Plume detection and estimate emissions for biomass burning plumes from TROPOMI Carbon monoxide observations using APE v1.1

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Abstract.

This paper presents the Automated Plume Detection and Emission Estimation Algorithm (APE), developed to detect CO plumes from isolated biomass burning events and to quantify the corresponding CO emission rate. APE uses the CO product of the Tropospheric Monitoring Instrument (TROPOMI) aboard the Copernicus Sentinel-5 Precursor (8-5PS5P) satellite, launched in 2017 and collocated active fire data of from the Visible Infrared Imaging Radiometer Suite (VIIRS), the latter flying 3 min ahead of S-5P. After identifying appropriate fire events using VIIRS data, an automated plume detection algorithm based on traditional image processing algorithms selects plumes for further data interpretation. The approach is based on several thresholds which are tuned thresholds optimised for data over the US on United States in September 2020. Subsequently, the CO emission rate is estimated using cross-sectional flux method, which requires horizontal wind fields at the height of the plume. To infer proper wind fields from ECMWF reanalysis 5 data, we test three-plume height. Three different plume heights . We consider a constant plume height at 100 m and at the injection height provided by Global Fire Assimilation System (GFAS) and we simulate the were considered and the ECMWF reanalysis v5 data (ERA5) was used to compute emissions. A varying plume height in downwind direction with a the downwind direction based on 3D Lagrangian modelsimulation was considered appropriate. APE is configured to optimize performance of the US and its performance is verified for observations over Australia and Siberia. APE identified 5562 fire clusters and only 1327 cases was deemed to have good For all fire sources identified by VIIRS, only 16% of the data corresponded to clear sky TROPOMI CO data . 882 plumes were detected in 1327 cases and only 378 plumes were considered as 309 plumes were short and about 195 had multiple sources of fire in them. Finally, the emissions were estimated for 226 cases in 378 cases out of which 5 were further visually filtered leading to 221 cases. The constant plume height at 100 m provides unacceptable emission estimates. The difference in emission estimates using a plume height at the GFAS injection height and using Lagrangian simulation, was found < 4% for Siberia but is significant for several cases for US and Australia. Therefore, APE will employ the simulated varying plume height in downwind direction for the automated algorithm with plume signature. Furthermore, the quality filters of APE resulted in emission estimations for 26% of the TROPOMI CO data with plume signatures. Visual filtering of the APE's output showed a true-positive confidence level of 97.7%. Finally, we provide a first an estimate of the emission uncertainties. The assumption of a constant emission over the time of the plume formation and the spatial under-sampling of

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the greatest contribution of error comes from the uncertainty in GFAS injection height that leads to emission errors < 100%, followed by systematic errors in the ERA5 wind data. The assumption of constant emission during plume formation, and spatial undersampling of CO column concentration by TROPOMIyields error < 20 %Mostly, an emission plume created by a burning and uncertainties in the GFAS injection height may cause errors, yield an error of <100 %. Errors in the 20%. The randomised errors from the ensemble ERA5 wind data are considered to be significant but could not the quantified because of the lack of uncertainty information in the used wind fields found to be less than 20% for 97% of the cases.

1 Introduction

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Carbon monoxide (CO) is an air pollutant and in high concentrations , it causes harmful healtheffects. Moreover, it harms human health. It is an indirect greenhouse gas and a contributor to increase in several greenhouse gases in the atmosphere as it alters atmospheric OH, thus leading to an increase in the lifetime of methane (Spivakovsky et al., 2000). CO is produced mainly due to incomplete combustion, for example, Andreae et al. (1988); Watson et al. (1990) have showed that the biomass burning increases CO in atmosphere. Shi et al. (2015) quantified the total COemissions from vegetation burning, fuel-wood combustion, and human waste in different tropical regions. Alsoby incomplete combustion and Andreae et al. (1988); Watson et al. (1990) showed that biomass burning is a significant source of atmospheric CO. Furthermore, Granier et al. (2011); Crippa et al. (2018); Hoesly et al. (2018) show that the CO emission due showed an increase in CO emissions due to fossil fuel burning has increased since 2000. Additionally, CO emitted by localized sources at localised sources on the ground leads to a prominent footprint in the atmosphere , namely plumes, due to its lifetime ranging from days to several months (Holloway et al., 2000). Thus, it becomes essential to understand. These atmospheric characteristics can be observed by satellites, which can provide essential information to improve our understanding of the effect of CO on air-quality and climateby measuring and quantifying it accurately on a global and local sealesair quality and climate.

The TROPOspheric Tropospheric Monitoring Instrument (TROPOMI) aboard the Sentinel-5 Precursor (S5P) satellite, launched in 2017, monitors CO daily on a global scale (Borsdorff et al., 2018) and at with a high spatial resolution of $7 \times 7 \text{ km}^2$, improved to $5.5 \times 7 \text{ km}^2$ in August 2019. Due to it's the high spatial resolution and the daily coverage the CO Level-2 dataset, the CO emissions by daily coverage, CO emissions from cities (Borsdorff et al., 2019a, 2020; Lama et al., 2020), wildfires forest fires (Schneising et al., 2020; Li et al., 2020; Magro et al., 2021; van der Velde et al., 2021) and industrial sources (Tian et al., 2021) have been quantified and studied investigated and quantified. Rowe et al. (2022) compared the TROPOMI CO measurements to with in-situ aircraft measurements for different biomass burning plumes in 2018 and found that the difference between the two measurements is about differences of < 7.2%. This shows that the illustrating the potential of TROPOMI CO measurements can be use to quantify the biomass burning events to quantify CO emissions from these sources.

Furthermore, most of the above studies estimate CO flux on large-scale regions Most of the studies mentioned above estimated CO fluxes on large regional scales (Schneising et al., 2020; Magro et al., 2021; van der Velde et al., 2021) and mega-city scales (Borsdorff et al., 2019a, 2020; Lama et al., 2020). However, not many single point emissions are quantified So far, the single-point CO emissions estimated from the TROPOMI CO dataset data have received less attention. Tian et al.

(2021) showed CO emissions based on TROPOMI for single-point single-point industrial sources from India and China. They were able to perform performed a statistical study for three years, as the geo-location is known for of the industrial source is known. A similar study analysis to quantify emissions from single-point biomass burning (fires) using TROPOMI CO data has not been shown in the literatureas geo-locations for fires are not known a priori. The geo-locations for fires.

