Impact of deoxygenation and warming on global marine species in the 21st century

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Abstract. Ocean temperature and dissolved oxygen shape marine habitats in interplay with species’ physiological characteristics. Therefore, the observed and projected warming and deoxygenation in the 21st century of the world’s oceans may strongly affect species’ habitats. Here, we implement an extended version of the Aerobic Growth Index (AGI), which quantifies whether a viable population of a species can be sustained in a particular location. We assess the impact of projected deoxygenation and warming on the contemporary habitat of 47 representative marine species covering the epipelagic, mesopelagic/bathypelagic, and demersal realms. AGI is calculated for these species for the historical period and into the 21st century using bias-corrected environmental data from six comprehensive Earth System Models. While habitat viability decreases nearly everywhere with global warming, impact of this decrease is strongly species-dependent. Most species lose less than 5% of their contemporary habitat volume over the 21st century even at 3°C of global warming relative to preindustrial, although some individual species are projected to incur losses 2-3 times greater than that. We find that the contemporary spatiotemporal variability of O2 and temperature (and hence AGI) provides a quantifiable measure of a species’ vulnerability to change. Species’ vulnerability is the most important indicator for large (>5%) potential habitat losses - not relative or absolute changes in habitat viability (i.e., AGI\text{rel} or ΔAGI), temperature or O2. Loss of contemporary habitat is for most epipelagic species driven by warming of ocean water and is therefore expanded with increased levels of global warming. In the mesopelagic/bathypelagic and demersal realms habitat loss is also affected by pO2 decrease for some species. Our analysis is constrained by the uncertainties involved in species-specific critical thresholds, which we quantify, by data limitations on 3D species distributions as well as by high uncertainty in model O2 projections in equatorial regions. Focus on these topics in future research will strengthen our confidence in assessing climate-change driven losses of contemporary habitat across the global oceans.

1 Introduction

Ocean temperature and dissolved oxygen (O2) are strongly linked by physical and biogeochemical processes as well as through their effects on the aerobic performance of aquatic ectotherms (Pörtner, 2010; Verberk et al., 2011; Breitburg et al., 2018; Oschlies et al., 2018; Seibel et al., 2021). Indeed, temperature and O2 are both found to be central in shaping species’
distributions and are important climatic stressors to marine species worldwide (Perry et al., 2005; Doney et al., 2011; Poloczanska et al., 2013; Breitburg et al., 2018; Penn et al., 2018; Deutsch et al., 2020; Clarke et al., 2021). Observed and projected warming and deoxygenation are therefore likely to impact species. Robust observational evidence of anthropogenically-forced deoxygenation is now becoming available as long-term O$_2$ changes emerge from their natural variability (Frölicher et al., 2009; Long et al., 2016; Stramma et al., 2020; Buchanan and Tagliabue, 2021; Sharp et al., 2022). Specifically, an increase in the temporal and spatial resolution of observational data has allowed for the discovery and quantification of a ~2% decline in the global top-1000m O$_2$ inventory since the 1960s (Ito et al., 2017; Oschlies et al., 2017; Schmidtko et al., 2017; Breitburg et al., 2018). This negative trend is projected to continue during the 21st century for all climate scenarios (Bopp et al., 2013; Kwiatkowski et al., 2020). More than 10% of deep ocean O$_2$ is likely to be lost even if CO$_2$ emissions were stopped in the year 2020 (Oschlies, 2021). Earth System Model simulations project an O$_2$ loss between 100-600 meter depth of 13.27±5.28 mmol m$^{-3}$ for a high-emission low-mitigation SSP5-8.5 scenario and 6.36±2.92 mmol m$^{-3}$ loss for a high-emission low-mitigation SSP1-2.6 scenario by the end of the 21st century (2080–2099 mean values relative to 1870–1899; Bopp et al., 2013; IPCC, 2019; Kwiatkowski et al., 2020). However, simulated trends seem to underestimate trends in observations (Andrews et al., 2013; Oschlies et al., 2017; Oschlies et al., 2018; Buchanan and Tagliabue, 2021) and models poorly represent tropical Pacific Oxygen Minimum Zones (Cocco et al., 2013; Cabré et al., 2015), indicating even higher possible impacts of deoxygenation toward the future. Ocean temperatures are changing as well: Ocean warming is a direct effect of atmospheric warming, as the ocean takes up approximately 90% of the excess heat from anthropogenic activities (Von Schuckmann et al., 2020). Global mean sea surface temperatures are observed to have increased by ~0.5 °C since the 1960s (Hersbach et al., 2020). Earth System Model simulations project sea surface warming of 3.5±0.8°C for a high-emission low-mitigation SSP5-8.5 scenario and 1.42±0.32°C warming for a low-emission high-mitigation SSP1-2.6 scenario by the end of the 21st century (2080–2099 mean values relative to 1870–1899; Kwiatkowski et al., 2020). Most marine species will thus experience further warming of their habitat, considering that chances of limiting global atmospheric warming to 1.5°C are low even if all and unconditional pledges by countries are implemented in full and on time (IPCC, 2021; Meinshausen et al., 2022). Models that include more complex representations of species biology and ecology show that every tenth of degree of global warming increases impacts on marine biodiversity, transforming species assemblages through changes in abundance, biomass and catch potential (Cheung et al., 2016). Moreover, the warming signal penetrates the deep ocean where it has major potential to affect marine ecosystems together with deoxygenation and ocean acidification (Levin and Le Bris, 2015).

From a biogeochemical perspective, changes in O$_2$ content of a water parcel are driven by a combination of a) changes in O$_2$ solubility due to changes in temperature, b) changes in ventilation and stratification of the water column and associated changes in O$_2$ supply, c) changes in pO$_2$ due to gas diffusion rates that depend on temperature, and d) changes in large-scale biological consumption of O$_2$ (Keeling et al., 2010; Kwiatkowski et al., 2020; Buchanan and Tagliabue, 2021; Oschlies, 2021; Pitcher et al., 2021). The relative importance of these mechanisms for deoxygenation varies in space and time (e.g., Frölicher et al., 2009; Keeling et al., 2010; Frölicher et al., 2020), which makes it challenging to attribute local deoxygenation to a single driver.
Generally however, O2 has reduced over the past ~60 years due to a combination of a-c in the extratropical oceans, while changes in large-scale biological consumption of O2 (d) also contributed to changes in O2 in low-O2 equatorial zones (Buchanan and Tagliabue, 2021; Oschlies, 2021). Solubility effects dominate the top-200 meter deoxygenation while biological processes and especially ventilation changes increase their importance with depth (Schmidtko et al., 2017).

