics in springs and headwater streams with agricultural catchments in southwestern Germany. *Science of The Total Environment* **722**: 137858 DOI: 10.1016/j.scitotenv.2020.137858

Weitere M, Altenburger R, Anlanger C, Baborowski M, Bärlund I, Beckers L-M, Borchardt D, Brack W, Bräse L, Busch W, et al. 2021. Disentangling multiple chemical and non-chemical stressors in a lotic ecosystem using a longitudinal approach. *Science of The Total Environment* **769**: 144324 DOI: 10.1016/j.scitotenv.2020.144324

Wendland F, Blum A, Coetsiers M, Gorova R, Griffioen J, Grima J, Hinsby K, Kunkel R, Marandi A, Melo T, et al. 2008. European aquifer typology: a practical framework for an overview of major groundwater composition at European scale. *Environmental Geology* **55** (1): 77–85 DOI: 10.1007/s00254-007-0966-5

Winter C, Lutz SR, Musolff A, Kumar R, Weber M, Fleckenstein JH. 2021. Disentangling the Impact of Catchment Heterogeneity on Nitrate Export Dynamics From Event to Long-Term Time Scales. *Water Resources Research* **57** (1): e2020WR027992 DOI: 10.1029/2020WR027992

Winter C, Tarasova L, Lutz S, Musolff A, Kumar R, Fleckenstein J. 2022. *Explaining the Variability in High-Frequency Nitrate Export Patterns Using Long-Term Hydrological Event Classification*. DOI: 10.1002/essoar.10507676.1

Yang J, Heidbüchel I, Musolff A, Reinstorf F, Fleckenstein JH. 2018. Exploring the Dynamics of Transit Times and Subsurface Mixing in a Small Agricultural Catchment. *Water Resources Research* **54** (3): 2317–2335 DOI: 10.1002/2017WR021896

Zhang X, Yang X, Jomaa S, Rode M. 2020. Analyzing impacts of seasonality and landscape gradient on event-scale nitrate-discharge dynamics based on nested high-frequency monitoring. *Journal of Hydrology* **591**: 125585 DOI: 10.1016/j.jhydrol.2020.125585

Zhi W, Li L, Dong W, Brown W, Kaye J, Steefel C, Williams KH. 2019. Distinct Source Water Chemistry Shapes Contrasting Concentration-Discharge Patterns. *Water Resources Research* **55** (5): 4233–4251 DOI: 10.1029/2018WR024257

Zink M, Kumar R, Cuntz M, Samaniego L. 2017. A high-resolution dataset of water fluxes and states for Germany accounting for parametric uncertainty.

