Particle size distribution and PM concentrations during synoptic and convective dust events in West Texas

Karin Ardon-Dryer[1], Mary C. Kelley[1]

[1]Department of Geosciences, Atmospheric Science Group, Texas Tech University, Lubbock, TX, USA

5   *Correspondence to*: Karin Ardon-Dryer (Karin.ardon-dryer@ttu.edu)

Abstract. Dust events are an important and complex constituent of the atmospheric system that can impact Earth's climate, the environment, and human health. The frequency of dust events in West Texas has increased over the past two decades, yet their impact on air quality in this region is still unclear as there is only one air quality monitoring station that measures only $PM_{2.5}$ concentrations (Particulate matter with aerodynamic diameter < 2.5 μm), and there is no information on other PM sizes or particle size distribution. The Aerosol Research Observation Station (AEROS) unit provides insight into the local variation of particle concentration during different dust events and allows for a better understanding of the impact of dust events on air quality. Since the west Texas area is prone to dust events, we were wondering if dust events generated by different meteorological causes (synoptic vs convective) will present similar particle concentrations or particle size distributions. In this project, three different dust events were measured by AEROS and compared. Each dust event originated from a different direction and lasted a different duration. One of the dust events was synoptic (April 10, 2019) and two were convective (June 5 and 21, 2019). Measurements of particle mass and number concentration, size distribution, and meteorological conditions for each dust event were compared. The Synoptic dust event (of April 10) was longer (12h) and had stronger wind speed conditions (up to 22.1 m sec^{-1}), while the two convective dust events lasted only 20 and 30 minutes and had lower wind speeds (up to 16.5 and 13.4 m sec^{-1} for June 5 and 21, respectively). Observation of PM based on daily and hourly values showed an impact on air quality, yet measurements based on daily and hourly values underestimate the impact of the convective dust events. Observations based on a shorter time scale (10-minutes) reveal the true impact of the two convective dust events. A comparison of particle size distribution showed that all three dust events had an increase of particles in the size range of 0.3 to 10 μm. Comparison of particle concentration for particles >5 μm and >10 μm show very high particle concentrations during the dust events. Some particle sizes even increase the concentration by ~2 orders of magnitude compared to the time before the dust event. Leading us to speculate that the impact on air quality of convective dust events in this region is underestimated with the current (hourly basis) method.

1 Introduction

Atmospheric dust particles, generated during dust events, are the second-largest contributor to the global aerosol burden (Goudie and Middleton, 2006; Textor et al., 2006). Dust events are common in arid and semi-arid environments (Goudie and Middleton, 2006; Goudie, 2014), occur when strong winds pick up loose dust particles, making them suspended in the

atmosphere (Goudie, 2014; Middleton, 2017). Dust events can be generated by two main meteorological disturbances, synoptic and convective. Synoptic is an upper-level disturbance including warm and cold fronts, low and high-pressure systems, troughs, and ridges, while convective caused by thunderstorms, including thunderstorm outflow boundaries and thunderstorm downbursts (Knippertz, 2014).

35

Among the various regions common to dust events including Africa, Asia, and Australia,. ~~Dust~~ dust events in the US contribute only 5% of the global dust emissions (Miller et al., 2004), where most of the dust events formed in the western portion of the US (Goudie, 2014; Rublee et al., 2020). Studies have shown that dust events in the central and ~~southeastern~~south-eastern US have increased in the last decade (Hand et al., 2016; Tong et al., 2017; Kelley and Ardon-Dryer, 2021), and climate models

40 predict that these will increase even more with climate change (Pu and Ginoux, 2017; Achakulwisut et al., 2018; Brey et al., 2020).

Dust events are an important and complex constituent of the atmospheric system as they can impact the Earth's climate, the environment, and human wellbeing in different ways. Atmospheric dust particles affect climate directly through scattering and

45 absorption of solar radiation (Wang et al., 2009; Lau et al., 2020) and indirectly by acting as cloud condensation nuclei and ice nuclei particles (Chen et al., 2019; Ardon-Dryer and Levin, 2014). Dust particles can influence the atmospheric vertical electric ~~filed~~ field (Ardon-Dryer et al., 2021), impact the economy (Tozer and Leys, 2013; Al-Hemoud et al., 2019; Abdullaev and Sokolik, 2020), as well as human well-being and health (Goudie, 2014; Bhattachan et al., 2019; Ardon-Dryer et al., 2020).

50 During dust events, the particle concentration ~~concertation~~ may exceed the health-recommended daily PM_{10} and $PM_{2.5}$ (particulate matter with an aerodynamic diameter <10 μm and <2.5 μm, respectively) threshold values by the World Health Organization (WHO; 50 and 25 µg m^{-3}, respectively; WHO, 2006, or 45 and 15 µg m^{-3}, respectively; based on the updated air quality guidelines WHO, 2021) and US Environmental Protection Agency (150 and 35 µg m^{-3}, respectively; EPA, 2016). Information on atmospheric dust particle concentrations and their sizes during dust events are important, as dust events and

55 the high particle concentration ~~concertation~~ have significant public health impacts (Aghababaeian et al., 2021). Epidemiological studies have demonstrated that there is a direct link between exposure to high amounts of dust particles and the number of daily hospitalizations and death cases (Karanasiou et al. 2012; Rublee et al., 2020; Herrera-Molina et al., 2021). Exposure to dust particles during dust events can cause respiratory and cardiovascular problems (Zhang et al., 2016; Toure et al., 2019; Rublee et al., 2020), increase the probability of low birth weight, and premature birth (Dastoorpoor et al., 2018;

60 Jones, 2020; Bogan et al., 2021), cause different diseases such as meningitis (Diokhane et al., 2016), valley fever (Middleton, 2020), and in rare cases end in death (Crooks et al., 2016; Zhang et al., 2016). Information on the increase of particle concentrations, change of particle~~s~~ sizes, with the degradation of air quality during dust events can help understand the impact these events have on people who are exposed to them.

65   The air quality in the Southern High Plains of West Texas is good overall, with a daily PM$_{2.5}$ value of around 7.1 ± 7.5 µg m^{-3} (Kelley et al., 2020). This area experiences many dust events, with an annual average of ~21 dust events per year, mainly in the spring and early summer (Kelley and Ardon-Dryer, 2021). Dust events in this region occur due to the strong wind speed, low surface cover, and low moisture conditions (Stout, 2001). Analysis of 420 dust events in this region by Kelley and Ardon-Dryer (2021) showed that most of the dust events are only one hour long and very few exceeded the regulatory recommended

70   PM$_{2.5}$ daily threshold. Out of these dust events, only synoptic dust events exceeded the daily threshold, but none of the convectively driven dust events exceeded the daily threshold. It is unclear whether convective events are less intense and have lower particles concentrations or whether the methods used to evaluate their impact (daily and hourly measurements) were not sensitive enough to detect these events, as many of the convective dust events in West Texas are of short duration.

75   To better understand the impact of dust events on air quality, as well as to examine if different types of dust events (synoptic vs. convective) have a similar particle mass and number concentration, additional measurements are needed. The Aerosol Research Observation Station (AEROS) was designed for that purpose. AEROS which has been operational since March 2019 allows for continuous monitoring of particle mass and number concentration including mass concentration of different PM sizes, and particle size distribution (Ardon-Dryer et al., 2022). Three dust events were captured by AEROS during the study

80   period, one synoptic and two convective. A comparison between these dust events based on particle number and mass concentrations and particle size distribution and their impact on air quality will be presented.

2 Method

2.1 Research area and measurement station Subsection (as Heading 2)

Measurements were conducted in Lubbock Texas which is located in the Southern High Plains of West Texas (33°35'12.5"N

85   101°52'31.3"W; Fig. 1). The flat urban area is a rural area surrounded by numerous agricultural fields in a semi-arid region, at approximately 1 km above sea level. AEROS is located on a Texas Tech University campus on a building rooftop at 9.8 m above the ground. The aerosol unit includes a shad shed that is temperature controlled by an air conditioning unit that keeps a continuous temperature of 22 °C. It has four rain-protected inlet units at 2.9 m from the rooftop floor (1 ± 0.01 m from the station rooftop), each inlet connected to a stainless-steel tube (0.013 m diameter - 1/2 inch tube) that is connected to a custom-

90   built in-line dryer unit, which is used to remove condensed-phase water from the collected particles. Swagelok reducer connects each dryer to an aerosol instrument (one for each instrument) using a 0.0064 m diameter (1/4 inch tube) stainless steel tube. The fourth inlet is used for aerosol collection using a filter holder. The three aerosol instruments include the TSI 3330 Optical Particle Sizer (OPS), a DustTrak DRX aerosol monitor (TSI 8533EP, Shoreview, MN, USA; TSI, 2021), and Grimm 11-D system Portable Aerosol Spectrometer (Grimm Aerosol Technik GmbH & Co. KG, Germany; Grimm 11-D, 2021).

95

The OPS measures total particle number concentration and particle size distributions in 16 channels from 0.3 to 10 µm, at a time resolution of 60 sec, using a flow rate of 1.0 L min^{-1}. The DustTrak DRX measures aerosol mass concentrations at various PM sizes (PM$_1$, PM$_{2.5}$, PM$_4$, and PM$_{10}$) at a time resolution of 60 sec, using a flow rate of 1.0 L min^{-1}. The Grimm 11-D measures total particle number concentrations, mass concentration (e.g., PM$_1$, PM$_{2.5}$, PM$_4$, and PM$_{10}$), and size distribution

100   over a size range of 0.25 – 35.15 µm in 31 channels (bins). Data are recorded every 60 sec at a flow rate of 1.2 L min^{-1}. Data from the three units were collected each minute and calculated using MATLAB for 10-minutes, hourly, and daily average values. Since the Grimm 11-D provide the concentration of particles for each bin size, calculations of size distributions for number (dN/dlogD$_p$) concentrations were performed from the instrument output using Matlab. All instrument time was synchronized and converted to local Central Standard Time (CST). Additional information on AEROS and each of the aerosol

105   instruments including an intercomparison analysis can be found in Ardon-Dryer et al. (2022).