Consequences of the observed and projected deoxygenation and warming for marine species can be understood from the biogeochemical and physiological balance between \( pO_2 \) supply and demand, both of which depend on temperature (Pörtner and Knust, 2007; Verberk et al., 2011; Deutsch et al., 2015; Deutsch et al., 2020; Clarke et al., 2021). \( pO_2 \) supply to a water parcel and hence a species is governed by a-d, while a species’ O2 demand (respiration) increases with temperature (Sect. 2.1). This study uses the Aerobic Growth Index (AGI; Clarke et al., 2021) to quantify the combined effects of deoxygenation and warming on marine species in the 21\textsuperscript{st} century. AGI is a species-specific ratio between \( pO_2 \) supply and demand and is interpreted as a measure of habitat viability. Viable habitat is characterized by a \( pO_2 \) supply over \( pO_2 \) demand ratio (i.e., AGI) sufficient for feeding, movement, defense, as well as growth and thus allows for sustainably maintaining a certain species’ population (Clarke et al., 2021). Considering the ongoing deoxygenation and warming, AGI and habitat viability are expected to decrease (Deutsch et al., 2015; Clarke et al., 2021; Gruber et al., 2021; Oschlies, 2021). Our approach newly includes depth variability as well as temporal variability of temperature and O2 in calculating AGI (Sect. 2.1), which we apply to 47 representative species thanks to the generalized temperature dependence of \( pO_2 \) demand in AGI (Sect. 2.2). For environmental data of temperature and \( pO_2 \) we use bias corrected Earth System Model projections from the latest version of the Coupled Model Intercomparison Project Phase 6 (CMIP6) (Sect. 2.3). Through this approach we assess the potential loss of viable contemporary habitat volume due to warming and deoxygenation for a representative selection of species - as well as identifying the drivers of such losses (Sect. 3).

### 2 Methods and data

#### 2.1 Aerobic Growth Index

We apply the Aerobic Growth Index (AGI; Clarke et al., 2021) to quantify species-specific impacts on habitat viability in response to changes in temperature and \( pO_2 \). AGI integrates growth theory, metabolic theory, and biogeography to calculate a theoretical ratio of \( pO_2 \) supply over \( pO_2 \) demand for each species \( i \) (Eq. 1; rewritten from Eq. 14 in Clarke et al., 2021).

\[
AGI_i = \frac{pO_{2,\text{supply}}}{pO_{2,\text{demand}}} = \frac{pO_{2,\text{supply}}}{pO_{2,\text{threshold}}} W^{\frac{1}{2.4}} \left[ \frac{1}{\int_0^\infty \exp \left( \frac{E_i}{2T} \right) - \frac{E_i}{2T} \right] \quad (1)
\]

Here, the environmental state is described by \( pO_{2,\text{supply}} \) (mbar) and \( T \) as in-situ temperature (K). The variables \( j_1 \) (the anabolism activation energy divided by the Boltzmann constant, 4500K), \( j_2 \) (the catabolism activation energy divided by the Boltzmann constant, 8000K), and \( d \) (the metabolic scaling coefficient, 0.7) are species-independent to facilitate its application to a large number of species (Pauly, 2010; Cheung et al., 2011; Pauly and Cheung, 2018; Clarke et al., 2021). We note that the ratio
between the mean species’ weight $W_i$ (g) and the species’ asymptotic weight $W_{\infty,i}$ (g), $\frac{W_i}{W_{\infty,i}}$ reduces to $\frac{1}{3}$ as $W_i = \frac{1}{3} \cdot W_{\infty,i}$ (Clarke et al., 2021).

We newly consider both vertical and temporal variability in $pO_2$ and temperature in the calculation of the species’ $pO_2$ threshold ($pO_2^{\text{threshold}}$, mbar) and preferred temperature ($T^{\text{pref}}$, K). Critical $pO_2$ values as well as preferred temperatures are highly species-dependent (Vaquer-Sunyer and Duarte, 2008; Pörtner and Peck, 2011; Seibel, 2011). Following Clarke et al. (2021) and Penn et al. (2018), we take $pO_2^{\text{threshold}}$ ($T^{\text{pref}}$) as the volume-weighted 10th (50th) percentile of all in-habitat $pO_2$ (temperature) values. AGI can therefore be calculated for any species for which we have distribution data as well as environmental data for temperature and $pO_2$. Temporal variations in $pO_2$ and temperature are considered by using monthly climatological mean data from the World Ocean Atlas 2018 (WOA18 average of all available decades; Boyer et al., 2018; Garcia et al., 2019; Locarnini et al., 2019; Zweng et al., 2019). The horizontal distribution data are extended over the full depth range of each species (0-200m for epipelagic species, 200-1000m for mesopelagic and bathypelagic species and bottom layer of the ocean data for the demersal species, thereby covering both deep and shallow demersal habitats; see Fig. C1 and Sect. 2.3) to include the vertical variability of $O_2$ and temperature in our estimate of $pO_2^{\text{threshold}}$, $T^{\text{pref}}$ and hence AGI. This approach facilitates the estimation of species-specific, temperature-dependent critical $pO_2$ levels and $T^{\text{pref}}$ despite the lack of observational data from multi-stressor laboratory experiments that apply to field conditions (Boyd et al., 2018; Collins et al., 2022). Different iterations of the metabolic index (Deutsch et al., 2015; Penn et al., 2018; Deutsch et al., 2020), which require laboratory-based estimates of temperature dependent critical $pO_2$ levels, agree well with AGI in their assessment of habitat loss (Clarke et al., 2021) despite the much fewer data needed to calculate AGI. For additional details on the calculation of AGI we refer to Clarke et al. (2021).

An AGI of one implies that there is sufficient $O_2$ supply for feeding, movement, and defence, but not growth and reproduction. To sustain a viable population, additional aerobic scope is needed until AGI is above its critical value ($AGI^{\text{crit}}$) for a particular species. Following Clarke et al. (2021), $AGI^{\text{crit}}$ is the 10th percentile of all AGI values in a species’ habitat including vertical and temporal variability as done similarly for $pO_2^{\text{threshold}}$ and $T^{\text{pref}}$. In this study a species is deemed impacted by changes in temperature and $pO_2$ whenever AGI drops below $AGI^{\text{crit}}$ on an annual mean basis. All species information is listed in Table A1. The coarse resolution and the imperfect harmonization between the biogeographical, temperature and $O_2$ data may affect the accuracy of the estimated $AGI^{\text{crit}}$, as indicated by some species having $AGI^{\text{crit}}$ at or below 1. We discuss in Sect. 4 how such biases may affect the results and conclusion, and how future studies can build on our results to improve the accuracy of the analysis.

Relative changes in AGI ($AGI^{\text{rel}}$) between time=$t_1$ and time=$t_0$ can be estimated from Eq. 2:

$$AGI^{\text{rel}} = \frac{AGI_{t_1}}{AGI_{t_0}} = \frac{pO_2(t_1)}{pO_2(t_0)} \cdot \exp\left(\left(t_2 - t_1\right) \cdot \left(\frac{1}{T(t_2)} - \frac{1}{T(t_1)}\right)\right) - 1 \quad (2)$$

Relative changes are thus entirely species-independent (in contrast to the metabolic index of Deutsch et al., 2015). We maintain a reference period 1995-2014 throughout this study (i.e., AGI($t_0$) is the mean AGI over the years 1995-2014).
2.2 Species data

AGI can be calculated for nearly a 1000 commercially exploited marine species due to the generalized temperature dependence of the $pO_2^{\text{demand}}$. This broad applicability of AGI allows us to select 47 representative species (n=47), which are chosen such that depth level, climatic zone (tropical and temperate) and body size are represented. In addition, we selected some pelagic and deep-water wide-ranging species that inhabit both tropical and temperate regions, as well as the hypoxia tolerant species Dosidicus gigas. Three depth groups are represented through our selection: Epipelagic species (n=23) in the 0-200 meter depth range, mesopelagic/bathypelagic species (n=6) in the 200-1000 meter depth range and demersal species (n=18) at the sea floor (bottom wet layer of the models; Sect 2.3). The representativeness of our species selection is assessed in Sect. 3.4. Species’ $pO_2^{\text{threshold}}$, $T^{\text{prev}}$ and $AGI^{\text{est}}$ are listed in Table A1. The contemporary species distributions are based on gridded product from Close et al. (2006) and are assumed to represent the 1995-2014 mean state (Fig. C1).