Fire locations can be detected using the Visible Infrared Imaging Radiometer Suite (VIIRS) 375m thermal anomalies/active fire product (Schroeder et al., 2014). The VIIRS instrument is aboard the joint NASA/NOAA Suomi National Polar-orbiting Partnership (Suomi NPP) satellites satellite and flies in the same orbit as S5P, in loose formation with a temporal separation of 3.5 minutes between them. This short time difference helps in collocating allows us to collocate observations of TROPOMI CO data and the VIIRS active fire data productfor this studyproduct.

The CO plumes in the TROPOMI data can be used to estimate CO emission by fires and different methods Different methods of estimating emissions are discussed in the literature, namely, the inversion methods coupled with Gaussian dispersion models (Krings et al., 2011; Nassar et al., 2017; Lee et al., 2019) or, different Chemical Transport Models (CTM) (Brasseur and Jacob, 2017), Cross-sectional Flux Methods (CFM) (White et al., 1976; Beirle et al., 2011; Cambaliza et al., 2014, 2015; Kuhlmann et al., 2020) and integrated mass enhancement (IME) method (Frankenberg et al., 2016). The An inversion coupled with a Gaussian plume model can be is used for flux inversions of an isolated single plume event assuming steady and uniform wind conditions. The This method fits an analytically computed Gaussian plume to TROPOMI CO column observations. The method and can only be applied to observations under very specific wind conditions (Varon et al., 2018). The **EM** uses an empirical linear relation between IME method relates the emission and the integrated mass in the observed plumethat was established, and Frankenberg et al. (2016) showed that the relation is linear based on aircraft data (Frankenberg et al., 2016) . Current implementations are only applicable for single emission events for methane plumes and. However, no such relationship has been established for CO measurements in fires. Thus, we do not consider IME around fires. Hence, IME is not considered for the present work. The inversion methods using CTMs, such as, the Weather Research and Forecasting model coupled to Chemistry (WRF-Chem) (Grell et al., 2005), GEOS-Chem (Bey et al., 2001), and others, can reduce uncertainties thereby predicting and thus predict emissions more accurately. They Although these methods can be applied to complex emission events, but the corresponding simulations are complex, computationally expensive and difficult to automate, in particular particularly for a large number of fires with different geolocations which is the objective of this study. The CFM is suited for the interpretation of isolated plumes but does not assume a particular plume shape. It needs comparatively well suited to the present work, as it requires less computational power and is easier to automate. H-CFM is based on the mass conservation of the pollutant transport in the downwind direction of the plume. The CO-emission is estimated from corresponding fluxes across different planes perpendicular to the direction of the plume using the wind velocity at the plume height. The Brunner et al. (2019) showed that the plume height depends upon on different aspects, namely, meteorology, emission height, etc(Brunner et al., 2019)., and may not be explicitly available. MoreoverFurthermore, the CFM breaks down when diffusion is dominant, i.e., that is when the wind velocity is $< 2 \text{ ms}^{-1}$ (Varon et al., 2018).

The present work aims at developing an automated scheme to detect single and spatially isolated emission of emissions from biomass burning events on the spatial scale of in TROPOMI observations and to estimate the corresponding CO emissions. For

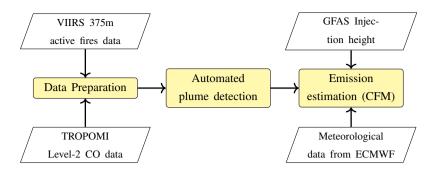


Figure 1. High-level flow chart of the APE algorithm.

this purpose, we employ and improve the CFMas it has the potential to be applied in an operational data processing employed and improved the CFM. First, VIIRS fire data and satellite data are—were prepared for automated plume detection which is discussed in section 2.1. Plume Section 2.1. The plume detection algorithm from a single point source using VIIRS fire counts is the subject of section Section 2.2. Section 2.3 describes the emission estimation using cross-sectional flux method the CFM where an appropriate choice of the plume height and the wind fields are is discussed. The study results results of our study are discussed in Section 3and finally, and finally, Section 4 concludes our study and sets recommendations for the future work.

100 2 Methodology

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Figure 1 illustrates a high-level flowchart of APE. Corresponding pseudo-code flow diagram of the automated plume detection and emission estimation algorithm (APE), and the corresponding pseudocode is given in Appendix B algorithm description Algorithm 1. It-APE is divided into three parts, namely data preparation, automatic plume detection, and emission estimation. During data preparation, the algorithm identifies single point The data preparation algorithm identifies single-point fire sources from the VIIRS 375 m active fire data product (Schroeder et al., 2014) and subsequently selects and extracts TROPOMI CO data around every located fire source. The Thereafter, the plume detection algorithm searches for a plume in the extracted CO datawhich is required, and a detected plume serves as an input for emission estimation. The emission estimation algorithm initially computes the background CO, which is the usual observed CO concentration at that location without any CO emissions due to from the fire. Subtracting the The background allows us to obtain the enhanced CO which further is used to estimate the emissions by cross-sectional flux methodCO enhancement, which is used by the CFM to estimate CO emissions. These three parts of the algorithm are discussed in detail in the following sections.

2.1 Data preparation

2.1.1 Selection of fire events

Fire events are inferred from the VIIRS 375m active fire data product (Schroeder et al., 2014) provided by the Fire Information for Resource Management System (FIRMS) which; FIRMS is operated by NASA's Earth Science Data and