2.2 Meteorological measurements

Meteorological information, such as 5-minute to hourly ambient temperature, relative humidity, wind speed, wind direction, wind gust, visibility, pressure, and precipitation were retrieved from the local National Weather Service (NWS) Automated Surface Observation System (ASOS), available via the METeorological Aerodrome Reports (METARs) which is located ~9.8

110   km North-East from AEROS (33° 39' 48.96" N. 101° 49' 22.8" W, Fig. 1). Observations of meteorological conditions (e.g., thunderstorms, rain, haze, and dust) were retrieved for that period using the "Present Weather Code". All times were converted to Central Standard Time.
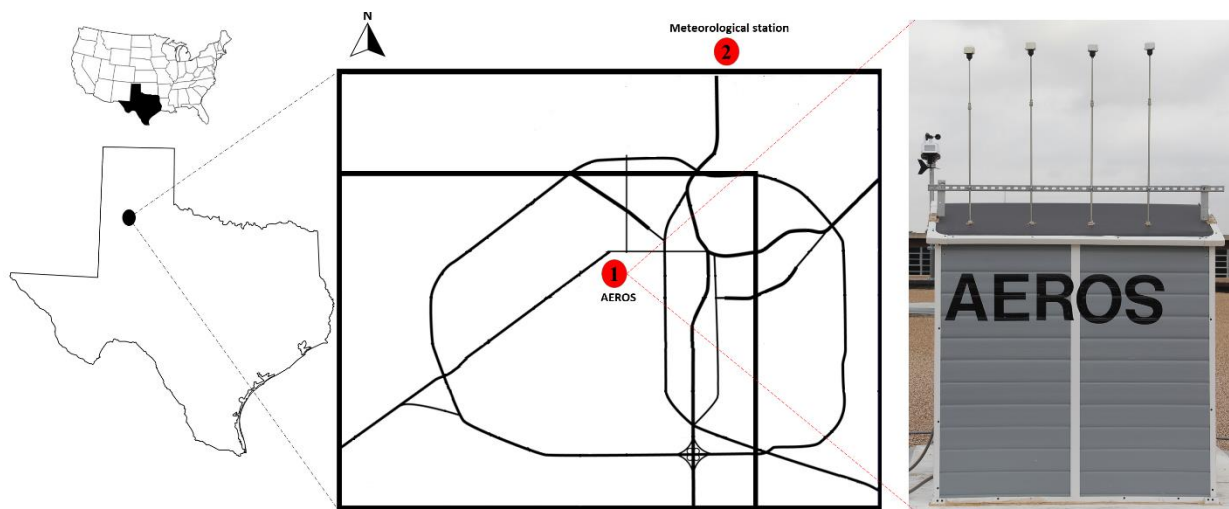
Figure 1. Location of AEROS (1) and meteorological station (2) in the South High Plains of West Texas. Photos show the AEROS aerosol
115   measurements unit.

3 Results and discussion

3.1. Different meteorological conditions initiate each dust event

Three different dust events were captured by AEROS and compared in this work. The dust events occurred on April 10, June 5, and June 21, 2019. All dust events were defined in the METARs as blowing dust events (BLDU). The April 10 dust event was caused by a synoptic disturbance, while the two June dust events were convective. A description of the meteorological conditions that initiate each dust event is described below. Each of these dust events originated from a different direction as shown in Fig. S1.

The synoptic event on April 10 was caused by a low-pressure system that was located between Colorado, New Mexico, and Texas. As the system moved east, strong winds were generated from the west and moved toward AEROS. Strong winds started around 11:00, but a reduction in visibility did not start until 13:30. The highest wind speed and wind gust (22.1 m sec^{-1} and 26.7 m sec^{-1}, respectively) were measured at ~17:50, while the lowest visibility reached 0.8 km minutes afterward (at 17:53). The cold front associated with the low-pressure system moved through the area in the late evening hours resulting in the continuation of strong winds until the early morning hours the following day (April 11). Visibility during this day behaved according to the wind speed, as can be seen in Fig. 2A. A movie showing the dust event taken from the 12th floor of the Atmospheric science group in the Media and Communications building at Texas Tech University campus presents a continuous flow of dust particles (Movie S1A).

The second dust event that occurred on June 5 was a convective event that resulted from an outflow boundary from thunderstorms that formed west of the measurement site. An outflow boundary from one of the thunderstorms moved through the area causing a dust wall (Haboob type). This convective event lasted for a short duration (20 minutes), ~~has~~ had a sharp increase of wind speed up to 16.5 m sec^{-1} (from 2.6 m sec^{-1} measured three minutes earlier), wind gust up to 23.6 m sec^{-1} with a decrease of visibility down to 0.4 km (Fig. 2B). The thunderstorms generated moderate precipitation that started at 18:41 and lasted until 20:50. After the precipitation ended the visibility increased again to 10 km (Fig. 2B). A movie showing the dust event with the Haboob and the precipitation that follows can be found in Movie S1B.

The third dust event was also a convective dust type, that occurred on June 21, 2019. During the morning hours, a dryline moved east towards the area where it stalled just to the west of the research area later in the afternoon. Thunderstorms developed east of the dryline, southeast to AEROS, ~ 70 km from the measurements station (see Fig. S2), due to the lift that occurred in front of the boundary. An outflow boundary from the thunderstorm moved northwest through the observational domain ~~generated~~ generating the dust event. This dust event was short (30 minutes) as can be seen by the sharp and short increase of wind speed from 4.1 to 10.8 m sec^{-1} from 20:40 to 20:45, wind speed by 21:00 reach~~ed~~ 13.4 m sec^{-1} (with a gust of

16.5 m sec^{-1}) and visibility decreased to 4.8 km (Fig. 2C). A movie showing the dust particle's outflow can be found in Movie S1C.

150
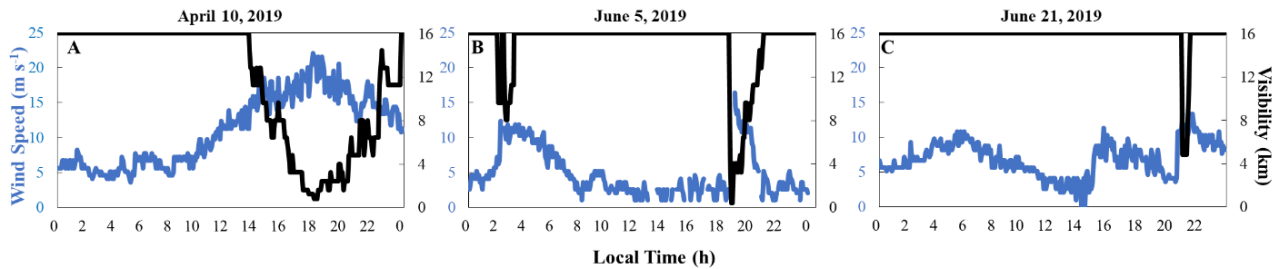
Figure 2. Changes in meteorological conditions wind speed (blue) and visibility (black) as measured on each of the dust event days.

3.2. PM concentration during the three dust events

PM concentrations were measured during each of the dust events. Unfortunately, Grimm-11D was not operational on April 10, and therefore PM comparisons were only made based on the DustTrak unit. A comparison of daily average values of each day shows relatively high SD values for each of these three dust days (Table 1). April 10 had a higher daily concentration for all PM sizes compared to June 5 and 21. The daily average values on April 10 were 70.1 ± 111 µg m^{-3} and 125 ± 182 µg m^{-3} for $PM_{2.5}$ and PM_{10}, respectively. These values exceed the WHO health recommended values of $PM_{2.5}$ and PM_{10} daily threshold, and the EPA recommended daily threshold for $PM_{2.5}$, but not for PM_{10}. PM_1 daily average for the April 10 dust event was 62.37 ± 101 µg m^{-3}, since there is no standard set for PM_1 we could not evaluate their impact (Griffiths et al., 2018). The $PM_{2.5}$ and PM_{10} daily concentrations on June 5 were 22.2 ± 126 µg m^{-3} and 29.5 ± 184 µg m^{-3}, respectively. These daily average values did not exceed EPA daily threshold or WHO threshold, but the $PM_{2.5}$ values were above the daily threshold based on the new WHO air quality guidelines (WHO, 2021). PM_1 daily average for June 5 dust event was 21 ± 121 µg m^{-3}. On June 21 the $PM_{2.5}$ daily average was 27.7 ± 99 µg m^{-3}, which was above WHO health recommended values but not EPA's. The PM_{10} daily average value was 37.8 ± 129 µg m^{-3} and did not exceed the WHO or EPA's thresholds. PM_1 daily average for the June 21 dust event was 26.5 ± 95 µg m^{-3}. If observation would have been made only based on the daily average values, without the ability to look at the high SD values, one would not have suspected a dust event occur on the two June days, and these days would have been classified as clean days. These findings are similar to those found by Kelley and Ardon-Dryer (2021) where dust events during synoptic days surpass the EPA recommended daily threshold for $PM_{2.5}$, while convective dust events were under the $PM_{2.5}$ threshold.