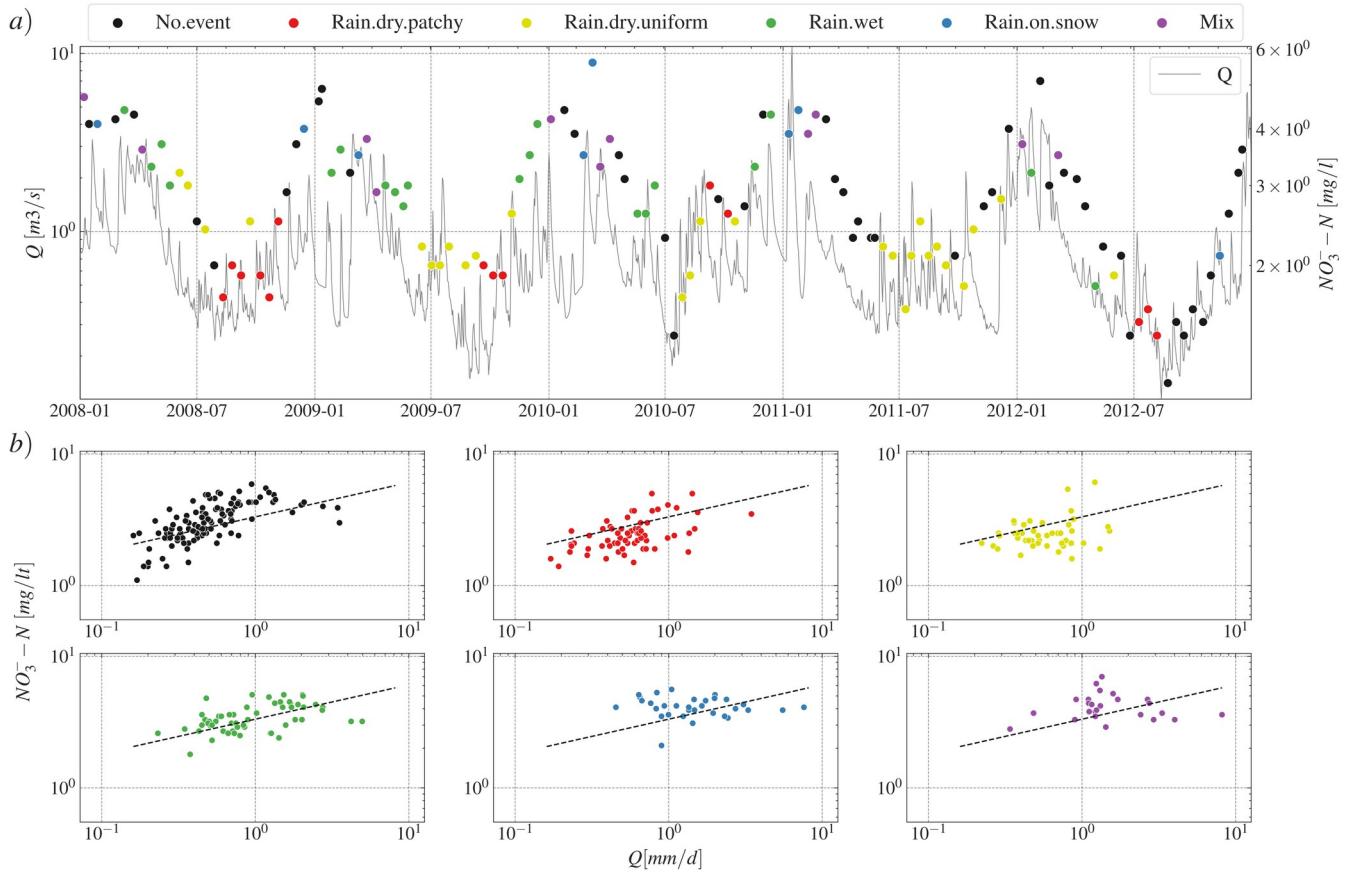