Information System -(https://earthdata.nasa.gov/active-fire-data). The data contains different include various parameters such as fire radiative power (FRP), temperature and the time of measurement defined at-in latitude-longitude coordinates which correspond to the center. Each of these coordinates corresponds to the centre of a $375 \times 375 \text{ m}^2$ ground pixel. From now on, each of these latitude longitude coordinates will be referred to and is referred to in this paper as a fire count in this paper. Mostly, an emission plume created by burning or a VIIRS pixel. In most cases, a fire within a single VIIRS pixel cannot 120 be detected by TROPOMI with its pixel size of 7×7 or 5.5×7 . Only clusters create a CO signature spanning multiple TROPOMI pixels due to the detection limit of the satellite. Only larger fires with a cluster of VIIRS fire counts can lead to a detectable CO plume in the observations. To identify appropriate fire clusters, we employ TROPOMI observations. We used the Density-Based Spatial Clustering of Applications with Noise (DBSCAN) (Ester et al., 1996; Schubert et al., 2017) algorithm 125 algorithm (Ester et al., 1996; Schubert et al., 2017) from the scikit-learn library (Pedregosa et al., 2011) to identify fire clusters. It separates the areas which areas that are densely packed with fire counts from the low density areas and has an areas of low density and therefore has the ability to detect arbitrarily-shaped clusters. DBSCAN requires two inputs, first-clusters of any shape. DBSCAN takes two inputs; the first is the maximum search radius $-r_{max}$ around a fire countand second, and the second is the minimum number of fire counts within the area $-n_{\min}$. r_{max} is set to 4 km which is approximately half about half 130 the size of the TROPOMI pixel size. The minimum number of fire counts have been empirically set to $n_{\min} = 10$. For further analyses, we convert each cluster to analysis, we converted each cluster into a single point source using the fire radiative power (FRP) as weights the weight of the individual fire counts. This single point source will single-point source will henceforth be referred to as fire source from now on and serves as an input to a fire source and will be used as input to the TROPOMI CO data preparation.

135 2.1.2 TROPOMI CO data preparation

For the identified fire sourcesfrom VIIRS dataVIIRS fire sources, the corresponding TROPOMI orbit (See orbits (see Table B1 for the version of the L2 productversion) is selected. Figure 2 illustrates the collocated information for part of TROPOMI orbit number 10254 over Australia. The orbit is—) were selected and the orbit was corrected for stripes (see the Fast Fourier Transformation algorithm of Borsdorff et al. (2019b)). For each fire source, we extract Figure 2 shows an example of the collocated information for a part of a TROPOMI orbit over Australia. Then, we extracted a data granule of 41 × 41 TROPOMI CO pixels centered around the centred on each fire source. The granule size is minimum minimum granule size of 220 km is chosento capture the distance of the emission traveled from the source to the granule edges within 6h was chosen, as an air mass with an average velocity of 5 ms⁻¹ takes 6 h to reach the edges of the granules from the centre. After extraction, two data quality filters are applied.

- DP-1 The maximum TROPOMI CO pixel size due to distortion in the swath direction is restricted to < 12 km to avoid large pixel size and its variation within the granule.
 - DP-2 A sufficient number of CO pixels with For a data granule, 80% of all CO data must meet a data quality $q_a > 0.5$ (Apituley et al., 2018). $q_a > 0.5$, which corresponds to clear-sky, elear-sky-like and mid-level elouds observations.

Within the entire data granule 80% of all pixels must fulfil this quality criteria. Additionally cloud observations. Furthermore, we require 85% of 'good' the pixels in an area of 7×7 pixels centered around 7×7 pixels centered on the fire source to meet the above criterion. The more usable pixels around plume the source, the better the plume can be disentangled from the atmospheric background (see discussion in Sec Sect. 2.3.1).

The threshold values are empirically determined for a reference data set from September 2020 over the US-United States and verified for two other data sets over Australia and Siberia (see the Tab. B1 in Appendix B for a detailed specification of the data sets). Finally, the selected CO scene corresponding to a fire source is forwarded as an is passed on as input to the plume detection algorithm.

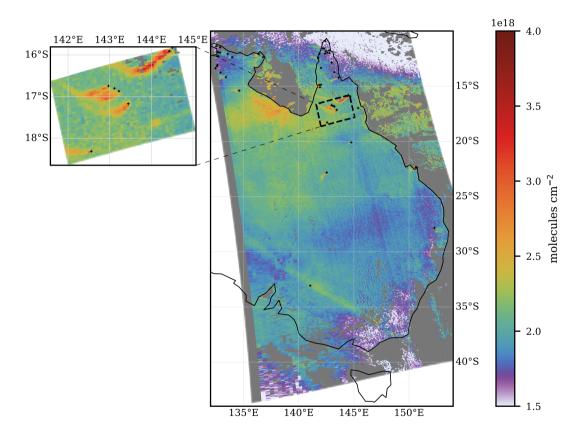


Figure 2. 49 detected fire sources represented by black '+' on 2019-10-06 overlapped with the TROPOMI level 2 CO data for orbit 10254. The dashed region represents a 41×41 pixel granule.

2.2 Plume detection algorithm

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The next step of APE identifies in APE is to identify the plumes within each selected CO data granule. Kuhlmann et al. (2019) developed a plume detection algorithm based on statistical methods, and Finch et al. (2021) used machine learning to detect plumes. A In the present study, a machine learning approach is not considered in the present study mainly due to non-availability

the unavailability of data containing detected plumes and their sources for training. Instead, our plume detection approach is based on traditional image processing algorithms (van der Walt et al., 2014).

Using the extracted CO TROPOMI data, a plume is detected by a region-based segmentation algorithm, where pixels with similar properties are clustered together to form a homogeneous region. One of the most commonly used and classic region-based segmentation algorithm algorithms is the 'marker-based marker-based watershed transform method' (Beare, 2006; Gao et al., 2004). The CO column concentration represents metaphorically metaphorically represents the altitude of a topographic map. Thus, the watershed algorithm segments the regions into valleys and mountains (CO enhancements) based on a given marker and a gradient map. In the following paragraphs, we describe the plume detection in more detail using an example.

The marker-based watershed algorithm in the scikit-image package (van der Walt et al., 2014) takes two inputs to segment an image, one. One is the 'gradient map' I_{grad} where the, which emphasises changes in altitude are emphasized and homogeneous regions are dampened and attenuates homogeneous regions. The second input is a marker image I_{mark} which that provides the seed points referred for the algorithm, referenced by an integer label for the algorithm. The definition of both inputs is discussed in the following paragraphs.

We start with the extracted CO TROPOMI granule of a size 41×41 pixel size pixels $\mathbf{I}(i,j)$ with $i,j=1,\cdots,41$. An example is shown in Figure 3(a). First, high frequency components or noise the high-frequency components of the CO-image are reduced by a 2D Gaussian filter with standard deviation a standard deviation of $\sigma = 0.5$ (pixel), which was pixels, chosen empirically. The smoothened image is referred to as smoothed image is called \mathbf{I}_s . From this image, the gradient map \mathbf{I}_{grad} is computed using a Sobel operator (Sobel and Feldman, 1990; van der Walt et al., 2014) as shown below, namely

$$\mathbf{I}_{\text{grad}} = \sqrt{\mathbf{G}_x + \mathbf{G}_y} \tag{1}$$

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$$\mathbf{G}_{x} = \begin{bmatrix} 1 & 0 & -1 \\ 2 & 0 & -2 \\ 1 & 0 & -1 \end{bmatrix} * \mathbf{I}_{s} \qquad \mathbf{G}_{y} = \begin{bmatrix} 1 & 2 & 1 \\ 0 & 0 & 0 \\ -1 & -2 & -1 \end{bmatrix} * \mathbf{I}_{s}, \tag{2}$$

where * represents the convolution operator. Here, the gradient \mathbf{I}_{grad} emphasizes emphasizes the edges of a plume, as shown in Figure 3b.