2.3 Earth system model data

Environmental data of O$_2$, potential temperature and salinity for the years 1850-2100 are taken from the Coupled Model Intercomparison Project Phase 6 (CMIP6) multi-model ensemble (Eyring et al., 2016). Scenarios ‘historical’, ‘SSP5-8.5’ and ‘SSP1-2.6’ as well as the pre-industrial control simulation ‘piControl’ were used from the six models for which all data were available: CNRM-ESM2-1, MPI-ESM1-2-HR, UKESM1-0-LL, IPSL-CM6A-LR, GFDL-ESM4, and CanESM5 (Appendix Table A2). All model data were horizontally regridded to a 1° regular grid before further post-processing.

To account for mean errors and model drift, both a drift correction and bias correction were performed. First, the bottom layer (‘seafloor’) O$_2$ ocean data were linearly detrended for piControl drift over the piControl years corresponding to the scenario years (1850 to 2100) as these trends are more than 10% of the scenario signal for some models. Drift in bottom layer temperature and salinity as well as upper water column O$_2$, salinity and temperature were negligible as compared to the scenario signals and therefore not accounted for. Secondly, we performed a mean bias correction (e.g., Maraun, 2016) on all model data by subtracting the ‘present-day’ monthly mean climatological bias (the difference between WOA18 data vertically interpolated to the respective model levels and the respective model’s 1995-2014 monthly climatology) from the entire simulated timeseries. We used the available WOA18 climatological mean product for the sea floor data because WOA18 climatologies are only available at monthly resolution until 1500m depth. The extracted original spatial resolution of the WOA18 data is 1° for O$_2$ and 0.25° for temperature and salinity but these were all regridded to 1° to match the regridded model data grid. Note that to calculate the temperature biases, the model potential temperature was converted to in-situ temperature.

Finally, $pO_2$ was calculated following Sect. E by Bittig et al. (2018), which is based on earlier work (Benson and Krause, 1984; Garcia and Gordon, 1992; Sarmiento and Gruber, 2006), and includes pressure correction (Taylor, 1978) and the correction for water vapor pressure (Weiss and Price, 1980) in the calculation of $pO_2$ (Appendix B).

All results are presented at global warming levels (e.g., Hausfather et al., 2022). In order to do so we first bias-correct modeled surface air temperatures such that the 1995-2014 global mean air temperature increase since 1850-1900 is 0.87 °C as observed
(HadCRUT.5.0.1.0; Morice et al., 2021) in order to be consistent with the bias corrected ocean temperature and oxygen data. We then find the first year where the 15-year running mean of these bias-corrected global mean surface air temperature data is greater than or equal to the warming level of interest and calculate the 15-year running over the data for the analyses.

3 Results

3.1 Global changes in warming, deoxygenation and habitat viability

Ocean temperature is projected to increase (Fig. 1a) and $pO_2$ to decrease (Fig. 1b) with global warming of the atmosphere. These changes occur in all three depth layers considered here, and for all CMIP6 models. Noticeably, the response of ocean warming and deoxygenation to global atmospheric warming is approximately linear (Fig. 1). From a linear fit to the mean CMIP6 changes in Fig. 1 we find that the epipelagic realm warms the most with 0.50±0.03 °C per degree of atmospheric warming (standard deviation given across the individual model fits). This signal is dampened toward depth to 0.18±0.02 °C per degree of global warming in the mesopelagic/bathypelagic realm and 0.08±0.01 °C per degree of global warming in the demersal realm (Fig. 1a). In addition to the warming, we find that the epipelagic realm loses 0.40±0.55 mbar of $pO_2$, the mesopelagic/bathypelagic loses 1.35±0.89 mbar of $pO_2$ and the demersal realm loses 1.17±0.97 mbar of $pO_2$ on average and per degree global warming of the atmosphere (Fig. 1b) hence having the largest changes realized at depth in contrast to ocean warming. The warming (Fig. 1a) and deoxygenation (Fig. 1b) drive a reduction in AGI$^{rel}$ which we interpret as a loss of habitat viability (Sect. 2.1; Fig. 1c) that is independent of species (Eq. 2). In the epipelagic, AGI$^{rel}$ is reduced by 2.17±0.69 % per degree of global warming (Fig. 1c). The global mean reduction in habitat viability (i.e., AGI$^{rel}$) in the mesopelagic/bathypelagic realm is 2.33±1.64 % per degree of global warming. Last, demersal loss of habitat viability is 0.86±0.48 % per degree of global warming is the least pronounced of the three studied depth intervals. The approximately linear response of marine warming, deoxygenation, and loss of habitat viability to global atmospheric surface warming (Fig. 1) highlights and confirms that all additionally realized atmospheric warming will affect the marine environment (Cheung et al., 2016).

The projected changes are independent of emission pathway and only depend on the amount of global warming to a first degree. Nevertheless, a sharp contrast exists between maximum changes realized in scenario SSP1-2.6 and those changes of scenario SSP5-8.5: Relative to the 1850-1900 mean, global multi-model mean warming by 2081-2100 in SSP1-2.6 is limited to 1.14±0.28 °C in the epipelagic, 0.62±0.07 °C in the mesopelagic/bathypelagic and 0.22±0.02 °C in the demersal realm. For SSP5-8.5 the changes are approximately doubled to 2.70±0.76 °C, 1.00±0.15 °C and 0.40±0.06 °C of warming, respectively. Deoxygenation is also much reduced in the low-emission scenario as compared to the high-emission SSP5-8.5 scenario by 2081-2100, although model uncertainty is larger here: Global mean $pO_2$ is reduced by at most 0.88±1.42 mbar in the epipelagic, 3.23±2.96 mbar in the mesopelagic/bathypelagic, and 5.42±3.80 mbar in the demersal realm for SSP1-2.6 while maximum global mean deoxygenation is projected to be stronger in SSP5-8.5 with 2.27±2.85 mbar, 7.13±4.22 mbar and 5.61±4.23 mbar $pO_2$ loss respectively. The loss of habitat viability is 6.39 % lower in SSP1-2.6 than in SSP5-8.5 by 2081-2100 in the epipelagic
(-11.58±4.48 % under SSP5-8.5 vs. -5.18±2.08 % under SSP1-2.6), 4.62 % lower in the mesopelagic/bathypelagic (-11.48±7.50 % under SSP5-8.5 vs. -6.86±5.60 % under SSP1-2.6), and 0.90 % lower in the demersal realm (-4.23±1.90 % under SSP5-8.5 vs. -3.33±1.69 % under SSP1-2.6).