Table 1: Daily average values for PM_{10}, $PM_{2.5,}$ and PM_1 measured by the DustTrak during each day.

Daily average values (µg m^{-3})	April 10	June 5	June 21

PM$_1$	62.3 ± 101	21 ± 121	26.5 ± 95
PM$_{2.5}$	70.1 ± 111	22.2 ± 126	27.7 ± 99
PM$_{10}$	125 ± 182	29.5 ± 184	37.8 ± 129

The daily average values found in these three events were lower than those measured in other locations such as the
175 Mediterranean (Alghamdi et al., 2015; Krasnov et al., 2016; Saraga et al., 2017), Asia (Tsai et al., 2012; Sarkar et al., 2109)
and Africa (Kandler et al., 2009; Bouet et al., 2019). The proximity of these different locations to large dust sources, compared
to this region, could explain the higher PM values. Although several dust events in Southern Tunisia had daily
PM$_{10}$ concentrations in a similar range to those measured during the synoptic day of April 10 (~125 µg m^{-3}), many other
Southern Tunisia dust events had much higher daily PM$_{10}$ values (>1500 µg m^{-3}; Bouet et al., 2019). Even other locations in
180 the US (e.g., Arizona) had higher PM$_{10}$ concentrations daily with daily values up to 1,972 µg m^{-3} (Hyde et al., 2018). PM$_{10}$
daily values for the synoptic day were in the same range as those measured during dust events in this region by Stout (2001).
For PM$_{2.5}$, the daily concentrations PM$_{2.5}$ values of on these three days were in a similar range to those measured in the same
location in previous dust events (Kelley and Ardon-Dryer, 2021) but higher than those measured in the Great Basin region
(Hahnenberger and Nicoll, 2012).

185

Next, PM$_1$, PM$_{2.5}$, and PM$_{10}$ values were calculated for hourly and 10-minute time intervals to capture changes in the PM
concentrations over a shorter duration (Fig. 3), mainly since the convective dust events were shorter (20 - 30 minutes long).
Dust particles during the synoptic event (April 10) were present in the atmosphere for 12 h, starting from noon when an increase
in PM was observed until midnight when PM concentration decreased back to background levels (Fig. 3A, 3D). The highest
190 PM concentration (hourly average ± SD) was measured at 5 pm with a concentration of 298 ± 116 µg m^{-3} for PM$_1$, 327 ± 123
µg m^{-3} for PM$_{2.5}$, and 539 ± 186 µg m^{-3} for PM$_{10}$. Calculation based on 10-minutes shows that the highest PM concentrations
were measured at 19:10, PM$_1$ was 483 ± 58 µg m^{-3} while PM$_{2.5}$ and PM$_{10}$ were 533 ± 61, and 850 ± 90 µg m^{-3}, respectively
(Fig. 3D). The fluctuations in PM measurement based on the 10-minute average were caused by the fluctuation of wind speed.
Comparison between the wind speed to the 10-minute average PMs concentrations had high R^2 values (> 0.67). Unlike the
195 synoptic dust event, the two convective events were shorter in duration and had different PM concentrations. The highest
hourly PM concentrations on June 5 were measured at 18:00, with 187 ± 408 µg m^{-3} for PM$_1$, 196 ± 425 µg m^{-3}, and 280 ±
621 µg m^{-3} for PM$_{2.5}$ and PM$_{10}$, respectively. Lower hourly PM concentrations were measured than those measured during the
synoptic dust event. But when the observation was made based on a 10-minute average, the highest PM concentrations, which
were measured at 18:10, had PM$_1$ of 922 ± 577 µg m^{-3}, PM$_{2.5}$ and PM$_{10}$ of 964 ± 600 µg m^{-3} and 1403 ± 884 µg m^{-3}, respectively,
200 higher than those measured during the synoptic event (based on 10-minutes). The dust event on June 21 had the highest hourly
PM concentrations at 21:00, with PM$_1$ of 170 ± 251 µg m^{-3} and PM$_{2.5}$ and PM$_{10}$ of 178 ± 261 µg m^{-3} and 238 ± 340 µg m^{-3},
respectively. These hourly values were also lower than those measured on the synoptic dust event. But when observations were

made based on the 10-minutes, PM values were higher than those during the synoptic event. The highest 10-minutes PM concentrations which were measured at 21:00 had PM_1 of 684 ± 153 µg m^{-3}, and $PM_{2.5}$ and PM_{10} of 710 ± 159 µg m^{-3} and 927 ± 192 µg m^{-3}, respectively. The high SD values of the PM concentrations ~~values~~ for the two convective events can reflect the short duration of these dust events. It should be noted although the calculation of PM based on 5-minutes enhanced the difference between these three dust events, no statistical difference (based on the ANOVA test) was observed between 5-minutes and 10-minutes (Fig. S3).
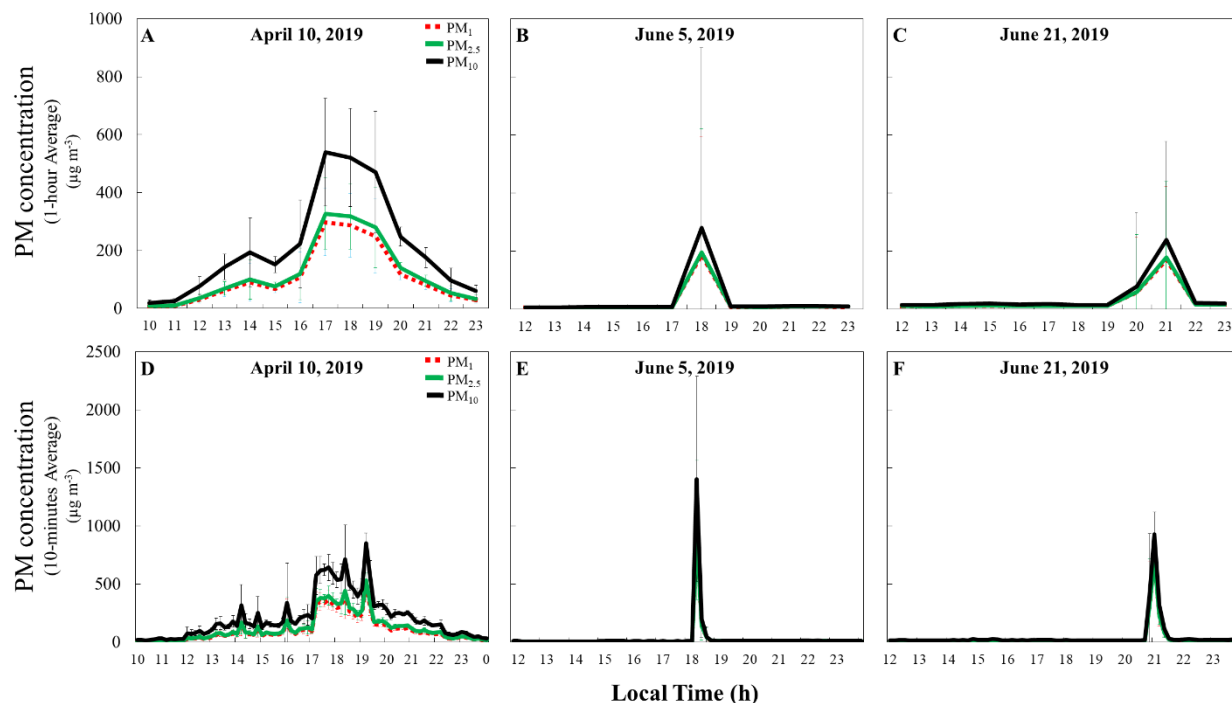
Figure 3. Changes in PM concentration (PM_1 in red, $PM_{2.5}$ in green, and PM_{10} in black) were measured by DustTrak during the three dust events, April 10 (A, D), June 5 (B, E), and June 21 (C, F), for hourly average (upper panel) and 10 minutes average (lower panel).

It is interesting to note that since the synoptic dust event had much stronger wind speeds it was expected that it will generate more particles and have higher PM concentrations than those in the convective events. It is expected that the rate of increase of PM concentrations from the background level (before the dust) could reflect that change. Calculations were made to evaluate how much PM values increased in each of these events. Comparison of the change of $PM_{2.5}$ and PM_{10} concentration from before the dust event to the highest measured during the dust event (based on the 10-minutes average) were performed. We expected that the synoptic dust event will have the highest increase but to our surprise, on April 10, the increase of $PM_{2.5}$ and PM_{10} was 533 and 848 µg m^{-3}, respectively. This increase was lower compared to those calculated on the two convective days. The increase of $PM_{2.5}$ and PM_{10} were 958 and 1397 µg m^{-3}, respectively on June 5, and 701 and 915 µg m^{-3}, respectively on June 21.

All three dust events had an increase in the PM values in all the three PM sizes examined (PM_1, $PM_{2.5}$, and PM_{10}). The hourly and daily $PM_{2.5}$ values were in the same range as those measured in this region during previous dust events (Kelley et al., 2020; Kelley and Ardon-Dryer, 2021). As for PM_{10}, the hourly values measured in each of the three dust events were in the same range as those measured during dust events in El Paso TX (Novlan et al., 2007) and Phoenix AZ (Hyde et al., 2018). While an increase in $PM_{2.5}$ and PM_{10} was expected, the increase in PM_1 values was interesting as the increase was observed in each of the different dust events observed in this study. While few studies measured PM_1 concentrations during dust events, some of them did not observe an increase in PM_1 during dust events (Jaafari et al., 2018; Claiborn et al., 2000), while others measured an increase in PM_1 concentrations (Alghamdi et al., 2015) but not as strong as presented in this work.