By default, the marker image (\mathbf{I}_{mark}) is initialized initialised to zero and subsequently then two different seeds are defined.

One seed indicates regions which do not have without CO enhancements and another refers to regions of with clear CO enhancements given by $\mathbf{I}_{\text{mark}}(i,j) = 1$ and $\mathbf{I}_{\text{mark}}(i,j) = 2$, respectively. The seeds are defined as follows:

1. The regions of no CO enhancements without CO enhancement: A pixel I_{mark} (i, j) has no does not have CO enhancement if it is either below the median of I_S or below the mean of the 15x15 pixels centered 15 x 15 pixels centred at I_S(i, j). The size of 15x15 pixels is empirically 15 x 15 pixels was chosen to account for the background variability. The pixels Pixels corresponding to no enhancement can be seen in Figure 3c represented by the label '1' and the image is referred to as a preliminary marker image.

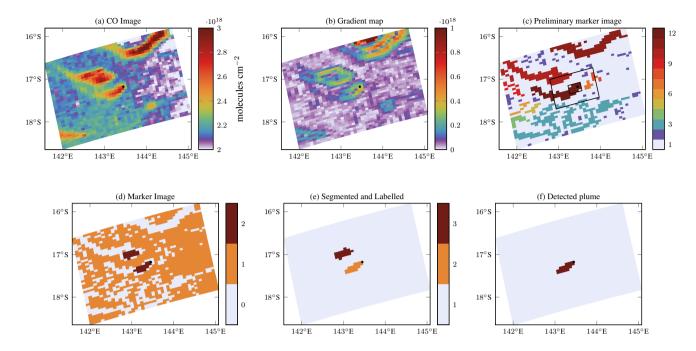


Figure 3. Plume detection algorithm. Figure shows An example showing different steps of the algorithm for an example plume for a fire source on 6th October 2019 in Australia.

2. The regions with CO enhancements of CO enhancement: Using the preliminary marker image with labels '0' and '1', we identify identified all connected pixels with the same marker value (hereafter referred to as connected regionshereafter) using the 'label' algorithm (Fiorio and Gustedt, 1996) in of the scikit-image package (van der Walt et al., 2014). Each connected region is identified by an a unique integer value per pixel (not to be confused with the seed marker). Next, we zoom in an zoomed in on a 5 × 5 pixel area around the fire source, and extract and extracted all connected regions as a potential plumes. We extend Further, the potential plumes were expanded by going to a 15 × 15 pixels around the fire source using pixels of with the same label. We extract all labeled Then, the labelled CO data in this area and calculate were extracted and a CO threshold was calculated as their mean CO value. We mark value. Lastly, all pixels within the 15 × 15 pixel area with a seed were marked with a label '2' which have CO values, if their CO value was above this threshold, which further. This yields the remaining seed points which are defined only in a 15 × 15 pixel area around the fire source.

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The above selection process is illustrated in Fig. 3c with the different labels of the connected regions. The final marker image is shown in Figure 3d. Finally, the watershed algorithm ealeulates computes a segmented image for the entire domain using the gradient map image $I_{\rm grad}$ and the marker image $I_{\rm mark}$. Figure 3e gives shows an example of a segmented image. Using the gradient map, the watershed algorithm has decided that the two areas of enhanced CO values are not connected and so, and therefore do not belong to the same plume. From the segmented image, we extract extracted the correct plume, which should

originate from emissions at the source location. Therefore, we consider only those labeled labelled areas that overlap with the center centre 7×7 pixels. Figure 3f shows the detected plume. The detected plume in this case seem appears to be shorter, however in this case, but the tail end of the plume, i.e., around $< 143^{\circ}$ E, will fail background subtraction due to similar enhancements as the background. This can also be seen in gradient map the gradient map, where no gradient is detected on at the top side of the plume.

Finally, the extracted plume is evaluated regarding its suitability suitability of the extracted plume for further processing and the plume length is calculated for each extracted plume. Only the data are is evaluated and the length of the plume is calculated. The plume is provided to the emission estimation module if:

PD-1 the plume length is > 25km. If the plume length

PD-2 If there are not more than nine non-clustered fire counts or any other identified fire cluster within 0.05° distance from or in the identified plume.

If the length of the plume is < 25 km, then the detected plume is flagged as a short plume and will be ignored for further processing. The short plumes are difficult to quantify in an automated way as they can have different shapes, which makes it difficult to identify the plume direction. Before starting the emission estimate module, we remove direction of the plume. The second criterion (PD-2) removes all plumes with multiple fire sources, as the aim of this paper is to quantify fires with single sources. This is done by the following filter:

PD-2 Check if there are not more than 9 non-clustered fire counts or any other identified fire cluster within 0.05° distance from or in the identified plume.

2.3 Emission estimation

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For the detected plumes, the emissions were estimated using the cross-sectional flux method (CFM) (White et al., 1976; Beirle et al., 2011; Cambaliza et al., 2014, 2015; Kuhlmann et al., 2020). The CO emission $E_{-}E_{-}$ is defined as the mean flux through n cross sections perpendicular to the downwind direction of the plume, namely

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$$E = \frac{1}{n} \sum_{i=1}^{n} Q_i$$

$$Q_i = \int \delta C_{\text{co}}^i(s, t_0) \cdot v^i(z_i, s, t_0) \cdot ds$$
(3)

where Q_i (in kgs⁻¹) is the CO flux through eross section cross-section i, $\delta C_{\rm co}^i$ (in kgm⁻²) is the background subtracted CO values along a cross-section i and v^i (in ms⁻¹) is the velocity perpendicular to the cross-section i. The wind Wind velocity $v(z,s,t_0)$ at the plume height z, the cross-section position s and at the observation time t_0 is are obtained from the European Center for Medium-Range Weather Forecasts (ECMWF) data of the European Centre for Medium-Range Weather Forecasts Reanalysis v5 (ERA5) data (Hersbach et al., 2017). For error characterization characterisation, we define the standard error