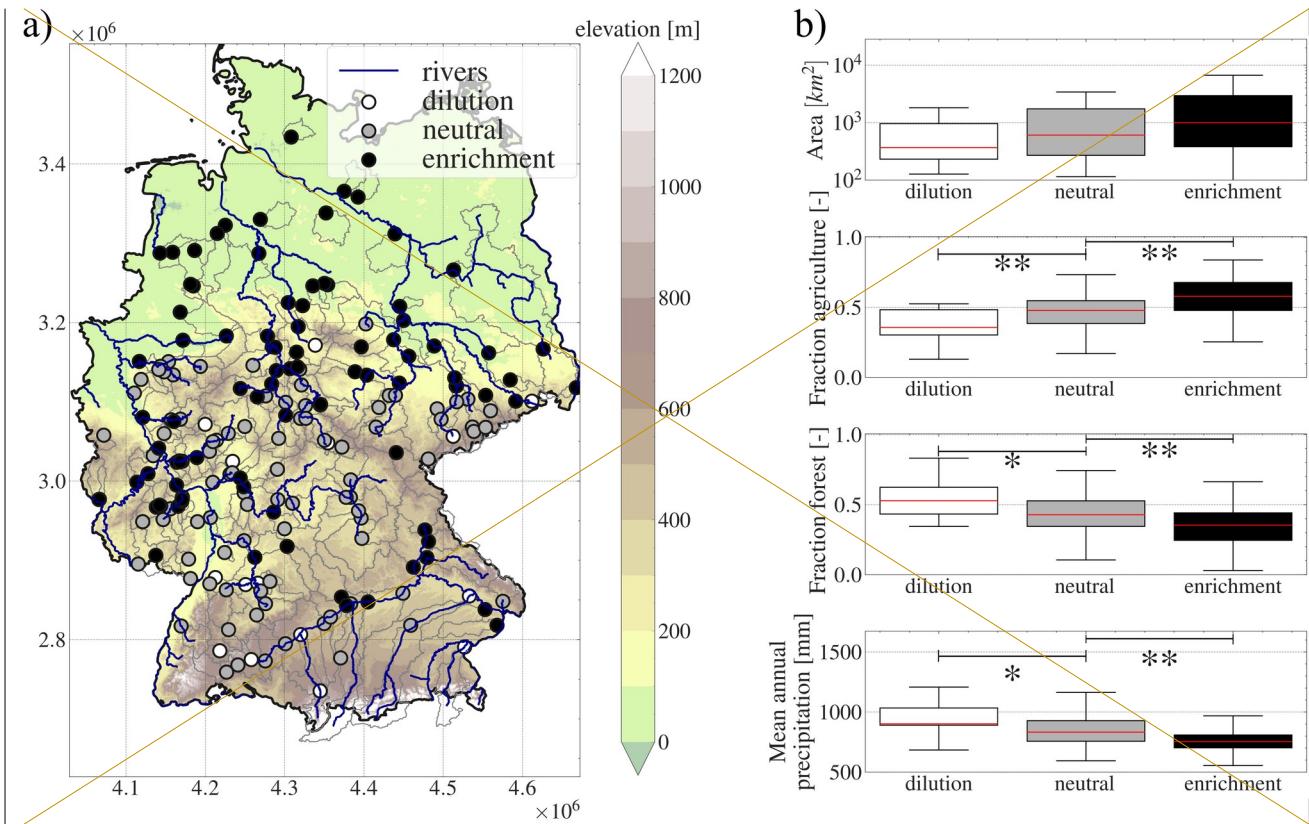