Most of the studies that examine human exposure thresholds are based on annual or daily values (e.g., WHO and EPA threshold), but there are no guidelines or accepted thresholds for short-term (15-min to 1-h) exposure (Griffiths et al., 2018). The need for the development of short-term PM exposure guidance has become more crucial due to the increase of periods with short-term high PM concentrations (Deary and Griffiths, 2021). Many studies use short term observation (mainly based on 1-h) as a key to understanding the impact of episodic air pollution events on the exposure to PM, but majority of the studies focus on wildfires and anthropogenic pollution (Griffiths et al., 2018; Brilli et al., 2021; Deary and Griffiths, 2021). The idea of the short-term threshold is that the hourly sampling interval can be used to characterize short-term exposure to critical PM concentration and provide information on the number of critical episodes and their persistence (Brilli et al., 2021) which could be critical for convective dust events.

Observation of PMs during dust events is normally reported on an hourly basis (Hahnenberger and Nicoll, 2012; Krasnov et al., 2014; Hyde et al., 2018). Fewer studies measured changes in particles during dust events using high temporal (5-minute, or 10-minute) intervals. One of these studies is Bouet et al. (2019) who presented similar findings to those shown in this work. In their study, they compared two dusty days that had similar daily PM_{10} values, yet when the observations were made based on a shorter time scale (5-minutes) one of the dust events had very high PM_{10} values, which were about two orders of magnitude higher from the daily average values. Although no information about the dust type of the dust event was provided, the fast and sharp increase and decrease of PM concentrations and wind speed suggest that the examined dust event was convective. The observation from Bouet et al. (2019) and those presented in this work validate the assumptions made by Kelley and Ardon-Dryer (2021) that hourly values mask the actual PM concentrations of short duration convective dust events. It seems that observation based on a daily and even hourly basis underestimates the impact of short-duration dust events and will not allow reflecting the true impact these might have on atmospheric particles and air quality. It is important to examine the impact of short-term events as it has been suggested that short-term exposure to high PM concentrations could cause various health issues (e.g., cardiovascular disease Martinelli et al., 2013). Several studies also found a correlation between short-term exposure to coarse PM and various health issues (Brunekreef and Forsberg, 2005; Host et al., 2008; Pérez et al., 2008; Graff

et al., 2009; Linares et al., 2010; Malig and Ostro, 2009; Tobias et al., 2019). While in extreme cases, people and animals caught in severe dust events have lost their lives due to suffocation (Idso et al 1976; Middleton, 2017). The fact that some dust events are short-term raises the question of their impact on air quality and human health. As suggested in Bouet et al. (2019), the health consequences of such intense but relatively short exposure needed spatial definition of air quality standards accounting for the intermittency of dust emission which should be encouraged in regions where such phenomenon strongly controls the air quality.

3.3 Comparison of particle size distribution and total concentration

Based on the differences in PM concentrations we were wondering how the particle size distributions of these dust events will be, whether they will have more large or small (e.g., inhalable) particles. A comparison of the size distribution for the size range of 0.3 to 10 µm using the OPS which was operated in during all three dust events was performed. Since the Grimm-11D did not operate during April 10 dust event, it could not be used for this comparison, yet it should be noted that the OPS and Grimm-11D measured similar concentrations and size distribution during the two convective days when similar particle size range (0.3 to 10 µm) was used (Fig. S4).

To evaluate how the particle size distribution changes on each of these three days, different calculations were made. The first, one represents the entire day (daily averaged), the second, 10-minutes averages of the time before the dust and during the peak of the dust event (represented by the time of the highest PM concentrations; 19:10 for April 10, 18:10 for June 5, and 21:00 for June 21). All three dust events' particle size distribution during the peak had much higher particle concentrations compared to times before the dust event or the daily average (Fig.4). For particles larger than 1 µm the difference in concentration was ~2 orders of magnitudes. For the April 10 dust event, the high particle concentration was observed for all particle sizes (Fig. 4A), while for the two convective dust events the differences were high only for particles >0.6 µm (Fig. 4B and 4C). As discussed earlier, the daily average for the convective days does not represent the true size distribution of these convective dust events, since the convective events lasted only 20 or 30 minutes, these durations only represent 2% of the time of the day. The size distributions during the peak of the dust (black line in Fig. 4) were compared between the different dust events (Fig 4D). This comparison showed that the two convective dust events had very similar size distributions. A difference between these convective dust events was observed for particles in the size range of 0.3 to 0.37 µm, where June 21 had much higher concentrations of small particles compared to those measured on June 5. The synoptic dust event had much higher particle concentrations in the size range of 0.4 - 4 µm, but for particles > 4 µm the three dust events had similar concentrations, in the same range.

Generally, dust events are characterized by high concentrations of coarse particles >2.5µm (Clements et al., 2013). However, these three dust events show that dust events in this region can also contain small particles (<2 µm). Particle concentrations of

these sizes were higher by an order of magnitude than those measured before the dust event. These findings are similar to those measured during dust storms in Israel (Ardon-Dryer and Levin, 2014). Coarseourse particles also increased in each of these dust events. The increase of particles concentration during the dust peak compared to the time before the dust, (for particle size >5 -10 µm, based on OPS) ranges from 30 up to 300 times more during the peek peak of the dust, and all three dust events had similar concentrations of particles at that size range (Fig. 4D). Next, an examination of the size distribution of coarse particles of the two convective events, using Grimm 11-D which tracks particles up to 35.15 µm was performed (Fig. 4E). It should be noted that the unit was not operated during the synoptic dust event. Observation of particle size distribution of particles larger than 10 µm (Fig. 4E) shows that some of the coarse particle sizes concentration increased by more than two orders of magnitude compared to the time before the dust event. In addition, an increase in particle concentration was observed for both convective dust events in the larger size bin (35.15 µm). The size of dust particles during dust events is vital for our understanding of how far these dust particles can travel from the source, as well as their health implications. The Mainly as the size distributions of dust particles during dust events are not well understood (Mahowald et al., 2014). Moreover, recently it has been stated that the atmosphere contains four times more coarse dust particles (>5 µm) than what is currently simulated in climate models, which ends in a substantial underestimation of the impact of coarse dust particles may have on the Earth system (Adebiyi and Kok, 2020). Therefore, Mahowald et al. (2014) suggested that models should improve their ability to capture the evolution of dust size distribution which should be based on additional cross-comparison of differing observational methods. These might be impacted by the proximity of the measurement location to the dust source (closer to the source meaning more coarse particles) as well as to the meteorological conditions that generated the dust event. This study provides measurements of particles >5 µm during three different dust events (of different types) showing that the concentration of coarse particles (>5 µm) may increase even by two orders of magnitude during dust events.

All three dust events seem to have multimodal distribution, with the first mode at the lowest size measured (0.3 µm), the second at 1 µm, and the third at 1.94 µm. Gillette et al. (1974) who examine the size distribution of particles (in the range of 1 - 20 µm) collected from Amarillo Texas (~170 km north of AEROS), using a wind tunnel, found a mode at 1 µm. Dust in Morocco also had multimodal distribution but at different sizes (Kandler et al., 2009). Dust in Lebanon had a single mode at much smaller particle sizes (~0.28 µm) compared to those found in this work (Jaafar et al., 2014). Analysis of dry deposition dust at different locations around the world found the mode to be at much larger particles sizes than those measured in this work (Reynolds et al., 2020; Katra and Krasnov, 2020), but the collection and analysis method most likely contribute to these differences. The size distribution from these three dust events (Fig 4D) was also in a similar range to those measured in Sahara (D'Almeida and Schütz, 1983) but much lower than those measured during dust in Asia (Chun et al., 2001) or Africa (Pio et al., 2014). The synoptic dust event (at the peak) had a size distribution in a similar range to those measured by Kandler et al. (2009) in Morocco, at least for particles size < 2 µm. But all three dust events had slightly higher concentrations for particles larger than 2 µm. The size distribution of the three dust events was similar to measurements during dust events in Israel, at least for particles up to 1 µm (Reicher et al., 2019). The higher concentration of larger particles in this study compared to those

in Reicher et al. (2019) could be attributed to the fact that measurements from this work were taken close to the source while dust collection in Israel was after dust particles had to travel a long distance to reach the measurements site. Daily values from the synoptic days were in a similar range to the size distribution measured in Israel during dust days (Ardon-Dryer and Levin, 2014). Both daily distributions of the convective days had similar concentrations as to clean days measured in Ardon-Dryer and Levin (2014) at least for size ranges of 0.2 to 0.6 μm. For particles >0.6 μm, the daily distribution of the convective days had higher particle concentrations.

Using the OPS total number concentration values comparison between the dust events was performed for particles in the size range of 0.3 to 10 μm. The total number concentration values measured at the highest PM concentrations on April 10 (19:10) was 156 # cm^{-3}, on June 5 (18:10) it reached up to 82 # cm^{-3}, while on June 21 (21:00) the maximum total number concentration was 93 # cm^{-3} at 21:00 (Fig. 4E4F). The total number concentration values behaved similar to the changes of PM, but in this comparison, April 10 event had a much higher total number concentration than the convective events. It should be noted that the calculation of the total number concentration based on a 5- or 10-minutes average did not show any significance between the two (data not shown). The total particle number concentrations in these dust events were much higher than those measured during dust events at Storm Peak Laboratory in CO (Hallar et al., 2011), but the proximity to the source in this region could attribute to that. The total particle number concentrations during these dust events were much lower than those measured during biomass burning events (Reid et al., 2005; Ordou and Agranovski, 2019), the emitted particles' sizes are the main cause for that difference.