Figure 1 Global mean changes in ocean temperature (a), pO₂ (b) and AGI_rel (c) for different global warming levels, where global warming is calculated as global surface air temperature increase relative to the 1850-1900 mean. The 20-year multi-model running mean is given in opaque blue (SSP1-2.6) and red (SSP5-8.5) and has for several decades the corresponding multi-model mean year labeled. Individual models are shown in transparent blue and red without taking a running mean. AGI_rel is given relative to the 1995-2014 mean as in the remainder of the manuscript, and the mean contribution from temperature only (excluding the small effect of temperature on pO₂) is indicated by the black line in panel (c). AGI_rel (%) is entirely species-independent (Eq. 2) and values that exceed 1000% or are below -1000% were excluded during the calculation of the global mean AGI changes to omit several grid-cells with extreme outliers caused by very small absolute changes in O₂ causing very large changes in AGI_rel.

3.2 Local changes and drivers of habitat viability

A relative reduction in habitat viability (i.e., a reduction in AGI_rel; Fig. 1c) is projected to occur almost everywhere (Fig. 2a-c), indicating that for most habitats and therefore species we expect a reduction in habitat viability. The reduction in habitat viability is generally larger in the epipelagic and mesopelagic/bathypelagic realms (Fig. 1c, 2a, b), but the larger spatial heterogeneity at mesopelagic/bathypelagic depths reveals that locally mesopelagic/bathypelagic AGI_rel reductions may far exceed those in the epipelagic, particularly in the North Pacific (Fig. 2b). Hence the location of a species’ habitat, both vertically and horizontally, is key to projected changes in habitat viability.

When considering the contribution from the two drivers of AGI change, pO₂ and temperature changes, AGI_rel is driven mostly by temperature in the epipelagic and by pO₂ in the mesopelagic/bathypelagic and demersal realms (Fig. 2d-i). The AGI_rel due to pO₂ is -0.16±5.12% for the epipelagic, -2.52±6.96% for the mesopelagic/bathypelagic, and -0.62±2.02% for the demersal realm (Fig. 2d-f), while the respective AGI_rel due to temperature are -2.32±1.36%, -0.91±1.18%, and -0.39±0.95% (Fig. 2g-i). In the mesopelagic/bathypelagic the drivers of loss in habitat viability depend more strongly on location (Fig. 2e,h). Globally, on average 87% of AGI_rel is driven by the relatively pronounced warming in the epipelagic (black lines in Fig. 1c) since changes in pO₂ are minor (Fig. 1b) because the epipelagic realm is generally well ventilated with O₂-rich surface waters. For the mesopelagic/bathypelagic (demersal), AGI_rel is for only 24% (37%) explained by warming (black lines in Fig. 1c).
Nevertheless, the larger contribution from \( pO_2 \) to AGI\(^{rel} \) increases uncertainty for the mesopelagic/bathypelagic and demersal realms because model projections are uncertain for \( pO_2 \) (Fig. 1b, C2).

Figure 2 Multi-model mean AGI\(^{rel} \) relative to the 1995-2014 mean at 2 °C global warming (using SSP5-8.5 simulations), vertically averaged over the epipelagic and mesopelagic/bathypelagic realms, and shown for the demersal realm (a-c). AGI\(^{rel} \) is split up into the contribution from (d-f) \( pO_2 \) and (g-i) temperature. Data are hatched where more than 2 out of the 6 models disagree about the sign of change. Note that sea floor depth and thus demersal depth depends strongly on location. Contributions from \( pO_2 \) (temperature) are calculated by keeping temperature (\( pO_2 \)) constant at its 1995-2014 mean state when calculating AGI\(^{rel} \). Further note that since [\( O_2 \)] depends on temperature too, the contribution to AGI\(^{rel} \) from \( pO_2 \) also contains a minor temperature component.

AGI\(^{rel} \) is uncertain for species having a large part of their habitat in eastern-boundary equatorial upwelling regions or around Antarctica at epipelagic depths, the western equatorial Pacific at mesopelagic/bathypelagic depths, north of the equator in the Indian Ocean at epipelagic and mesopelagic/bathypelagic depths or regions scattered across all ocean basins for demersal depths (hatched areas in Fig. 2a-c). Most of this uncertainty is coming from \( pO_2 \) (Fig. 2 d-f, Fig. C2), with temperature contributing to uncertainty in the North-Atlantic south of Greenland and in the western equatorial Pacific at mesopelagic/bathypelagic depths.

Exceptions to the decrease in AGI\(^{rel} \) are limited to some smaller parts of the world’s oceans including equatorial regions and the North-Atlantic south of Greenland in the epi- and mesopelagic/bathypelagic, and around the Antarctic continent in the epipelagic. Model disagreement is generally large in these regions of AGI\(^{rel} \) increase and is mostly attributable to projected increases in \( pO_2 \) which have large uncertainties (hatching in Fig. 2a-f and model range in Fig. C2).
Notably, in some regions the effects of \( pO_2 \) and temperature on AGI\textsubscript{rel} may compensate each other and result in negligible changes in AGI: We find examples of this in the Northern Indian Ocean at epipelagic depths, in the Gulf of Guinea at mesopelagic/bathypelagic depths, and in the North Atlantic around Iceland at demersal depths. Note that the patterns in each of the panels of Fig. 2 remain similar for higher degrees of global warming, only the intensity of change increases (not shown), which agrees with the approximately linear response of AGI\textsubscript{rel} to global warming (Fig. 1c).

### 3.3 Impacts of AGI changes on habitat volume of individual species

The overall decrease in AGI\textsubscript{rel} and hence habitat viability with global warming (Figs. 1c and 2) causes loss of contemporary habitat volume and hence local extinction for species at each of the studied depth ranges (Figs. 3 and 4). Habitat loss is expressed relative to its contemporary volume (Fig. 3) to facilitate comparison between wide-ranging and more narrowly distributed species. Loss of contemporary habitat is generally less than \(~10\%\) at 1.5°C global warming but increases to up to \(~25\%\) for individual species at 3°C (Fig. 3). Wide-ranging epipelagic species (e.g., Acanthocybium solandri, Coryphaena hippurus, Katsuwonus pelamis, Thunnus obesus, or Elagatis bipinnulata: Fig. C1) experience losses of contemporary habitat volume of less than 5\% for any of the analyzed warming levels, while more narrowly distributed species experience the largest losses of up to \(~25\%\) of their contemporary habitat at 3°C global warming (e.g., Micromesistius poutassou, Thunnus atlanticus, Sebastes mentella, or Anarhichas denticulatus). Notably, species that have the largest contemporary habitat loss at 1.5°C generally are those species that also lose the most at 3°C of global warming, which is in line with the earlier findings of approximately linear response of habitat viability to warming and deoxygenation (Fig. 1). Any early (i.e., 1.5°C) response of a species to warming and deoxygenation is therefore a warning indicator for additional loss of contemporary habitat at increased levels of global warming. We separately assess the impact of the uncertainty of AGI\textsubscript{crit} on these results by calculating habitat loss with an AGI\textsubscript{crit} of a) minimum AGI, b) 5\(^{th}\) percentile, c) 10\(^{th}\) percentile (i.e., the default case), d) 15\(^{th}\) percentile and e) 20\(^{th}\) percentile of in-habitat AGI. We find that even when including much higher thresholds (AGI\textsubscript{crit} as 20\(^{th}\) percentile), our results are similar with a few species having large losses but most losing less than 5\% at 2°C of warming relative to the 1995-2014 state (Fig. C3).