Figure 1: a) Time series of daily discharge and bi-weekly grab sample nitrate concentration during event and no event conditions in the Naab River at the gauge of Unterkoebitz, Bavaria over a period of five years. Event types are differentiated by colors (see Figure 3 for details). b) Double logarithmic plot of C-Q pairs for samples (from 2000 to 2012) taken during different event types and no event conditions. Black dashed lines show the long-term C-Q relationship (same line in each subplot) obtained from linear regression in a double logarithmic plot of C-Q values for all available samples.

685

690



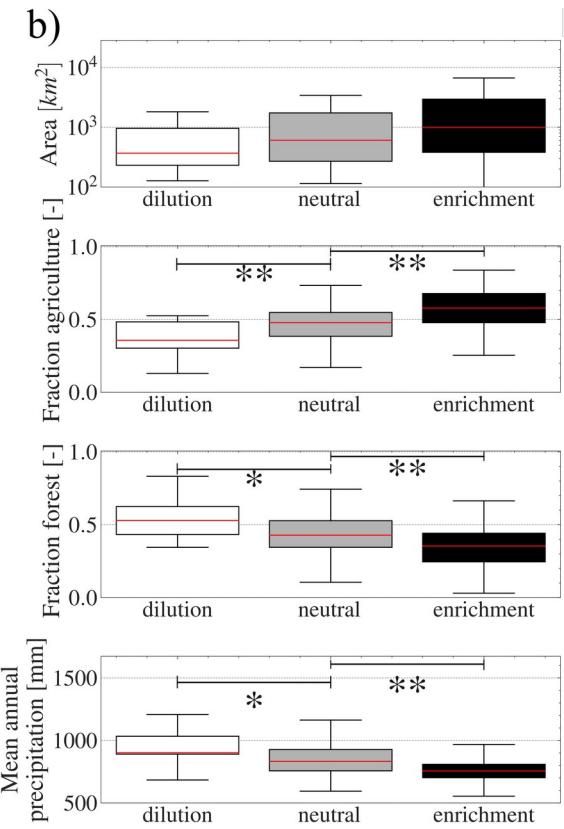
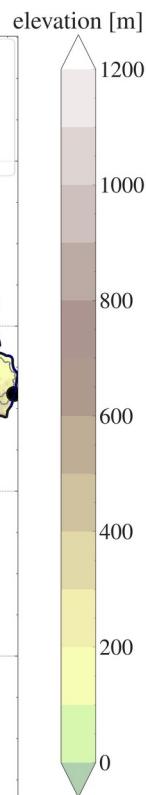
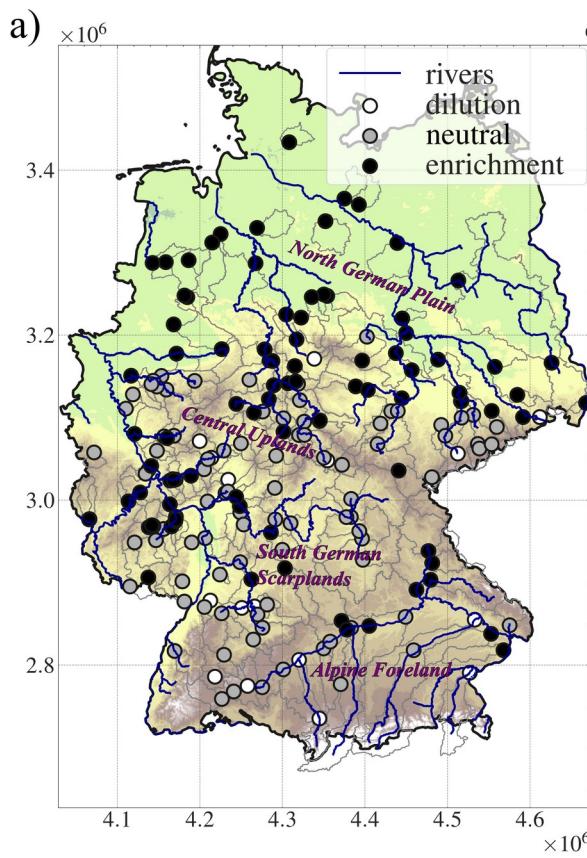


Figure 2: a) Study area and stations of nitrate concentration measurements in stream water. Gray lines show catchment boundaries. Catchment outlets (points) are color-coded according to the long-term export pattern (dilution, neutral and enrichment). Blue lines show the main rivers. The background colormap corresponds to the elevation. Purple labels indicate German natural regions b) Area, fraction of agriculture, fraction of forest and mean annual precipitation of study catchments grouped according to export patterns (dilution, neutral and enrichment). Red lines show medians of boxplots and significance of median differences between adjacent boxplots was estimated using Kruskal-Wallis test (displayed as * for $p<0.05$ and ** for $p<0.01$).