Comparison of total number concentration for coarse particle size, as measured by Grimm 11-D, was also performed for the two convective dust events. During the June 5 dust event (based on a 10-minutes average at the peak of the dust) the total number concentration for particles 5 to 35.15 μm was 4.6 ± 2.9 cm^{-1} while for particles in the size range of 10 to 35.15 μm the total number concentration was 1.3 ± 0.9 cm^{-1}. The increase in total number concentration was more than 350 to 675 times higher than the total number concentration right before the dust reached the station for particles > 5 and > 10 μm, respectively. The total number concentration during the June 21 convective dust event (based on a 10-minutes average at the peak of the dust) was slightly lower than those measured on June 5 with 2.3 ± 0.7 cm^{-1} and 0.5 ± 0.3 cm^{-1} for particles range of 5 to 35.15 μm and 10 to 35.15 μm, respectively. The increase in total number concentration was more than 141 to 318 times higher than the total number concentration right before the dust reached the station for particles > 5 and > 10 μm, respectively.

This study provides measurements for particle size distribution and total number concentrations of particles > 5 μm during the three different dust events (of different types) and for particles > 10 μm for the two convective events showing that the concentration of these coarse particles may increase by more than two orders of magnitude during dust events. These findings are in line with recent studies that found coarse particles during dust events near the source (Ryder et al., 2019; O'Sullivan et al., 2020), and even thousands of kilometers from the sources (Weinzierl et al., 2017; van der Does et al., 2018). Moreover,

recently it has been stated that the atmosphere contains four times more coarse dust particles than what is currently simulated in climate models, which ends in a substantial underestimation of the impact coarse dust particles may have on the Earth system (Adebiyi and Kok, 2020). Therefore, Mahowald et al. (2014) suggested that models should improve their ability to capture the evolution of dust size distribution which should be based on additional cross-comparison of differing observational methods. Such effort has taken place in recent years, yet many of these studies indicate that models still cannot capture some of the super coarse particles due to their deposition process which is still unclear (Drakaki et al., 2022; Meng et al., 2022). In addition, some of the differences between measurements and models might be impacted by the proximity of the measurement location to the dust source (closer to the source meaning more coarse particles) as well as to the meteorological conditions that generated the dust event.
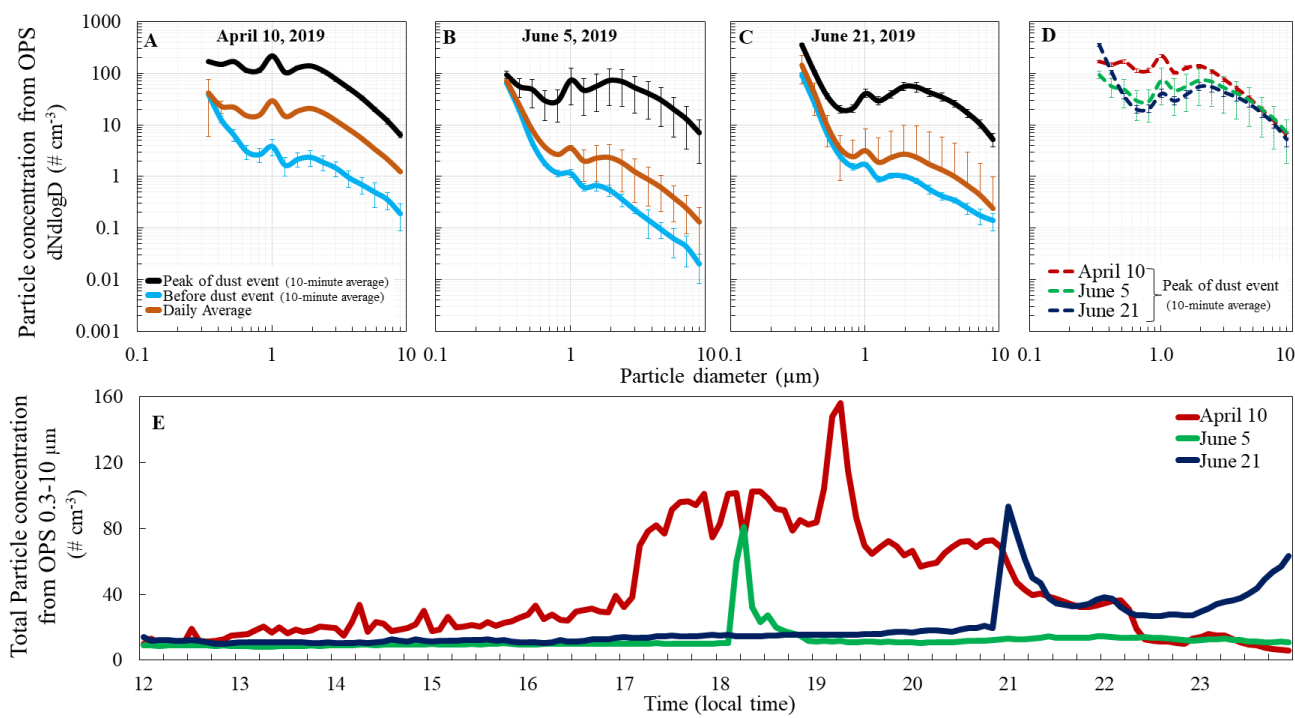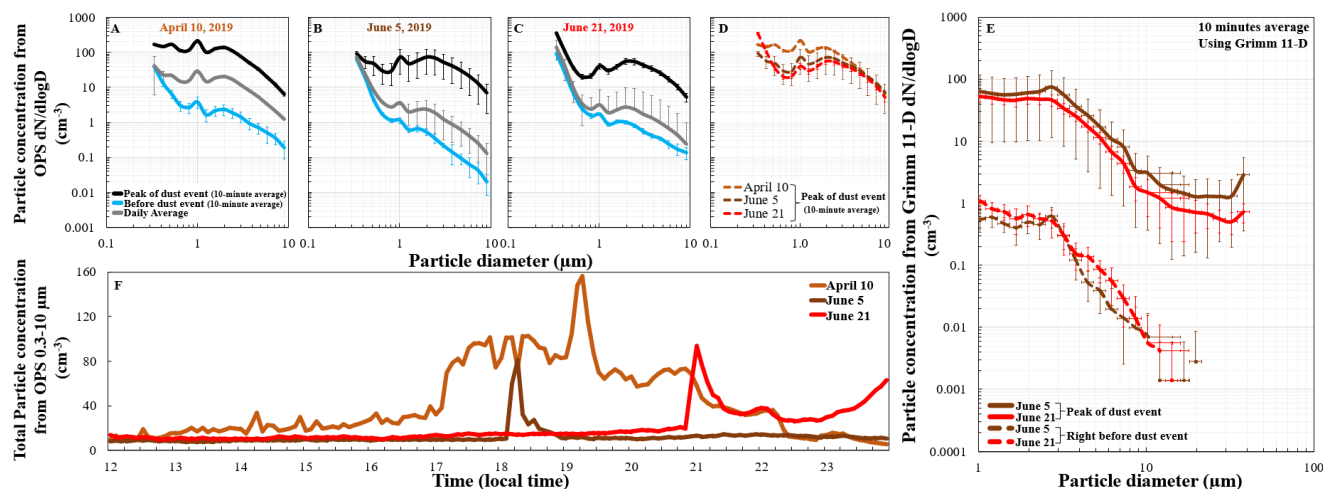
Figure 4. Changes in particle size distribution based on optical particle diameter, as measured by OPS, during the three dust events, April 10 (A), June 5 (B), and June 21 (C). The peak of the dust (10-minutes average for time with the highest concentration (black), a time before dust reached the station (10-minutes average in dark light blue), and daily average (~~orange~~gray). Comparison of the three size distributions at peak of the dust (10-minutes average for time with the highest concentration) in D. Comparison of 10 minutes average size distribution from June 5 (dark brown) and June 21 (red) as measured, with Grimm 11-D, at the peak of the dust event (straight line) and right before the dust event (in dashed line) in E. The particle's total number concentration (0.3 to 10 μm) from OPS for each of the dust events (~~E~~F) for April 10 (~~red~~light brown), June 5 (dark brown~~green~~), and June 21 (~~dark blue~~red).

While the measurements from these three dust events provide an insight into the changes in PM concentration, size distribution, and total particle number concentrations of dust events in West Texas. Additional measurements during dust events in this region and across the world are needed to help improve our understanding of how dust events impact particle concentrations and sizes, which might cause a different impact on air quality and human health. In addition, more measurements during different types of dust events (convective vs synoptic) ~~are needed~~ will improve our understanding of their implications.

4 Conclusions

Three dust events were captured by AEROS during April and June 2019. One of the dust events was synoptic (April 10, 2019) while the other two were convective (June 5 and 21, 2019). Measurements of particle mass and number concentration, size distribution, and meteorological conditions for each dust event were performed and compared. The Synoptic dust event (~~of~~ April 10) was longer (12h) and had stronger wind speed conditions (up to 22.1 m sec^{-1}), while the two convective dust events lasted less than 30 minutes and had lower wind speeds (up to 16.5 and 13.4 m sec^{-1} for June 5 and 21, respectively). Observation of PM based on daily and hourly values seems to underestimate the impact of the convective dust events. Observations based on 10-minutes reveal the true impact of the two convective dust events, w. ~~W~~ith PM concentrations even higher than the synoptic dust event. A comparison of particle size distribution showed that all three dust events had an increase in particle~~s~~ concentration in all sizes measured (0.3 to 10 μm). Some of the particle sizes had an increase in particle concentration of more

than ~2 orders of magnitude during the dust event compared to before it. All three dust events had similar particle concentrations for particle sizes > 5 µm. Further research is needed to determine the effects of additional short and intense dust events on particle concentration and sizes and on what impact they might have on air quality and human health.