Absolute losses in habitat volume show that small relative losses (Fig. 3) often correspond to the largest volumetric losses (Fig. 4). As an example, median Thunnus alalunga habitat loss is less than \(~2.5\%\) at any of the analyzed warming levels (Fig. 3), while absolute losses are the largest of all 47 species at \(\sim 0.25-1.5 \times 10^6\) km\(^3\) - depending on global warming level (Fig. 4). On the other hand, we find species like Sebastes mentella for which relative losses are large (median \(\sim 8-26\%\) of the contemporary habitat depending on global warming level; Fig. 3) while absolute losses are comparably small (\(\sim 0.6-1.8 \times 10^3\) km\(^3\)) because the contemporary volume of Sebastes mentella is relatively small (Fig. 4). Note that epipelagic species lose habitat volume in the order of a million km\(^3\): In comparison, the entire Black Sea has a volume of about 0.5 million km\(^3\). Depending on the location of viable contemporary habitat loss, for species of commercial interest such large absolute loss can be particularly impactful to local fisheries.
Figure 3 Percentage of remaining contemporary (1995-2014) habitat volume for different levels of global warming. Note the different y-axis scale for 3°C global warming. Habitat volume is considered lost when AGI<AGI_{crit} on an annual mean basis. For 1.5°C global warming both the SSP1-2.6 and SSP5-8.5 scenarios are included (number of datapoints \(n=2\) scenarios \(\times\) 6 models = 12 for each boxplot), while at higher levels of global warming we use SSP5-8.5 as not all models reach these warming levels under the SSP1-2.6 scenario (\(n=6\) models). The species are ordered such that species with the largest median losses at 1.5°C global warming are on the left-hand side for each realm subplot. Each boxplot indicates the median in orange and a box bounded by the interquartile range (IQR; the 25th to 75th percentiles) and the whiskers extending to the data range with a maximum of 1.5×IQR, with outliers as open circles. Stars indicate the median contribution from temperature, the remainder is therefore due to \(pO_2\) changes. As changes are expressed relative to the contemporary viable habitat volume (which is by definition 90% of the total habitat volume), values up to 111% (=100/90) are possible.
Contemporary habitat loss (km$^3$) for different levels of global warming. Note the different y-axes for both the depth groups and warming levels. Habitat volume is considered lost when AGI$>$ AGI$^{cr}$. For 1.5°C global warming both the SSP1-2.6 and SSP5-8.5 scenarios are included (number of datapoints $n=2$ scenarios * 6 models = 12 for each boxplot), while at higher levels of global warming we use SSP5-8.5 as not all models reach these warming levels under the SSP1-2.6 scenario ($n=6$ models). Species are ordered in the same order as in Fig. 3. Each boxplot indicates the median in orange and a box bounded by the interquartile range (IQR; the 25th to 75th percentiles) and the whiskers extending to the data range with a maximum of 1.5×IQR, with outliers as open circles.

For most species, temperature is the main driver of habitat loss (black stars in Fig. 3). Exceptions exist for example in the mesopelagic/bathypelagic, where $pO_2$ drives about half of the habitat loss for the two species with the largest loss (i.e., Sebastes mentella and Aphanopus carbo) as well as for the demersal species Anarhichas denticulatus. Even though most of the realized
loss can be explained by warming, not all species have large losses despite warming being relatively uniform although dampened toward depth (Fig. 1a; Kwiatkowski et al., 2020). We understand these differences from the original spatial and temporal $pO_2$ and temperature variability in each species’ habitat, which shapes their vulnerability to change (Sect. 3.4).

### 3.4 Drivers of habitat volume loss of individual species

The differences in habitat loss between species as shown in Figs. 3 or 4 are better understood from the probability density of AGI for each species (Fig. 5 and C4). The spatial variability of $pO_2$ and temperature in each species’ habitat results in a species-specific probability density function (PDF) for AGI (Fig. 5a,b). Depending on this shape, a given reduction in AGI exposes a relatively large or small part of the species’ habitat to subcritical AGI values (Fig. 5a,b). We can quantify this “vulnerability” of a species to changes in AGI by calculating the cumulative sum of the PDFs (i.e., the cumulative density function, CDF, Figs. 5c,d and C5).

![Figure 5](https://doi.org/10.5194/egusphere-2022-1364)

**Figure 5** Conceptual figure based on *Thunnus atlanticus* (a,c) and *Thunnus obesus* (b,d) showing the difference in impact (change in volume $\Delta V$) of a mean reduction $\Delta AGI$ of 0.1. This difference is shown to be related to the shape of the PDF and the slope of the CDF at 0.1 (i.e., at $AGI_{\text{crit}}$), which we refer to as the species’ “vulnerability”.

The slope of the CDF at a cumulative density of 0.1 (i.e., 10% of the volume where AGI$_{\text{crit}}$ is defined) indicates the potential loss in habitat for a certain change in AGI (Fig. 5 and C5). If the slope of the CDF is steep at the critical threshold, the species
is relatively vulnerable to warming and deoxygenation: Only a small reduction in habitat viability (i.e., AGI) will push a relatively large volume below the critical threshold. In example is given in Fig. 5, where for an identical change in mean in-habitat ΔAGI of 0.1 just 1% of volume is pushed below AGI\textsuperscript{crit} for a species with a small slope of 0.14 (Fig. 5b,d, ‘Thunnus obesus’ schematic), while the same change results in 9% volume loss for a species with a large slope of 1.67 (Fig. a,c, ‘Thunnus atlanticus’ schematic). Changes in the slope of some species’ CDFs indicates that different vulnerabilities exist for different parts of that species’ habitat. Hence, in habitat areas which are represented by a part of the CDF with a relatively steep slope, a relatively small change in AGI is needed to bring a relatively large volume closer to AGI\textsuperscript{crit}. Nevertheless, only the CDF slope at AGI\textsuperscript{crit} relates directly to viable habitat volume loss as only AGI values below AGI\textsuperscript{crit} are considered to have an impact on habitat volume. Indeed, projected habitat volume loss increases with species’ vulnerability (i.e., the CDF slope at AGI\textsuperscript{crit}; Fig. 6d), as well as to a lesser extent with warming and deoxygenation (Fig. 6b,c). Notably, the largest absolute reductions of mean in-habitat AGI do not indicate those species who lose most contemporary habitat volume (Fig. 6a). On the contrary, the environmental state of the contemporary habitat as captured in the PDFs and thus the slope of the CDFs and vulnerability is the strongest indicator for impact: 87% of the variance in volume loss at 2 °C global warming can be explained by vulnerability (R\textsuperscript{2} of linear fit in Fig. 6d). The correspondent linear equation taken across all depth realms is volume loss (%) = 7.31 * vulnerability – 0.10. Habitat viability thus strongly depends on the variability of temperature and \textit{O}_2 in the habitat of the species as captured by the species’ vulnerability (Figs. C5 and 6d). Therefore, reports of relative losses in habitat viability based on \textit{pO}_2 supply over \textit{pO}_2 demand ratios (e.g., Deutsch et al., 2015; Oschlies, 2021) should not be interpreted as leading to actual reductions of viable habitat for individual species as they do not include species-specific thresholds nor their vulnerabilities.