695

700

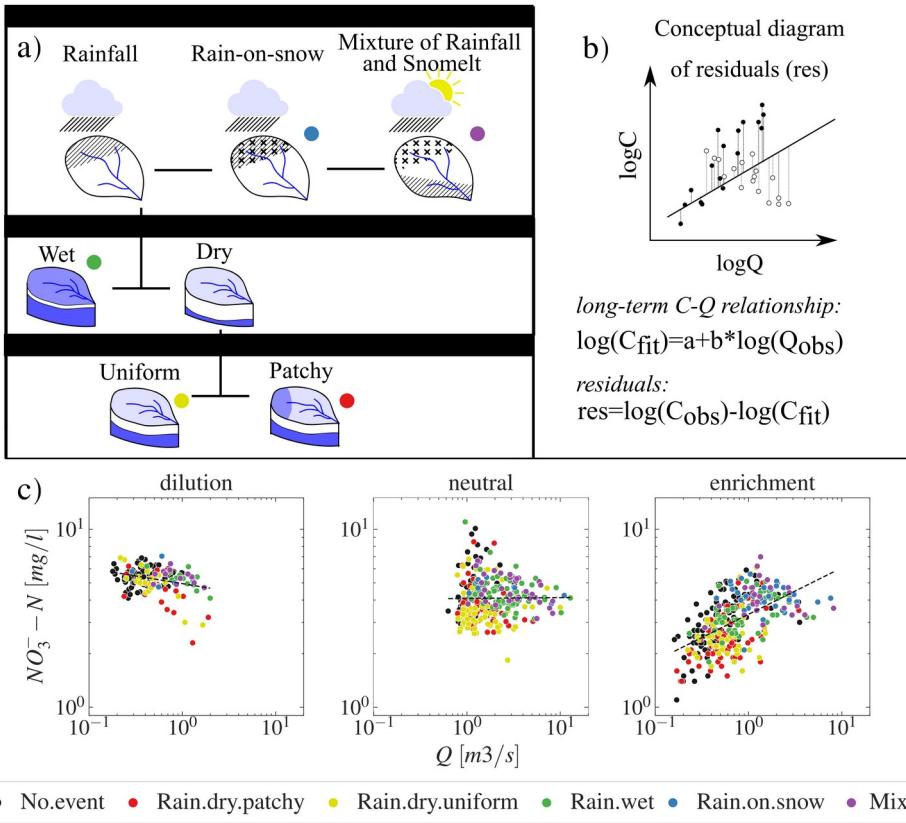


Figure 3: a) Hierarchical scheme for event classification (modified from Tarasova et al., (2020); classification criteria are provided in Table S1). Colored dots located next to the five different event types indicate their markers. b) Cobs and Qobs are observed concentration and discharge, Cfit is the nitrate concentration estimated from fitting the long-term C-Q relationship with a linear relation in log-log space, and res is the residual value. c) C-Q plots for three different catchments attributed to different long-term nitrate export patterns based on the $\log C - \log Q$ slope b, i.e., dilution ($b < 0.1$, the Würm River in Pforzheim), neutral ($b \sim 0$, the Wupper River in Opladen) and enrichment ($b > 0.1$, the Naab River in Unterkoeblitz).