395

Code availability. MATLAB codes can be obtained from the authors per request.

Data availability. All measurements are available per request.

400

Author contribution. KAD designed the experiments, supervised the entire process, and performed most of the analysis, in addition to writing the manuscript. MK maintained and managed AEROS during the sampling period. Both authors were actively involved in interpreting results and in discussions on the manuscript.

405 **Declaration of competing interest**. The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Acknowledgment. This research did not receive any specific grant from funding agencies in the public, commercial, or not-for-profit sectors. The authors would like to thank Texas Tech University for the support of Mary Kelley's scholarship and
410 Yuval Dryer for his help with the MATLAB codes.

References

Abdullaev, S.F., and Sokolik, I.: Assessment of the Influences of Dust Storms on Cotton Production in Tajikistan. In: Gutman, G., J. Chen, G. Henebry, and M. Kappas (eds.) Landscape Dynamics of Drylands across Greater Central Asia: People, Societies and Ecosystems. Landscape Series, Springer, Cham., 17, https://doi.org/10.1007/978-3-030-30742-4_6, 2020.
415 Achakulwisut, P., Mickley L., and Anenberg S.: Drought-sensitivity of fine dust in the US Southwest: implications for air quality and public health under future climate change. Environ. Res. Lett., 13, 2018.
Adebiyi, A.A., and Kok, J.F.: Climate models miss most of the coarse dust in the atmosphere. Sci. Adv. 6, 15, https://doi.org/10.1126/sciadv.aaz9507, 2020.
Aghababaeian, H., Ostadtaghizadeh, A., Ardalan, A., Asgary, A., Akbary, M., Yekaninejad, M.S., and Stephens, C.: Global
420 Health Impacts of Dust Storms: A Systematic Review. Enviro. Health Insights, 15, https://doi.org/10.1177%2F11786302211018390, 2021.

Alghamdi, M.A., Almazroui, M., Shamy, M., Redal, M.A., Alkhalaf, A.K., Hussein, M.A., and Khoder, M.I.: Characterization and elemental composition of atmospheric aerosol loads during springtime dust storm in western Saudi Arabia, Aerosol Air Qual. Res., 15, 440–453, https://doi.org/10.4209/aaqr.2014.06.0110, 2015.

425 Al-Hemoud, A., Al-Dousari, A., Misak, R., Al-Sudairawi, M., Naseeb, A., Al-Dashti, H., and Al- Dousari, N.: Economic impact and risk assessment of sand and dust storms (SDS) on the Oil and gas Industry in Kuwait. Sustainability 11 (1), 200, https://doi. org/10.3390/su11010200, 2019.

Ardon-Dryer, K., and Levin, Z.: Ground-based measurements of immersion freezing in the eastern Mediterranean, Atmos. Chem. Phys., 14, 5217–5231, https://doi.org/10.5194/acp-14-5217-2014, 2014.

430 Ardon-Dryer, K., Mock, C., Reyes, J., and Lahav, G.: The effect of dust storm particles on single human lung cancer cells. Environ. Res., 181, 108891, https://doi.org/10.1016/j.envres.2019.108891, 2020.

Ardon-Dryer, K., Chmielewski, V., Burning E., and Xueting X.: Changes of Electric Field, Aerosol, and Wind Covariance in Different Blowing Dust Days in West Texas, Aeolian Res., 100762, https://doi.org/10.1016/j.aeolia.2021.100762, 2021.

Ardon-Dryer, K., Kelley, M.C., Xueting, X., and Dryer, Y.: The Aerosol Research Observation Station (AEROS), Atmos.
435 Meas. Tech., https://doi.org/10.5194/amt-2021-270, preprint, 2022.

Bhattachan, A., Okin, G.S., Zhang, J., Vimal, S., and Lettenmaier, D.P.: Characterizing the role of wind and dust in traffic accidents in California. GeoHealth 3, 328–336, https://doi.org/10.1029/2019GH000212, 2019.

Bogan, M., Al, B., Kul, S., Zengin, S., Oktay, M., Sabak, M., Gumusboga, H., and Bayram, H.: The effects of desert dust storms, air pollution, and temperature on morbidity due to spontaneous abortions and toxemia of pregnancy: 5-year analysis.
440 Int. J. Biometeorol., 65(10), 1733-1739, https://doi.org/10.1007/s00484-021-02127-8, 2021.

Bouet, C., Labiadh, M.T., Rajot, J.L., Bergametti, G., Marticorena, B., Henry des Tureaux, T., Ltifi, M., Sekrafi, S., and Féron, A.: Impact of Desert Dust on Air Quality: What is the Meaningfulness of Daily PM Standards in Regions Close to the Sources? The Example of Southern Tunisia. Atmosphere., 10(8), 452, https://doi.org/10.3390/atmos10080452, 2019.

Brey, S.J., Pierce J.R., Barnes E.A., and Fischer, E.V.: Estimating the Spread in Future Fine Dust Concentrations in the
445 Southwest United States, J. Geophys. Res., 125, no. 21, https://doi.org/10.1029/2019JD031735, 2020.

Brilli, L., Carotenuto, F., Andreini, B.P., Cavaliere, A., Esposito, A., Gioli, B., Martelli, F., Stefanelli, M., Vagnoli, C., Venturi, S., Zaldei, A., and Gualtieri, G.: Low-Cost Air Quality Stations' Capability to Integrate Reference Stations in Particulate Matter Dynamics Assessment. Atmosphere. 12(8):1065. https://doi.org/10.3390/atmos12081065,2021

Brunekreef, B., and Forsberg, B.: Epidemiological evidence of effects of coarse airborne particles on health, Eur Respir J.,
450 26(2), 309-318, https://doi.org/10.1183/09031936.05.00001805, 2005.

Chen, Q., Yin, Y., Jiang, H., Chu, Z., Xue, L., Shi, R., Zhang, X., and Chen, J.: The roles of mineral dust as cloud condensation nuclei and ice nuclei during the evolution of a hailstorm, J. Geophys. Res. Atmos., 124, 14, 262–14,284. https://doi.org/10.1029/2019JD031403, 2019.

Chun, Y., Boo, K-O., Kim, J., Park, S-U., and Lee, M.: Synopsis, transport, and physical characteristics of Asian dust in Korea,
455 J. Geophys. Res. Atmos., 106, 18461-18469, https://doi.org/10.1029/2001JD900184, 2001.

Claiborn, C.S., Finn, D., Larson, T.V., and Koenig, J.Q.: Windblown dust contributes to high $PM_{2.5}$ concentrations, J. Air Waste Manag. Assoc., 50, 1440-1445, https://doi.org/10.1080/10473289.2000.10464179, 2000.

Clements, A. L., Fraser, M. P., Upadhyay, N., Herckes, P., Sundblom, M., Lantz, J., and Solomon, P.A.: Characterization of summertime coarse particulate matter in the Desert Southwest – Arizona, USA, J. Air Waste Manag. Assoc., 63, 764-772, https://doi.org/10.1080/10962247.2013.787955, 2013.

Crooks, J.L., Cascio, W.E., Percy, M.S., Reyes, J., Neas, L.M., and Hilborn, E.D.: The association between dust storms and daily non-accidental mortality in the United States, 1993–2005, Environ. Health Perspect., 124, 1735–43, https://doi.org/10.1289/ehp216, 2016.

D'Almeida, G.A., and Schutz, L.: Number, Mass and Volume Distributions of Mineral Aerosol and Soils of the Sahara, 22, 233-243, https://doi.org/10.1175/1520-0450(1983)022%3C0233:NMAVDO%3E2.0.CO;2, 1983.

Dastoorpoor, M., Idani, E., Goudarzi, G., Khanjani, N.: Acute effects of air pollution on spontaneous abortion, premature delivery, and stillbirth in Ahvaz, Iran: a time-series study, Environ. Sci. Pollut. Res., 25, 5447–5458, https://doi.org/10.1007/s11356-017-0692-9, 2018.

Deary, M., and Griffiths, S.: A novel approach to the development of 1-hour threshold concentrations for exposure to particulate matter during episodic air pollution events, J. Hazard. Mater., 418, https://doi.org/10.1016/j.jhazmat.2021.126334, 2021.

Diokhane, A.M., Jenkins, N., Manga, M.S., Drame, and Mbodji, B.: Linkages between observed, modeled Saharan dust loading and meningitis in Senegal during 2012 and 2013, Int. J. Biometeorol, 60 (4), 557–575, https://doi.org/10.1007/s00484-015- 1051-5, 2016.

Drakaki, E., Amiridis, V., Tsekeri, A., Gkikas, A., Proestakis, E., Mallios, S., Solomos, S., Spyrou, C., Marinou, E., Ryder, C., Bouris, D. and Katsafados, P.: Modelling coarse and giant desert dust particles, Atmos. Chem. Phys. Discuss., 2022, 1–36, doi:10.5194/acp-2022-94, 2022.

EPA (United States Environmental Protection Agency), 2016. NAAQS table. https://www.epa.gov/criteria-air-pollutants/naaqs-table. last access: 11 September 2021.