We highlight three groups of species for further discussion: (1) The most affected species due to high vulnerability despite small ΔAGI (red markers in Fig. 6): Micromesistius poutassou, Thunnus atlanticus, Sebastes mentella, Aphanopus carbo, Anarhichas denticulatus, Melanogrammus aeglefinus and Hippoglossoides platessoides, (2) the resilient species which have low losses despite high ΔAGI due to low vulnerability (blue markers in Fig. 6): Centrolophus niger, Hippoglossus hippoglossus and Theragra chalcogramma, and (3) the vulnerable species who lose <5% of their habitat volume due to small ΔAGI despite relatively high vulnerability (yellow markers in Fig. 6): Clupea harengus, Thunnus maccouii, Neocyttus rhombozal, Epinephelus nigritus, Gadus morhua and Nezumia aequalis. Considering the range captured in Fig. 6a-d we expect that our selection of species is representative of a wide range of marine ectotherms.

Interestingly, species with high vulnerability and loss (red markers in Fig. 6) all have a high \textit{pO}_2\textsuperscript{threshold} above 150 mbar (Table A1, Fig. C6). Thus, even though warming explains most of the loss of contemporary habitat (black stars in Fig. 3), loss is only high for vulnerable species (Fig. 6d) – which in turn all are sensitive to \textit{pO}_2 as evidenced by their \textit{pO}_2\textsuperscript{threshold} above ~150 mbar (Fig. C6). A high sensitivity to \textit{pO}_2 and hence a high \textit{pO}_2\textsuperscript{threshold} is therefore an indicator of vulnerability, although not all species with high \textit{pO}_2\textsuperscript{threshold} are vulnerable (Fig. C6). The high vulnerability for species with a high \textit{pO}_2\textsuperscript{threshold} shows that also species in well-oxygenated regions can be vulnerable to climate change as their natural \textit{pO}_2 range is limited. We further note...
that vulnerability does not depend on the depth realm of a species. Resilient species (blue markers in Fig. 6) have strong spatiotemporal variability of AGI (broad PDF in Fig. C4) such that even large mean changes of AGI (Fig. 6a) do not expose a large volume to subcritical AGI values. Noticeable is that the two species with a PDF skewed to the right (Fig. C4 Hippoglossus hippoglossus and Theragra chalcogramma) are both in this group, while all other species tend to have a left-skewed PDF of AGI values in their habitat. Whether AGI is the right metric for determining habitat viability for these two species need further investigation that goes beyond the scope of this study. The six species with relatively high vulnerability but small habitat volume loss (yellow markers in Fig. 6) experience relatively small AGI changes in their habitats even at 3°C global warming (Fig. C4) thereby preventing large habitat losses.

Figure 6 Multi-model mean in-habitat changes of (a) AGI, (b) temperature, (c) $pO_2$ at 2°C of global warming in SSP5-8.5 and (d) vulnerability (CDF slope at a cumulative density of 0.1 based on 1995-2014 mean data, Fig. C5) plotted against loss of contemporary habitat at 2°C of global warming (model range indicated by error bars) for each species. Species with >5% loss marked in red, more than −0.1 ΔAGI in blue, and volume loss <5% as well as vulnerability >0.3 in yellow. There is no uncertainty in the vulnerability calculation because all models have the same 1995-2014 CDF slope due to the WOA18 bias-correction. From a linear regression to the data which is plotted in dashed grey we find an $R^2$ of 0.0% for (a) which line is therefore not plotted, 18% for (b), 21% for (c) and 87% for (d).

### 4 Discussion

We introduce a new version of AGI that adds vertical temporal variability in the calculation of $pO_2$ threshold, $T_{pre}$, AGI crit - which makes it possible to assess volumetric habitat changes. The original AGI applies and assesses temporal variability in the horizontal (surface or bottom ocean layers for pelagic and demersal species, respectively) direction only (Clarke et al., 2021), as commonly practiced (e.g., Bryndum-Buchholz et al., 2019; Tittensor et al., 2021), i.e., either surface or sea floor data were applied for the calculation of $pO_2$ threshold, $T_{pre}$, AGI crit and hence AGI. To assess the differences between our new approach and the original approach we repeated the analysis as presented in Fig. 3 (Fig. C7), now using surface ocean data only for mesopelagic/bathypelagic and epipelagic species as well as calculating $pO_2$ threshold, $T_{pre}$, AGI crit from the surface monthly mean WOA18 data only. We find that the sensitivity to global warming of all species is higher for the original AGI as compared to our new approach which includes vertical and seasonal variability of temperature and $pO_2$. This is understood from the combination of a) limited spatial variability of surface ocean $pO_2$ as well as temperature, leading to higher $T_{pre}$ and $pO_2$ threshold.
estimates and therefore stronger sensitivity to warming and deoxygenation as compared to our new approach and b) larger AGI\textsuperscript{red} changes closer to the surface. We expect that including temporal and vertical spatial variability in calculating AGI provides a more realistic estimate of the \( pO_2 \) and temperature variability experienced by a species and therefore a better estimate of its sensitivity to warming and deoxygenation. Nevertheless, we acknowledge that further increasing spatiotemporal resolution (e.g., using daily-mean data and including interannual variability) can affect variability (Deser et al., 2009; Baumann et al., 2015) which can affect estimates of \( T^{\text{pref}} \), \( pO_2^{\text{threshold}} \) and \( \text{AGI}^{\text{crit}} \). Unfortunately, no established theory exists yet to decide what temporal variability in environmental parameters best captures species’ \( T^{\text{pref}} \), \( pO_2^{\text{threshold}} \) or \( \text{AGI}^{\text{crit}} \). By considering WOA18 monthly mean climatological data as the basis for our estimates of \( T^{\text{pref}} \), \( pO_2^{\text{threshold}} \) and \( \text{AGI}^{\text{crit}} \), we are consistent with the time resolution of the CMIP6 model data (monthly mean). Further, regarding the choice of the 10\textsuperscript{th} percentile threshold and impact of its uncertainty on our results (Fig. C3), we consider an \( \text{AGI}^{\text{crit}} \) threshold above the 20\textsuperscript{th} percentile of in-habitat AGI values unlikely as then by definition already 20 percent of the habitat would be unsuitable to sustain a viable population of that species. Nevertheless, for species where \( \text{AGI}^{\text{crit}} \) is very close to 1 or even below 1 (Table A1), a higher percentile may be warranted as a meaningful critical value. At the 10\textsuperscript{th} percentile, some uncertainty in the species-specific physiological parameters is considered. We find for most species that the 10\textsuperscript{th} percentile is located at an \( \text{AGI}^{\text{crit}} \) above which habitat volume steeply increases suggesting it acts as an appropriate threshold (Fig. C4).