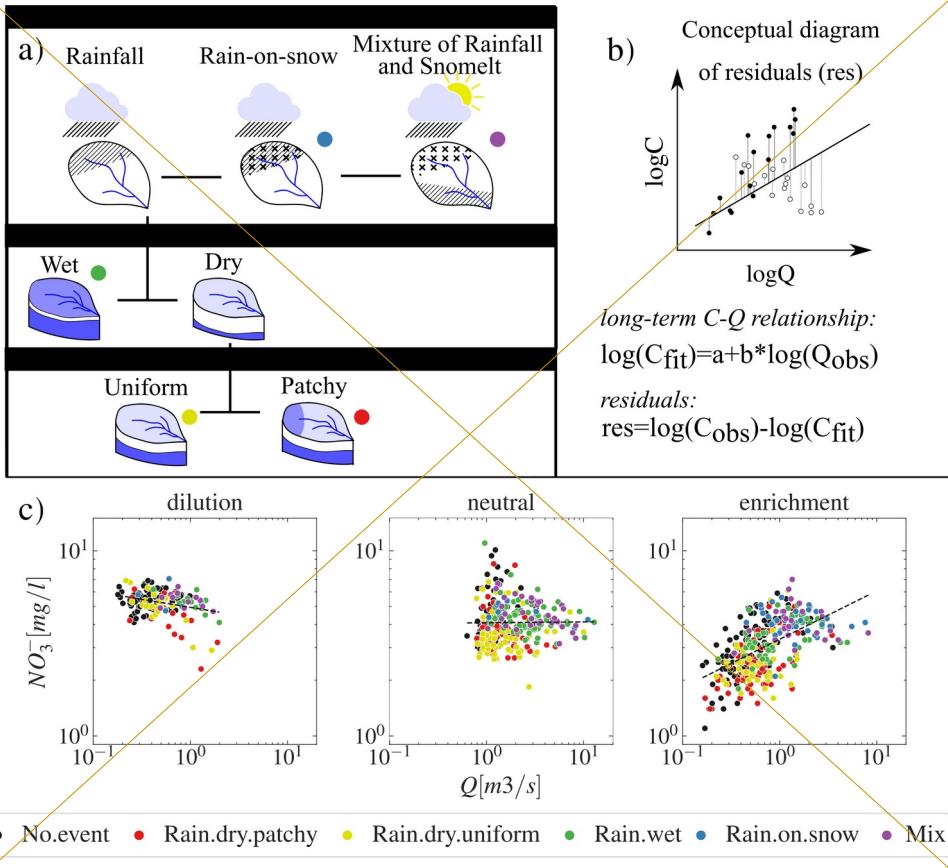
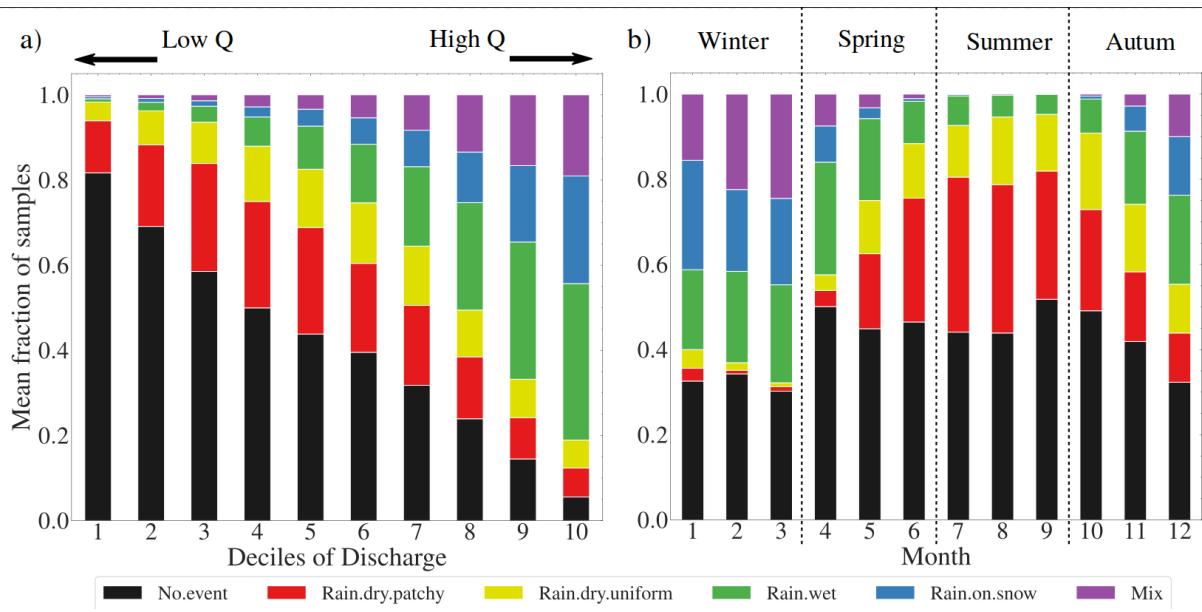
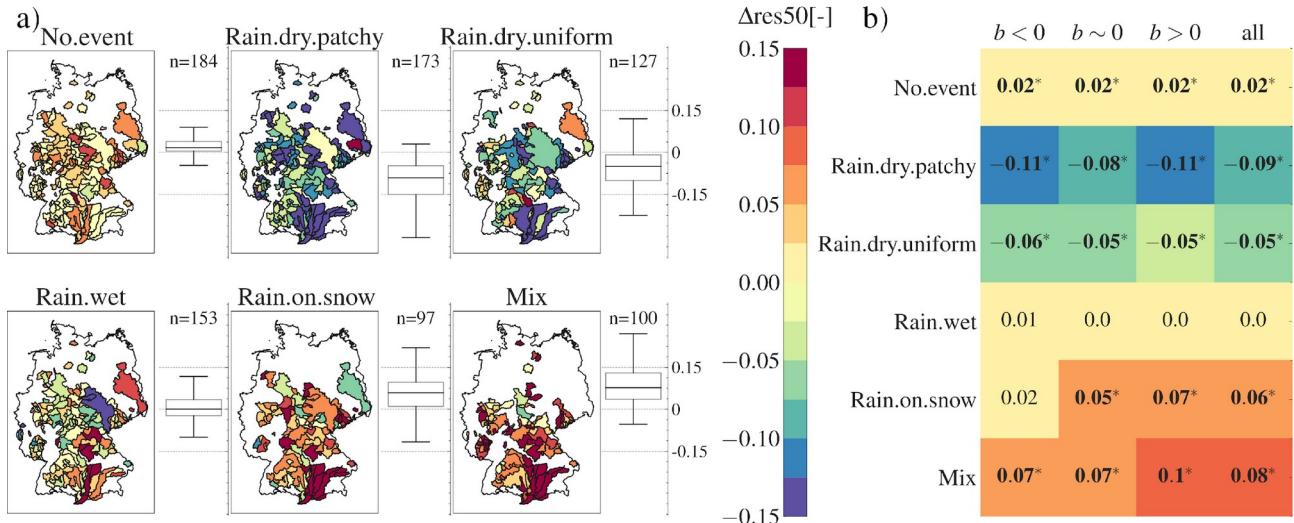