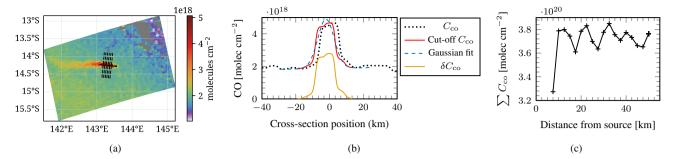


Figure 4. Plume on 2019-05-19 at 04:55 UTC. (a) Plume and every second transact lines drawn based on the detected plume separated by 5 km in downwind direction. (b) The black dotted line corresponds to the CO column along a transect in (a), and the red line shows re-centered re-centred and cut-off CO used for Gaussian fitting. Blue-dash dotted The blue dash-dotted line corresponds to the Gaussian fit, and the orange line represents the enhanced CO along the transect. (c) $\sum C_{co}$ along a transect against the distance from the source.

 $(\sigma_{\rm E})$ as

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$$\sigma_{\rm E} = \frac{1}{n} \sqrt{\sum_{i=1}^{n} (E - Q_i)^2}$$
 (4)

To determine the The cross-sections, hereafter referred to as transects, we first calculate the direction of the plume in were determined by calculating the plume direction in the downwind direction. The plume line results from a second-order curve fit through the pixel centers centres of the detected plume (see, e.g. the black solid line in Fig. 4a). Next, we calculate the transects at every 2.5 km perpendicular to the plume line were calculated and are illustrated as dashed lines in Figure 4a. Transects The transects are sampled at 2.5 km to reduce the errors due to interpolation discussed in discussed in the next paragraph. To calculate compute Q_i in Eq. 3, each transect is was sampled at distances of 500 m. The points over an transects are over-sampled to get Points over transects are over-sampled to obtain a smoother CO distribution, which further helps in the background subtraction discussed in SeeSect. 2.3.1. Along each transect, the CO column (C_{co}) is extracted by linear interpolation of the original CO data and is illustrated by a dotted black line in Figure 4b. This CO column is further used to compute calculate δC_{co} in Eq. 3. During the diagnostic diagnostic tests of our interpolation algorithm, an oscillation was observed in the CO columns integrated along the transects as a function of the downwind downward distance from the fire source (see Figure 4c). The oscillation is due to the under-sampling of the CO distribution by the TROPOMI instrument. The distance between two minima is approximately equal to the TROPOMI pixel size. This error is found to propagate propagated further into the CO enhancement δC_{co} which is computed from δC_{CO} , which was computed from the background subtraction algorithm.

2.3.1 Background Subtraction

To determine the atmospheric background of CO per transect, first we re-center the $C_{\rm CO}$, we re-centre the $C_{\rm CO}$ such that the maximum is at the origin to facilitate the Gaussian fit. The transect line is truncated at the first minima of CO on either side of

the origin, as illustrated by the red line in the Fig. 4b. To determine the background for each transect (red line), we assume that the column CO along the transect can be expressed as

$$C_{co} = H_0 + H_1 \cdot s + A_0 G(s)$$
 (5)

where H_0 and H_1 represents represent the background and the slope of change in the background over the transect, respectively. A_0 is the amplitude of the Gaussian distribution (G). We determined the background by fitting Eq. 5 through the CO data, which we subsequently subtracted from the $C_{\rm co}$ data to calculate the CO enhancement as shown below

$$\delta C_{\text{co}} = \max\{0, C_{\text{co}} - H_0 + H_1 \cdot s\}.$$
 (6)

Here, the negative CO column enhancements enhancements in the CO column are ignored. The blue-dashed and the orange line lines in Fig. 4b represent Gaussian fit and $\delta C_{\rm co}$, respectively.

2.3.2 Filtering during background subtraction

The background subtraction includes an important filtering mechanism to remove overlapping plumes. This is done during the background subtraction after the transect line is truncated. The filter <u>criteria</u> criterion is as follows

270 EE-1 Difference between the two The difference between the minima on either side of a truncated transect should be < 10% of the smallest value. This ensures a smooth background and so the absence of any interference with adjacent emission events. It should be noted that this filter also account for background situation which are too demanding for the CFM.

2.3.3 Plume height

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The plume height z_i at a transect/cross-section i is used to extract the appropriate wind velocity $v(z_i, s, t_0)$. For wildfires, Rémy et al. (2017) showed that the an-Integrated monitoring and modeling modelling System for wild-land Fires (IS4FIRES) injection height, $z_{\rm inj}$, from the Global Fire Assimilation System (GFAS) database is in a good agreement with the observations. Sofiev et al. (2012) showed the IS4FIRES injection height deviated by less than 500 m compared to the MISR Plume Height Project (MPHP), thus we consider 500m therefore we consider 500 m as plume height uncertainty. First, we assume that the plume height is equal to height of the plume is $z_{\rm inj}$ and is constant throughout the plume, which may hold. This may be true for stable meteorological conditions. The constant plume height will be referred as called $z_{\rm c}$ and the uncertainty at this plume height is given as $z_{\rm c}^p$ and $z_{\rm c}^m$ which correspond of $z_{\rm inj}$ + 500 m and $z_{\rm inj}$ - 500 $z_{\rm c}^p$ = $z_{\rm inj}$ + 500 m and $z_{\rm c}^m$ = $z_{\rm inj}$ - 500 m, respectively.

It should be noted that the injection height computed calculated in GFAS is for 24 h and may not be appropriate for a satellite plume which is a snapshot at a measurement time t_0 . Additionally In addition, the plume height might may vary due to meteorology in the downwind direction. Therefore, we simulated alternatively alternatively simulated particle trajectories starting at the fire location site around the injection height with a three-dimensional Lagrangian tracer

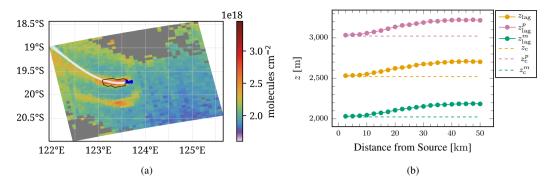


Figure 5. (a) The white band shows all tracer particles at the end of the Lagrangian simulation, and the blue dots show the fire counts on the detected plume. (b) Shows the plume height computed for different transects from Lagrangian simulations. The constant plume height (z_c) represented by an orange dashed line is 2521.87 m.

dispersion model. This allows us to estimate the The local plume height z_i by a vertical averaging of is then estimated by averaging the height of the tracers along the downwind direction. The This estimated plume height is referred as z_{lag} in the following, which and captures the change in height in the downwind direction.