Regarding species’ data, we assume that our results can be generalized to commercial fish and invertebrates worldwide, as they are based on representative species from different climatic zones (tropical, temperate, polar), vertical habitat (epipelagic, mesopelagic, bathypelagic, demersal), geographic range breadths, taxonomic groups (fish and invertebrates) and size classes. Species distribution ranges were generated by an algorithm developed by the Sea Around Us (see Close et al., 2006; Cheung et al., 2008). The resulting distributions, and the parameters used for their construction are available at http://www.searoundus.org. These distributions have been used to project climate-impacts on fishery resources in a great number of studies (Cheung et al., 2009; Cheung et al., 2010; Fernandes et al., 2013), and are assumed to represent species distributions in 1995-2014 (Tai et al., 2021). Our assumption to extend the 2D distributions provided by Close et al. (2006) over the entire depth range of each species’ depth realm is driven by data sparsity and reliability of 3D species distributions for our selection of species. When reliable 3D habitats can be constructed from species’ observations these could be included (e.g., distribution data are continuously collected in the Ocean Biodiversity Information System but currently are too sparse to provide 3D distribution data). In addition, species-specific thresholds \( pO_2^{\text{threshold}} \) and preference \( T^{\text{pref}} \) are calculated based on the in-habitat spatiotemporal variability of \( pO_2 \), temperature and \( \text{AGI} \) respectively. This is done in lack of observation-based thresholds and preferences that translate to field conditions (Boyd et al., 2018; Collins et al., 2022). Detrimental effects from deoxygenation such as visual hypoxia actually become relevant at much higher \( pO_2 \) than (near) lethal \( pO_2 \) levels (McCormick and Levin, 2017), while only the latter is often what is assessed in the lab. As an effect, the exact threshold of impact remains unknown and probably depends on many factors including the impact of interest, abruptness, magnitude, intensity, duration, heterogeneity and recurrence of exposure to subcritical values (Gruber et al., 2021), as well as timing of and adaptability to unfavorable temperatures, subcritical \( pO_2 \) and hence subcritical AGI.
Through the bias correction of the CMIP6 model data all monthly mean biases relative to WOA18 are removed from our analysis. We acknowledge the influence of observational uncertainties as well as resolution mismatch between our model and the observational WOA18 data used in our bias correction (Casanueva et al., 2020). More complex bias adjustment such as correction for variance biases is prevented by the spatial and temporal resolution of the model and observation data at the global scale. The ongoing effort to collect, compile, and quality-control O2 data in open-access repositories (e.g., Grégoire et al., 2021) will hopefully make it possible to do more advanced bias correction in the future. Until that time the strong temporal variability and spatial heterogeneity of O2 trends complicate the comparison between model and observational data. Nevertheless, the remaining forced response of the models likely underestimates deoxygenation (Andrews et al., 2013; Oschlies et al., 2017; Oschlies et al., 2018; Buchanan and Tagliahue, 2021) and overestimates atmospheric warming (Tokarska et al., 2020) and therefore ocean warming for some CMIP6 models. Part of these warming biases are due to the relatively high climate sensitivities in the CMIP6 models (Meehl et al., 2020). As a further measure to limit model uncertainty we therefore present results at different global warming levels such that they are insensitive to the differences in model climate sensitivity (Hausfather et al., 2022). We last acknowledge the relatively coarse resolution of the CMIP6 data (typically ca. 100km in the ocean) which for species with a highly local distribution (Fig. C1) may lead to higher model uncertainties, especially along the coasts where model disagreement is larger (Fig. C2).

Our approach may give a conservative estimate of contemporary habitat loss since a) crossings of the critical thresholds on timescales shorter than a year are excluded from our analysis, b) CMIP6 projections likely underestimate deoxygenation (Andrews et al., 2013; Oschlies et al., 2017; Oschlies et al., 2018; Buchanan and Tagliahue, 2021), but considering the importance of temperature in driving habitat loss (Fig. 3), especially in the epipelagic realm, the uncertainty of pO2 projections likely has a relatively small effect on our results, and c) we do not include other potential stressors in our analysis such as acidification, changes in ecosystem structure, overfishing, or eutrophication (Bindoff et al., 2019). Examples of crossings of the critical thresholds on timescales shorter than a year would be short hypoxic events and marine heatwaves (Frölicher and Laufkötter, 2018; Jacox et al., 2020; Cheung et al., 2021). Projected deoxygenation and particularly hypoxic or anoxic events have the potential to worsen and even surpass the effects of warming, marine heatwaves, and acidification (Gruber et al., 2021; Sampaio et al., 2021).

On the other hand, for some species the impact will be overestimated if they are able to adapt to future warming and deoxygenation (Cheung et al., 2009; Pinsky et al., 2013; Garcia-Molinos et al., 2016; Palumbi et al., 2019; Collins et al., 2021; Liao et al., 2021). Further note that we considered loss of contemporary habitat only: Mobile species have been observed to redistribute based on the rate and direction of climate change (Pinsky et al., 2013) which can preserve the species range area if they are able to expand into newly suitable areas - however this can alter the original ecosystem structure and function.

For most species we find a loss of habitat volume of less than 10%. It is found for example by Gotelli et al. (2021) that only a small percentage of species drives the observed changes in marine species assemblages, showing that even when only a few species experience large losses, impacts can be profound for the system as a whole. For the individual species however, the loss of only a small fraction of their habitat likely provides adaptation opportunities.
5 Conclusions

- Marine warming and deoxygenation are projected to intensify with global warming and drive a decrease in global habitat viability penetrating to all depths (Fig. 1 and 2).
- The relative decrease in habitat viability (i.e., $AGI_{rel}$) is dominated by warming at the surface while deoxygenation becomes increasingly important with depth (Fig. 2). Loss of contemporary habitat is driven mostly by warming in the epipelagic realm, while in the mesopelagic/bathypelagic and demersal realms reduced $pO_2$ is also contributing for some species (Fig. 3).
- The impact of reductions in habitat viability (i.e., $AGI$: Figs. 1c and 2) on lost habitat volume (Figs. 3 and 4) depends on species' vulnerability (Figs. 5, 6d, C5). Species’ vulnerability is shown to be the most important indicator for large (>5%) potential habitat losses - not relative or absolute changes in $AGI$, $pO_2$ or temperature (Fig. 6). A species’ $pO_2^{\text{threshold}}$ above 150 mbar is an indicator for high species vulnerability to warming (Fig. C6).
- We introduce an updated version of $AGI$: By including temporal and vertical spatial variability in the calculation of the species-specific $O_2$ thresholds and temperature preference, we include a more realistic representation of the in-habitat variability of $O_2$ and temperature and therefore likely the species’ tolerance to these. The updated $AGI$ has lower sensitivity than in the original $AGI$ of Clarke et al. (2021) (Figs. 3 and C7).

Appendix A Tables

Table A1 Species information, ordered alphabetically by species name. Group 1 is epipelagic; group 2 is mesopelagic/bathypelagic, and group 3 is demersal. $pO_2^{\text{threshold}}$ (mbar), $T^{\text{pref}}$ ($^\circ$C) and $AGI^{\text{crit}}$ (-) are based on the WOA18 monthly climatology in the habitat (Fig. C1) of the species except for species in the demersal group for which only a mean climatology is available (see also Sect. 2.2). The slope (change in fraction of total habitat volume per unit change in habitat-mean $AGI$ at $AGI^{\text{crit}}$) is calculated from the species’ CDF (Sect. 3.4).