Figure 4: a) Hierarchical scheme for event classification (modified from Tarasova et al., (2020); classification criteria are provided in Table S1). Colored dots located next to the five different event types indicate their markers. **b)** C_{obs} and Q_{obs} are observed concentration and discharge, C_{fit} is the nitrate concentration estimated from fitting the long-term C-Q relationship with a linear relation in log-log space, and res is the residual value. **c)** C-Q plots for three different catchments attributed to different long-term nitrate export patterns based on the $\log C / \log Q$ slope b , i.e., dilution ($b < 0.1$, the Würm River in Pforzheim), neutral ($b = 0$, the Wupper River in Opladen) and enrichment ($b > 0.1$, the Naab River in Unterkoeblitz).



715

Figure 5: a) Mean fraction of samples linked to each event type according to each catchment decile of discharge, and b) seasonal distribution of mean fraction of samples linked to each event type in the study catchments.



720

Figure 6: Median deviations of nitrate concentrations from the long-term C-Q relationships (Δres50) a) Δres50 values of different event types for each catchment. On the right-hand side of each map, boxplots show the distribution of Δres50 values across catchments for each event type (box limits represent the interquartile range and whiskers correspond to the 5th and 95th percentiles). b) Heatmap of Δres50 values averaged across different groups of catchments, considering all nitrate data for each event type and No.event. The three first columns of the heatmap correspond to one of the long-term export patterns (i.e., dilution (slope $b < 0$), neutral (slope $b \sim 0$) and enrichment (slope $b > 0$)) and the fourth column corresponds to all study catchments. Bold font and * indicates significant differences (Kruskal-Wallis test, $p < 0.05$) between median deviations across catchments for each event type and median deviation across catchments of all nitrate samples.