Gillette, D.A., Blifford Jr., I.H. and Fryrear, D.W.: The influence of wind velocity on the size distributions of aerosols generated by the wind erosion of soils, J. Geophys. Res., 79(27), 4068–4075, https://doi.org/10.1029/JC079i027p04068, 1974.

Goudie, A., and Middleton, N. J.: Desert dust in the global system. Springer, Dordrecht, 2006.

Goudie, A.S.: Desert Dust and Human Health Disorders, Environ. Int., 63, 101–13, https://doi.org/10.1016/j.envint.2013.10.011, 2014.

Graff, D.W., Cascio, W.E., Rappold, A., Zhou, H., Huang, Y.C., and Devlin, R.B.: Exposure to concentrated coarse air pollution particles causes mild cardiopulmonary effects in healthy young adults, Environ Health Perspect., 117(7), 1089-94, https://doi.org/10.1289/ehp0900558, 2009.

Griffiths, S.D., Chappell, P., Entwistle, J.A., Kelly, F.J., and Deary, M.E.: A study of particulate emissions during 23 major industrial fires: Implications for human health, Environ. Int., 112, 310–323, https://doi.org/10.1016/j.envint.2017.12.018, 2018.

Grimm 11-D: https://www.grimm-aerosol.com/products-en/dust-monitors/the-dust-decoder/11-d/, last access: 5 August 2021.

Hahnenberger, M., and Nicoll, K.: Meteorological Characteristics of Dust Storm Events in the Eastern Great Basin of Utah, U.S.A., Atmos. Environ., 60, 601–12, https://doi.org/10.1016/j.atmosenv.2012. 06.029, 2012.

Hallar, A.G., Chirokova, G., McCubbin, I., Painter, T.H., Wiedinmyer, C., and Dodson, C.: Atmospheric bioaerosols transported via dust storms in the western United States, Geophys. Res. Lett., 38, L17801, https://doi.org/10.1029/2011GL048166, 2011.

Hand, J.L., White, W.H., Gebhart, K.A., Hyslop, N.P., Gill, T.E., and Schichtel, B.A.: Earlier onset of the spring fine dust season in the southwestern United States, Geophys. Res. Lett., 43, 4001–4009, http://dx.doi.org/10.1002/2016GL068519, 2016.

Herrera-Molina, E., Gill, T.E., Ibarra-Mejia, G., and Jeon, S.: Associations between Dust Exposure and Hospitalizations in El Paso, Texas, USA, Atmosphere., 12(11), 1413, https://doi.org/10.3390/atmos12111413, 2021.

Host, S., Larrieu, S., Pascal, Blanchard, L., Declercq, M., Fabre, C., Jusot, P., Chardon, J. F., Le Tertre, B., Wagner, A., Prouvost, V., Lefranc, H.: Short-term associations between fine and coarse particles and hospital admissions for cardiorespiratory diseases in six French cities, Occupational and Environmental Medicine, 65, 544-551, https://doi.org/10.1136/oem.2007.036194, 2008.

Hyde, P., Mahalov, A., and Li, J.: Simulating the meteorology and PM10 concentrations in Arizona dust storms using the Weather Research and Forecasting model with Chemistry (Wrf-Chem), J. Air Waste Manag. Assoc., 68, 177-195, https://doi.org/10.1080/10962247.2017.1357662, 2018.

Idso, S. B.: Dust Storms, SciAm., 235, no. 4, 108-115, 1976.

Jaafar, M., Baalbaki, R., Mrad, R., Daher, N., Shihadeh, A., Sioutas, C., and Saliba, N.A.: Dust episodes in Beirut and their effect on the chemical composition of coarse and fine particulate matter, Sci. Total Environ., 496, 75–83, 2014.

Jaafari, J., Naddafi, K., Yunesian, M., Nabizadeh, R., Hassanvand, M.S., Ghozikali, M.G., Nazmara, S., Shamsollahi, H.R., and Yaghmaeian, K.: Study of PM10, PM2.5, and PM1 levels in during dust storms and local air pollution events in urban and rural sites in Tehran, Hum. Ecol. Risk Assess.: Int. J., 24, 482-493, https://doi.org/10.1080/10807039.2017.1389608, 2018.

Jones, B.A.: After the Dust Settles: The Infant Health Impacts of Dust Storms, J. Assoc. Environ. Resour. Econ., 7, https://doi.org/10.1086/710242, 2020.

Kandler, K., Schutz, L., Deutscher, C., Ebert, M., Hofmann, H., Jackel, S., Jaenicke, R., Knippertz, P., Lieke, K., Massling, A., Petzold, A., Schladitz, A., Weinzierl, B., Wiedensohler, A., Zorn, S., and Weinbruch, S.: Size distribution, mass concentration, chemical and mineralogical composition and derived optical parameters of the boundary layer aerosol at Tinfou, Morocco, during SAMUM 2006, Tellus B Chem. Phys. Meteorol., 61, 32-50, https://doi.org/10.1111/j.1600-0889.2008.00385.x, 2009.

Karanasiou, A., Moreno, N., Moreno, T., Viana, M., de Leeuw, F., and Querol, X.: Health effects from Sahara dust episodes in Europe: literature review and research gaps, Environ. Int., 47, 107–114, https://doi.org/10.1016/j.envint.2012.06.012, 2012.

Katra I, and Krasnov H.: Exposure Assessment of Indoor PM Levels During Extreme Dust Episodes, *Int. J. Environ. Res.,*
525   17(5), 1625, https://doi.org/10.3390/ijerph17051625, 2020.

Kelley, M.C. and Ardon-Dryer, K.: Analyzing two decades of dust events on the Southern Great Plains region of West Texas, Atmos. Pollut. Res., 12, 101091, https://doi.org/10.1016/j.apr.2021.101091, 2021.

Kelley, M.C., Brown, M.M., Fedler, C.B., and Ardon-Dryer, K.: Long-term Measurements of PM2.5 Concentrations in Lubbock, Texas, Aerosol Air Qual. Res., 20, 1306-1318, https://doi.org/10.4209/aaqr.2019.09.0469, 2020.

530   Knippertz, P.: Meteorological Aspects of Dust Storms, Mineral Dust, Springer Dorecht, https://doi.org/10.1007/978-94-017-8978-3_6, 2014.

Krasnov, H., Katra, I., Koutrakis, P., and Michael, D.F.: Contribution of dust storms to PM10 levels in an Urban arid environment, J. AirWaste Manag. Assoc., 64 (1), 89-94, https://doi.org/10.1080/10962247.2013.841599, 2014.

Krasnov, H, Katra, I, and Friger, M.: Increase in dust storm related PM10 concentrations: A time series analysis of 2001-2015,
535   Environ Pollut., 213, 36-42, https://doi.org/10.1016/j.envpol.2015.10.021, 2016.

Lau, W.K.M., Kyu-Myong, K., Chun, Z, Ruby, L.L. and Sang-Hun, P.: Impact of Dust-Cloud-Radiation-Precipitation Dynamical Feedback on Subseasonal-to-Seasonal Variability of the Asian Summer Monsoon in Global Variable-Resolution Simulations With MPAS-CAM5, Front. Earth Sci., 8, https://doi.org/10.3389/feart.2020.00226, 2020.

Linares, C., Tobías, A., and Díaz, J.: Is there new scientific evidence to justify reconsideration of the current WHO guidelines
540   for   particulate   matter   during   dust   intrusions?,   Sci   Total   Environ.,   408(10),   2283-4, http://dx.doi.org/10.1016/j.scitotenv.2010.02.005, 2010.

Mahowald, N., Albani, S., Kok, J.F., Engelstaeder, S., Scanza, R., Ward, D.S., and Flanner, M.G.: The size distribution of desert dust aerosols and its impact on the Earth system, Aeolian Res., 15, 53-71, https://doi.org/10.1016/j.aeolia.2013.09.002, 2014.

545   Malig, B.J., Ostro, B.D.: Coarse particles and mortality: evidence from a multi-city study in California, J. Occup. Environ. Med., 66, 832-839, http://dx.doi.org/10.1136/oem.2008.045393, 2009.

Martinelli, F., Reagan, R.L., Uratsu, S.L., Phu, M.L., Albrecht, U., Zhao, W., Davis, C.E., Bowman, K.D., and Dandekar, A.M.: Gene Regulatory Networks Elucidating Huanglongbing Disease Mechanisms. PLoS ONE 8(9): e74256. https://doi.org/10.1371/journal.pone.0074256, 2013.

550   Meng, J., Huang, Y., Leung, D.M., Li, L., Adebiyi, A.A., Ryder, C.L., Mahowald, N.M. and Kok, J.F.: Improved Parameterization for the Size Distribution of Emitted Dust Aerosols Reduces Model Underestimation of Super Coarse Dust, Geophys. Res. Lett., 49(8), e2021GL097287, doi:https://doi.org/10.1029/2021GL097287, 2022.

Middleton, M.J.: Desert dust hazards: A global review, Aeolian Res., 24, 53-63, https://doi.org/10.1016/j.aeolia.2016.12.001, 2017.

555   Middleton, N.: Health in dust belt cities and beyond—an essay by Nick Middleton, BMJ 371, m3089, https://doi.org/10.1136/bmj.m3089, 2020.

Miller, R.L., and Perlwitz, T.J.: Surface radiative forcing by soil dust aerosols and the hydrologic cycle, J. Geophys. Res. Atmos., 109, https://doi.org/10.1029/2003JD004085, 2004.

Novlan, D., Hardiman, J.M., and Gill, T.E.: A synoptic climatology of blowing dust events in El Paso, Texas from 1932–

560   2005, paper presented at 16th Conference on Applied Climatology, Am. Meteorol. Soc., San Antonio, TX, 2007.