The Lagrangian simulations are were performed using tracer particles. The motion of tracers is simulated according to

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$$\frac{d\mathbf{x}_p(t)}{dt} = \mathbf{v}(\mathbf{x}_p(t)) \tag{7}$$

where $\mathbf{v}(\mathbf{x}_p)$ represents the fluid velocity at the instantaneous particle position \mathbf{x}_p . The explicit forward Euler scheme (Butcher, 2003, p. 45) is employed was used to integrate the equation in time. The velocity on the right-hand side of the Equation. (7) is calculated by the tri-linear trilinear interpolation of the ERA5 velocity fields. The source locations for the Lagrangian simulations are based on the fire counts described in Section 2.1.1. At each source location, 3 are used as source locations for Lagrangian simulations. Three tracer particles are released at $z_{\rm inj}$ and $z_{\rm inj} \pm 500$ m and the particle trajectories are simulated. Particles are released at each source location. The release at $z_{\rm inj} \pm 500$ m is used for uncertainty analysis. The end time of the simulations is the TROPOMI measurement time t_0 , which is about $t_0 \approx 13:30h$ 30 h local time (Veefkind et al., 2012), and the simulation starts at $t_0 - 6$ h. The particles are released from the source locations every 2 minutes. Figure 5(a) shows a simulation of the tracer particles for one plume. The white band indicates particles at the tracer particle simulation for a plumeand the grayish band indicates them at TROPOMI measurement time.

The contribution to emissions from fires fire emissions is low in early morning the early morning, as shown in the ecosystem-specific diurnal cycles by Li et al. (2019) and so. Therefore, we ignore trajectory simulations before $t_0 - 6$ h. Additionally. Additionally, the process of heating due to fires is not accounted for in our Lagrangian simulation as we assume the ERA5 velocity fields contain some aspect of it-heating, as ERA5 assimilates skin surface temperatures.

At each intersect, the height In each transect, the heights of the tracer particles released at $z_{\rm inj}$ are were extracted and the mean height, $z_{\rm lag,i}$ is computed and $z_{\rm lag,i}$ was calculated. This is assumed to be constant along the transect. Figure 5 (b) shows the plume height height of the plume for different transects from the fire source which is used to compute velocity, that was

used to calculate the velocity v, in Equation 3. The uncertainty in plume height is defined as z_{lag}^p and z_{lag}^m and are computed was calculated from tracer particles that were released at height heights $z_{\text{inj}} + 500 \text{ m}$ and $z_{\text{inj}} - 500 \text{ m}$, respectively, and can be observed in Figure 5(b). Finally, the velocity, v, is was used to compute emissions.

2.3.4 Filtering during Lagrangian simulations

Related to the Lagrangian simulation, we apply three filters:

- EE-2 The injection height from GFAS must be available.
- EE-3 If the simulated trajectories are not aligned in the direction of the plume, then the plume is rejected.
 - EE-4 If the wind velocity at the TROPOMI measurement time which is used to compute emissions is less than $2 \, \text{ms}^{-1} \, \text{ms}^{-1}$ then the plume is rejected.

Filter EE-2 may become relevant due to the false detection of plume or a plume, false fire in the VIIRS active fire database or just the data is missing in the missing data in the GFAS database. Reason There are several potential origins for filter EE-3might be a rotation or the errors in the: the rotation, errors in ERA5 velocities or due to, the spatial and temporal resolution of velocity fields or inaccurate injection height. Finally, if the wind speed is below the specified value in EE-4, diffusion dominates the pollutant transport, and CFM is not appropriate to estimate the CO emission.

3 Algorithm application

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Table 1. Results for automated plume detection and emission estimation algorithm (APE v1.1) for four months in the US, Australia and Siberia

Regions	Fire Clusters	CO data	Plume detection	Emission estimation	Visual Inspection
US	1081	213	130	130 37	
AU	2013	385	266	129	128
Sib Jun	416	130	83	35	34
Sib Jul	2052	599	94	25	24
All Regions	5562	1327	378	226	221

Our The APE algorithm targets at global performance and includes several threshold values, which needs to be determined carefully need to be carefully determined for optimal performance. For the current version of APE V1.1-, we decided to determine the thresholds using the region encapsulating that encapsulates the United States of America (US) on Sept in September 2020. The algorithm is verified by applying it to other regions encapsulating that encompass Australia (AU) on Oct in October 2019 and Siberia (Sib) on Jun and Jul- in June and July 2021 (see Table- B1 for more details on the selected regions). It is

important to note that these regions are not used to configure APE and so can can therefore be used to test-verify the overall performance of the algorithm. The different time periods were chosen to focus on the regional burning season and so to maximize number of observed fires to maximise the number of fires observed. Table 1 shows the number of plume evaluated in the different modules of APE. cases evaluated by different APE modules. The columns Fire cluster and CO dataeolumns highlight the results of plume detection, and emission estimation show the results for data preparation (see Section. 2.1)part. The columns plume detection and emission estimation show the results for plume detection (sec. 2.2) and emission estimation (sec. 2.3)parts, respectively. Furthermore, the details corresponding to filtering can be found in Tables B2, B3 and B4 in appendixthe Appendix.

A total of 5562 fire clusters sources (see Table. 1) were identified in the VIIRS active fire data product for all regions based on the clustering method discussed in Section 2.1.1. For each fire cluster, source, the TROPOMI CO data was were filtered for maximum pixel size and quality (see SeeSect. 2.1.2). The filter on the TROPOMI pixel size filter (DP-1) filtered about rejected 1533 cases out of 5562 cases which that mostly belonged to Australia and USthe United States. The quality of TROPOMI CO data (DP-2) was found to be bad insufficient for about 2553 cases in 5562 cases, mainly due to the presence of clouds. For the Siberian region on July 2021, more than 50% of all fire clusters are flagged as bad quality data because of more cloudy observations at these latitudes. Finallybad-quality data for the same reason. Finally, the data preparation part yielded a total of 1327 good CO data granules for all regions for further processing.