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<th>$T^{\text{pref}}$</th>
<th>$AGI^{\text{crit}}$</th>
<th>slope</th>
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Appendix B Calculation of $pO_2$

$pO_2$ [mbar] at depth (Taylor, 1978; equation 5 rewritten; Bittig et al., 2018; Sect. E) can be written as a modified Henry’s Law:

$$pO_2 = \frac{[O_2]}{K_0} \cdot \exp \left( \frac{V_m P}{R T} \right),$$  \hspace{1cm} (B1)

with $K_0 = \frac{O_2^{\text{sat}}}{xO_2(1.01325-\rho H_2O)}$, and

$[O_2]$ as the in situ O$_2$ concentration (mol kg$^{-1}$),

$V_m$ the partial molar volume of O$_2$ (31.7⋅10$^{-6}$ m$^3$ mol$^{-1}$ (Enns et al., 1965)),

$P$ the approximated pressure (Pa; $P = 1025 \cdot 9.81 \cdot \text{depth}$ with depth in m)

$R$ the gas constant (8.3145 m$^3$ × Pa × K$^{-1}$ × mol$^{-1}$)

$T$ the absolute temperature (K)

$O_2^{\text{sat}}$ the saturation O$_2$ concentration (mol kg$^{-1}$)

$xO_2$ the dry air mole fraction of O$_2$ in air (0.20946; Glueckauf, 1951)

$\rho H_2O$ the water vapor pressure (mbar)

Where $\exp \left( \frac{V_m P}{R T} \right)$ (unitless) is the pressure correction term for $O_2^{\text{sat}}$. We then first calculate the saturation concentration of O$_2$ in seawater (Garcia and Gordon, 1992) in mol kg$^{-1}$ in Eq. B2.

$$O_2^{\text{sat}} = 10^{-6} \cdot \exp (l) \quad \text{with}$$

$$l = A_0 + A_1 \cdot T_{\text{scaled}} + A_2 \cdot T_{\text{scaled}}^2 + A_3 \cdot T_{\text{scaled}}^3 + A_4 \cdot T_{\text{scaled}}^4 + A_5 \cdot T_{\text{scaled}}^5$$

$$+ \text{salinity} \cdot (B_0 + B_1 \cdot T_{\text{scaled}} + B_2 \cdot T_{\text{scaled}}^2 + B_3 \cdot T_{\text{scaled}}^3)$$

$$+ C_0 \cdot \text{salinity}^2$$

Where $T_{\text{scaled}} = \ln \left( \frac{298.15 - T_{\text{insitu}}}{K_C + T_{\text{insitu}}} \right)$, $K_C=273.15$ K and using salinity (psu) and in situ temperature $T_{\text{insitu}}$ (°C). The A constants provide a correction for temperature, the B and C constants a correction for salinity (Benson and Krause, 1984; Garcia and Gordon, 1992; Sarmiento and Gruber, 2006) and are listed in Table B1.

**Table B1 Constants for the calculation of $O_2^{\text{sat}}$**

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We calculate the water vapor pressure $p_{H_2O}$ (mbar) following Weiss and Price (1980) (Eq. B3).

$$p_{H_2O} = 1013.25 \cdot \exp \left( D_0 + D_1 \cdot \frac{100}{T_{in situ} + KC} + D_2 \cdot \ln \left( \frac{T_{in situ} + KC}{100} \right) + D_3 \cdot \text{salinity} \right),$$

where $D_0=24.4543$, $D_1=-67.4509$, $D_2=-4.8489$, $D_3=-5.44 \cdot 10^{-4}$.
Appendix C Additional figures

Figure C1 Horizontal distributions for each species in dark grey as based on Close et al. (2006), with superscript indicating the species’ depth realm as in Table A1 (1=epipelagic, 2=mesopelagic/bathypelagic, 3=demersal). Species are ordered based on the slope of the CDF (Fig. C5). These 2D habitats were extended over the depth range of the respective species’ group for the analysis (Sect. 2.2).
Figure C2: Multi-model range of AGI\textsuperscript{rel} at 2°C global warming for the three depth intervals studied – as in Fig. 2.
Figure C3 Percentage of remaining contemporary (1995-2014) habitat volume for 2°C global warming shown similar to Fig. 3, including 5 levels of AGI$^{\text{crit}}$ in every species’ boxplot (number of datapoints $n=5$ AGI$^{\text{crit}}$ levels * 6 models = 30): AGI$^{\text{crit}}$ as the minimum in-habitat AGI value, the 5th percentile, the 10th percentile, the 15th percentile and the 20th percentile. Note the different y-axis when comparing to Fig. 3. Each boxplot indicates the median in orange and a box bounded by the interquartile range (IQR; the 25th to 75th percentiles) and the whiskers extending to the data range with a maximum of 1.5×IQR, with outliers as open circles.
Figure C4 Probability density of AGI for each species for the contemporary reference period 1995-2014 (in black) and for 3°C global warming (in red with shaded model range). The PDF is a kernel-density estimate using Gaussian kernels, calculated using Python's SciPy package function ‘gaussian_kde’ with grid-cell volume taken as weights, following Scott’s Rule, and evaluated at 50000 points from an AGI of 0 to 25. Species are ordered based on the slope of the CDF (Fig. C5).
Figure C5 Cumulative density function of AGI at 1995-2014 for each species with “vulnerability” annotated in upper right corner (slope at cumulative density of 0.1; i.e., at AGI=AGI\textsuperscript{crit}). The cumulative density is calculated as the cumulative sum of the probabilities in the PDF estimate (Fig. C4), normalized to a sum of 1. Species are ordered based on their slope.
For each species $T_{\text{pref}}$, $AGI_{\text{crit}}$ and $pO_2^{\text{threshold}}$ (see also Table A1) plotted against their vulnerability (slope at cumulative density of 0.1; i.e., at $AGI=AGI_{\text{crit}}$; see also Fig. C5, Fig. 6, Table A1). Colors as in Fig. 6: For 2 °C global warming, species with > 5% loss marked in red, more than −0.1 ΔAGI in blue, and volume loss < 5% as well as vulnerability > 0.3 in yellow.
Figure C7 Percentage of remaining contemporary (1995-2014) habitat volume of the surface layer (or bottom layer/sea floor for demersal species) for different levels of global warming. This figure is the same as Fig. 3, but here applying surface values only in calculating $p_{O_2}^{\text{threshold}}$, $T_{\text{pre}}$ and $AGI^{\text{crit}}$ and hence AGI and volume=AGI$^{\text{crit}}$. Demersal results are logically the same as in Fig. 3 as these consider only the sea floor. Note the different y-axes (also when comparing to Fig. 3). Each boxplot indicates the median in orange and a box bounded by the interquartile range (IQR; the 25th to 75th percentiles) and the whiskers extending to the data range with a maximum of 1.5xIQR, with outliers as open circles. As changes are expressed relative to the contemporary viable habitat volume (which is by definition 90% of the total habitat volume), values up to 111% (=100/90) are possible.

Data availability

Scripts used in our analysis will be shared with a DOI upon publication of this manuscript.
Author contributions

All authors contributed to the conceptualization of the study. AM and TF collected the model data, while WC and TC provided species data. AM executed the formal analysis, investigation, and validation of the study with input from all co-authors. The extended version of AGI methodology was developed through discussion between all authors. AM wrote the initial draft and visualized the results with input from all co-authors.

Competing interests

The authors declare that they have no conflict of interest.

Acknowledgments

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