Ordou, N. and Agranovski, I.E.: Contribution of Fine Particles to Air Emission at Different Phases of Biomass Burning. Atmosphere, 10, 278, https://doi.org/10.3390/atmos10050278, 2019.

O'Sullivan, D., Marenco, F., Ryder, C. L., Pradhan, Y., Kipling, Z., Johnson, B., Benedetti, A., Brooks, M., McGill, M., Yorks, J. and Selmer, P.: Models transport Saharan dust too low in the atmosphere: a comparison of the MetUM and CAMS forecasts

565   with observations, Atmos. Chem. Phys., 20(21), 12955–12982, doi:10.5194/acp-20-12955-2020, 2020.

Pérez, L., Tobias, A., Querol, X., Kunzli, N., Pey, J., Alastuey, A., Viana, M., Valero, N., Gonzalez-Cabre, M., and Sunyer, J.: Coarse particles from Saharan dust and daily mortality, Epidemiology., 19, 800-807, https://doi.org/10.1097/ede.0b013e31818131cf, 2008.

Pio, C.A., Cardoso, J.G., Cerqueira, M.A., Calvo, A., Nunes, T.V., Alves, C.A., Custódio, D., Almeida, S.M., and Almeida-

570   Silva, M.: Seasonal variability of aerosol concentration and size distribution in Cape Verde using a continuous aerosol optical spectrometer, Front. Environ. Sci., 2, https://doi.org/10.3389/fenvs.2014.00015, 2014.

Pu, B., and Ginoux, P.: Projection of American dustiness in the late 21st century due to climate change, Sci. Reps., 7, 5553, https://doi.org/10.1038/s41598-017-05431-9, 2017.

Reicher, N., Budke, C., Eickhoff, L., Raveh-Rubin, S., Kaplan-Ashiri, I., Koop, T., and Rudich, Y.: Size-dependent ice

575   nucleation by airborne particles during dust events in the eastern Mediterranean, Atmos. Chem. Phys., 19, 11143–11158, https://doi.org/10.5194/acp-19-11143-2019, 2019.

Reid, J.S., Koppmann, R., Eck, T.F. and Eleuterio, D.P.: A review of biomass burning emissions part II: intensive physical properties of biomass burning particles, Atmos. Chem. Phys., 5(3), 799–825, doi:10.5194/acp-5-799-2005, 2005.

Reynolds, R.L., Goldstein, H.L., Moskowitz, B.M., Kokaly, R.F., Munson, S.M., Solheid, P., Breit, G.N., Lawrence, C.R., and

580   Derry, J.: Dust deposited on snow cover in the San Juan Mountains, Colorado, 2011– 2016: Compositional variability bearing on snow-melt effects, J. Geophys. Res. Atmos., 125, e2019JD032210, https://doi-org.lib-e2.lib.ttu.edu/10.1029/2019JD032210, 2020.

Rublee, C., Sorensen, C., Lemery, J., Wade, T., Sams, E., Hilborn, E., and Crooks, J.: Associations between dust storms and intensive care unit admissions in the United States, 2000–2015, GeoHealth, 4, https://dx.doi.org/10.1029%2F2020GH000260,

585   2020.

Ryder, C.L., Highwood, E.J., Walser, A., Seibert, P., Philipp, A. and Weinzierl, B.: Coarse and Giant Particles are Ubiquitous in Saharan Dust Export Regions and are Radiatively Significant over the Sahara, Atmos. Chem. Phys. Discuss., 1–36, doi:10.5194/acp-2019-421, 2019.

Saraga, D., Maggos, T., Sadoun, E., Fthenou, E., Hassan, H., Tsiouri, V., Karavoltsos, S., Sakellari, A., Vasilakos, C., and
590   Kakosimos, K.: Chemical Characterization of Indoor and Outdoor Particulate Matter (PM2.5, PM10) in Doha, Qatar, Aerosol
       Air Qual. Res., 17, 1156-1168, https://doi.org/10.4209/aaqr.2016.05.0198, 2017.

Sarkar, S., Chauhan, A., Kumar, R., and Singh, R.P.: Impact of deadly dust storms (May 2018) on air quality, meteorological,
and atmospheric parameters over the northern parts of India, GeoHealth, 3, 67–80, https://doi.org/10.1029/2018GH000170,
2019.

595   Stout, J.: Dust and environment in the Southern High Plains of North America, J. Arid Environ., 47, 425-441,
       https://doi.org/10.1006/jare.2000.0732, 2001.

Textor, C., Schulz, M., Guibert, S., Kinne, S., Balkanski, Y., Bauer, S., Berntsen, T., Berglen, T., Boucher, O., Chin, M.,
Dentener, F., Diehl, T., Easter, R., Feichter, H., Fillmore, D., Ghan, S., Ginoux, P., Gong, S., Grini, A., Hendricks, J., Horowitz,
L., Huang, P., Isaksen, I., Iversen, I., Kloster, S., Koch, D., Kirkevåg, A., Kristjansson, J.E., Krol, M., Lauer, A., Lamarque,
600   J.F., Liu, X., Montanaro, V., Myhre, G., Penner, J., Pitari, G., Reddy, S., Seland, Ø., Stier, P., Takemura, T., and Tie, X.:
       Analysis and quantification of the diversities of aerosol life cycles within AeroCom, Atmos. Chem. Phys., 6, 1777–1813,
       https://doi.org/10.5194/acp-6-1777-2006, 2006.

Tobias, A., Karanasiou, A., Amato, F., Roqué, M., and Querol, X.: Health effects of desert dust and sand storms: a systematic
review and metaanalysis protocol, BMJ Open 2019 9, e029876. https://doi.org/10.1136/bmjopen-2019-029876, 2019.

605   Tong, D.Q., Wang, J.X.L., Gill, T.E., Lei, H., and Wang, B.: Intensified dust storm activity and Valley fever infection in the
       southwestern United States, Geophys. Res. Lett., 44, 4304–4312, doi:10.1002/2017GL073524, 2017.

Toure, N.O., Gueye, A., Mbow-Diokhane, G.S., Jenkins, M., Li, M.S., Drame, K.A., Coker, R., and Thiam, K.: Observed and
modeled seasonal air quality and respiratory health in Senegal during 2015 and 2016, Geo Health, 3, 423–442,
https://doi.org/10.1029/2019GH000214, 2019.

610   Tozer, P., and Leys, J.: Dust storms - what do they really cost?, Rangel. J., 35 (2), 131–142, https://doi.org/10.1071/RJ12085,
       2013.

Tsai, J.H., Huang, K.L., Lin, N.H., Chen, S.J., Lin, T.C., Chen, S.C., Lin, C.C., Hsu, S.C. and Lin, W.Y.: Influence of an Asian
Dust Storm and Southeast Asian Biomass Burning on the Characteristics of Seashore Atmospheric Aerosols in Southern
Taiwan, Aerosol Air Qual. Res., 12, 1105-1115, https://doi.org/10.4209/aaqr.2012.07.0201, 2012.

615   TSI: DUSTTRAK™ DRX Aerosol Monitor Model 8533/8534/8533EP, Operation and Service Manual, P/N 6001898 Revision
       S, https://www.tsi.com/getmedia/3699890e-4adf-452f-9029-f3725612d5d1/8533-8534-DustTrak_DRX-6001898-Manual-
       US?ext=.pdf, last access: 4 June 2021.

van der Does, M., Knippertz, P., Zschenderlein, P., Giles Harrison, R. and Stuut, J.B.W.: The mysterious long-range transport
of giant mineral dust particles, Sci. Adv., 4(12), eaau2768, doi:10.1126/sciadv.aau2768, 2018.

620   Wang, C., Jeong, G.R., and Mahowald, N.: Particulate absorption of solar radiation: anthropogenic aerosols vs. dust, Atmos.
       Chem. Phys., 9, 3935–3945, https://doi.org/10.5194/acp-9-3935-2009, 2009.

Weinzierl, B., Ansmann, A., Prospero, J.M., Althausen, D., Benker, N., Chouza, F., Dollner, M., Farrell, D., Fomba, W.K., Freudenthaler, V., Gasteiger, J., Groß, S., Haarig, M., Heinold, B., Kandler, K., Kristensen, T.B., Mayol-Bracero, O.L., Müller, T., Reitebuch, O., Sauer, D., Schäfler, A., Schepanski, K., Spanu, A., Tegen, I., Toledano, C. and Walser, A.: The Saharan aerosol long-range transport and aerosol-cloud-interaction experiment: Overview and selected highlights, Bull. Am. Meteorol. Soc., 98(7), 1427–1451, doi:10.1175/BAMS-D-15-00142.1, 2017.

WHO (World Health Organization): WHO Air Quality Guidelines for Particulate Matter, Ozone, Nitrogen Dioxide and Sulfur Dioxide, Global Update 2005. Summary of Risk Assessment. WHO, Geneva, 2006.

WHO (global air quality guidelines): particulate matter ($PM_{2.5}$ and PM_{10}), ozone, nitrogen dioxide, sulfur dioxide and carbon monoxide. World Health Organization. https://apps.who.int/iris/handle/10665/345329. License: CC BY-NC-SA 3.0 IGO, 2021.

Zhang, X., Zhao, L., Tong, D.Q., Wu, G., Dan, M., and Teng, B.: A systematic review of global desert dust and associated human health effects. Atmosphere, 7(12), 158, https://doi.org/10.3390/atmos7120158, 2016.