The plume detection algorithm described in SeeSect. 2.2 identified a total of 378 plumes plume signature in 882 cases for all regions from available 1327 good CO data cases available. A total of 445 cases were found to have enhancements (see Table B3)which means that the CO enhancement, meaning that the enhancement of CO from these fires was not significant enough to be measured by below the detection limit of TROPOMI. In addition to this, about 882 cases, only 378 cases were considered good as the PD-1 filter flagged 309 identified plumes were flagged as short as the length of these plumes was found to be shorter than their plume length was < 25 km(PD-1 filter). Further, Furthermore, the PD-2 filter identified a total of 195 cases where other fire sources and clusters were present in the detected plumes.

The emission estimation algorithm takes took 378 plumes as an input and computes emissions using input and calculated emissions using the CFM for a total of 226 cases. So Therefore, a total of 152 plumes were rejected by filter the EE-1, EE-2, EE-3, and EE-4 during the filters during emission estimation (See Table see Table B4 for details). The EE-1 filter removed 29 cases due to overlap with other plumes. The injection height from the GFAS database was not available for 57 cases (EE-2) and 29 plumes failed because overlapping with other plumes (EE-1). Furthermore, particles-plume alignment filter, EE-3, removed a total of 51 cases. This is mostly due to bad can be attributed to poor plume detection, inaccurate velocities, or injection heights. LastlyFinally, the velocity filter, EE-4rejects about, rejected a total of 15 cases. Finally, the automated plume selection can be verified with a

As a final step, the 226 cases were verified by visual inspection. For the US region difference are small by construction as the threshold values were tuned for this data ensemble. For the verification areas Australia and Siberia we see similar We can see a good performance of the detection algorithm, and in total only 5 algorithm, as only five of the 226 plumes were flagged as false identifications cases were wrongly identified. Fig. 6 gives shows three examples of the false positive detection and cases.

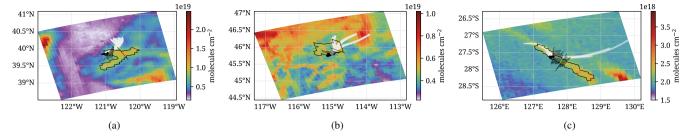


Figure 6. False positive detection of plumes Falsely identified cases by the APE algorithm. Black dots indicate the fire counts, white bands the tracer particles and the black polygon depicts the detected plume, respectively. The black dashed lines are transaction lines.

Thus, our analysis confirms the applicability of our algorithm to other areas with a confidence of based on the above analysis, we can conclude that 97.7% of the cases produced by the algorithm are good.

Overall, APE incorporates strict data filtering, which is mainly driven by the TROPOMI detection limit and data quality due to cloud coverage. However, this should not distract from the fact that TROPOMI is the first instrument that shows these emission features in satellite data. Secondly, the data yield is thinned out further by selecting data which are appropriate for the current APE inversion scheme, which are fire emission events by isolated single sources. Overall, we consider that there will be sufficient data yield for a new TROPOMI CO data product when applied to more than 6 years of global observations.

3.1 Cross-Sectional Massflux method (CFM)

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The CFM computed emissions for 221 cases. To compare the effect of plume heights, two variables are were defined, namely, the mean plume height \overline{z}_{lag} , which is the mean of z_{lag} of all transects along the downwind direction of the plume, and the maximum rise in plume height (δz) with respect to z_c . They are given as

$$\overline{z}_{\text{lag}} = \frac{1}{m} \sum_{i=0}^{m} z_{\text{lag},i} \tag{8}$$

$$\delta z = \max\{z_{\text{lag}}\} - z_c \tag{9}$$

The Figures 7 (a) and (c) show the mean plume height \bar{z}_{lag} is plotted against the constant plume height (z_c) in Figure 7(a) and (c) for US and for the United States, Australia, and Siberian region, respectively. The the Siberian region δz decreased and increased in the downwind direction for about 43 and 178 fires, respectively. Additionally, the Furthermore, δz in the downwind direction is was found to vary > 500 m for 30 fires in regions encapsulating Australia and US as represented Australia and the United States, as shown by the orange color colour in Figure 7 (a). However, no such cases were found in Siberia (see Fig. 7(c)). Among these 30 fires, about 11 fires had $\delta z > 1000$ m. As the This increase in plume height in the downwind direction can be attributed to the rising warm air, which may be heated by the fire. Furthermore, this heating can be related to total fire radiative power (FRP) and fire counts represent, since they describe the heat generated and the area burnt, a relation between plume height rise and these two variables can be expected as higher FRP means higher temperature which heats up the air, leading

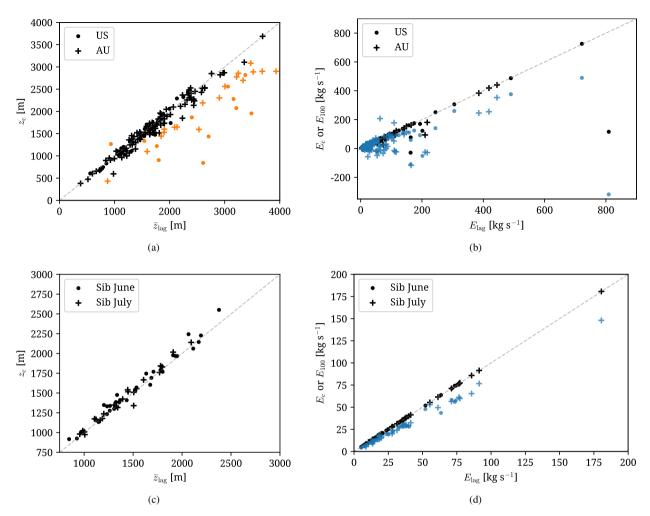


Figure 7. Plume height variation and emissions for regions encapsulating US and Australia (top figures) and Siberia (bottom figures). (Panel a and c) The mean plume height (see Eq. (8)) versus the constant plume height for each fire. The black color colour represents $\delta z \leq 500$ m, and the orange color colour indicates $\delta z > 500$ m (Panel b and d). Comparison between the emissions computed at plume height z_{lag} versus z_{c} represented by black color colour and z_{lag} versus a constant plume height of 100 m represented by blue color colour.

rise of the warm airburnt area, respectively. However, no such relation was observed and, as there were cases with low FRP, or low fire counts where the, where $\delta z > 1000$ m and vice versa. Additionally, it was difficult to find an appropriate reason for large rise challenging to find a suitable reason for a large increase in plume height in the downwind direction. Obviously, this plume height variation can influence the emissions due to the change in the velocity with height.