725

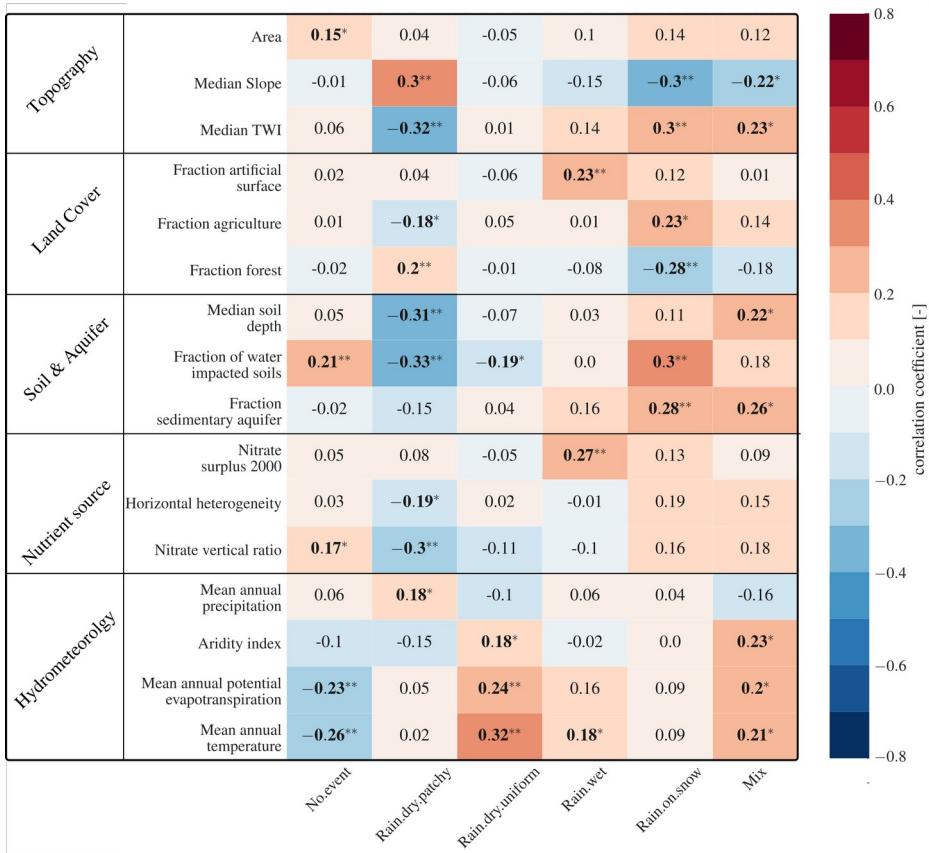


Figure 7: Spearman rank correlation coefficient between deviations of nitrate concentrations from the long-term C-Q relationships (Δ_{res50}) of a particular event type across study catchments and catchment descriptors. Significant correlations are indicated by bold font and * for $p < 0.05$ and ** for $p < 0.01$.

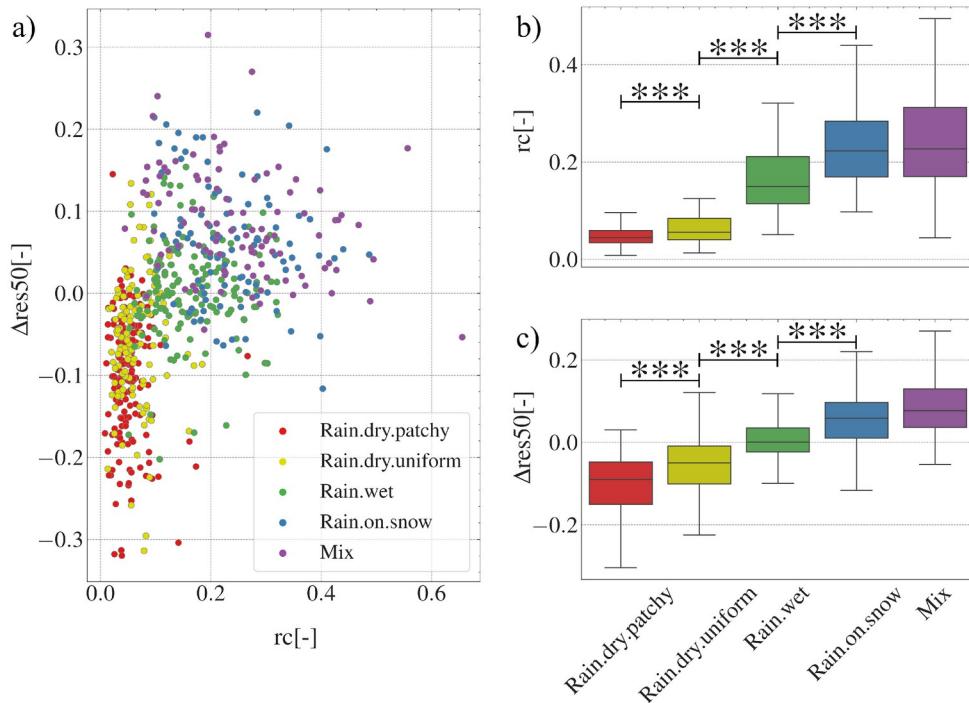


Figure 8: (a) Relationship between Δres50 for each catchment and event type and median runoff coefficient (rc); runoff coefficient is not defined for No.event. **(b)** Variability of runoff coefficients (rc) for each event type and **(c)** median residuals for each event type. Significance of median differences between adjacent boxplots was estimated using Kruskal-Wallis test (displayed as *** for $p<0.001$).