Figure Figures 7(b) and (d) compare the emissions computed from the Lagrangian plume height (E_{lag}) with the emissions computed from the constant plume height (E_c) represented by blackcolor and in black, and the 100 m plume height (E_{100}) represented by bluecolor. Combination in blue. A combination of all cases in figures Figures 7(b) and (d) show that the shows

that E_c varied less than 10% from the E_{lag} for a total of 198 cases. And the fires in Siberia varied by For Siberia fires, the corresponding variation is less than 4%. However, 23 cases in US and AU had variation larger than the United States and Australia show differences of >10%. Thus, the overall effect of the Lagrangian plume height to the constant plume height on the emission estimate is considered minor, however. However, we could identify several cases where the emissions estimate from Lagrangian plume height become becomes more reliable. For example, a US fire (black eolorcolour) on the bottom right of Figure 7(b) was found to have a high $E_{lag} = 809 \text{ kgs}^{-1}$ and low $E_c = 115.9 \text{ kgs}^{-1}$. The total fire radiative power (FRP) for this case was found to be the highest among all the detected plumes detected plumes, and the burnt area (number of fire counts in VIIRS data) was the third highest among all the detected cases. Additionally, the The CO enhancement was large, thus, also large, so a high emission estimate is expected. Furthermore, the a high FRP is correlated with higher temperatures, so an increase in the plume height in the downwind direction is normal. It should be noted that the Lagrangian simulations do not consider the heating. However, heating but we assume that the velocities meteorological data (velocities) in ERA5 cover this, as it assimilates the surface temperaturetemperatures. The increase in plume height is observed in the Lagrangian simulations as the z_{lag} increases by 1350 m compared to the constant plume height in the downwind direction at 32.5 km from the fire source. From this, one we can conclude that the E_{lag} ear may be more appropriate than E_c . A similar Similar reasoning can be used to explain why the E_{lag} was higher compared to E_c for the fires where the FRP on average was higher.

Figures 7(b) and (d) also compare emissions from Lagrangian plume height to a constant 100 m plume height. We considered 100 m plume height as three-dimensional velocity fields, which are required to compute the CO emissions based on the plume heights z_c and z_{lag} , amount amounting to a large quantity of data. Furthermore, computing emissions by scaling 100 m winds would simplify the the approach to a large extent. However, we found no correlation between the difference in the emissions $(E_{100} - E_{lag})$ or E_c and the variation in plume heights. Additionally, a total of 37 fires were found to have negative values for E_{100} due to a negative velocity at 100 m. This makes it difficult challenging to find an appropriate scaling to obtain emissions at z_{lag} from the velocities at 100 m. Thus, thus highlighting the importance of using three-dimensional velocity fields rather than surface near wind surface-near-wind fields at a fixed altitude. From all these observations, we conclude that the varying plume height is more reliable to compute emissions by an automated algorithm.

It should be noted that the emission estimations in this paper are not compared to any existing database as we are not aware of any in-situ studies except by Rowe et al. (2022) which compares the CO column from aircraft measurements to TROPOMI data.

420 3.2 Emission uncertainty

We estimate two

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Sherwin et al. (2023) validated satellite CH_4 data using controlled emission releases of point sources of methane for detection and quantification. For CO, such validation is not possible for single-point releases. (Rowe et al., 2022) have shown that the integrals of TROPOMI CO data along the plume transects were $\approx 7.2\%$ higher than the aircraft measurements after corrections for a few fires in the US. However, they do not report emissions. Thus, to the best of our knowledge, no independent, high-quality estimate of CO point-source emission is currently available. Hence, to demonstrate the data quality and enhance

Table 2. Maximum values of standard error and emission uncertainties due to plume height for different regions among all fires.

Region	$\sigma_{ m E}(z_{ m lag})\%$	$\sigma_{ m E}(z_{ m c})\%$	$\Delta E_{\text{lag}}^p \%$	$\Delta E_{\mathrm{lag}}^{m}\%$	$\Delta E_c^p\%$	$\Delta E_c^m\%$
US	15.11	35.51	71.28	82.27	246.93	163.54
AU	18.79	28.10	94.78	95.41	130.17	170.37
Sib Jun	18.21	18.13	17.34	14.95	17.37	8.84
Sib Jul	19.72	19.57	14.88	13.12	15.47	12.1

confidence in our data product, we instead perform a detailed uncertainty analysis disentangling (quasi-)random and systematic errors.

We estimate three different contributions to the uncertainty of the estimated emission, where we assumed constant emissions in time for each plumeconstant emissions over time are assumed for each fire case. First, we consider the relative variation of the CO fluxes through the different intersects Q_i is considered. Different error sources may cause this variation, and the corresponding error on in the flux estimate can be characterized characterised by the standard error σ_E in Eq. 4. Second, errors due to random uncertainties in ERA5 velocity fields are addressed. Finally, systematic errors that affect the different fluxes Q_i equally cannot be address by this approach, and so require a dedicated discussion require a different approach. These errors are TROPOMI bias CO column biases, ERA5 velocity bias, and injection height uncertainty. In addition to these two errors To complete our error classification, we verify the emission uncertainty in APE, i.e., the standard error, the APE algorithm using data from WRF simulations, where the wind velocities, CO and injection height are known.

3.2.1 Standard errors

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The standard error encompass of the emission estimate encompasses various uncertainty sources, e.g., the interpolation error due to the under-sampling of the CO field by TROPOMI (shown in Fig. 4(c)), relative wind errors with respect to the ERA5 velocity fields, the precision of the TROPOMI CO data, the uncertainty variation in defining the atmospheric CO background per intersect, and the temporal variation of the emission around its mean. $\sigma_{\rm E}$ does not allow us to disentangle these error sources, expect except for the TROPOMI CO precision. This is < 10%, which is specified for every TROPOMI observation. Overall, the precision of the CO column is < 10% per pixel, even for dark scenes over land (Landgraf et al., 2016). For the flux estimate, this yields an a negligible error contribution. To compare the standard error for different fires, TableTab. 2 reports the maximum relative standard error for the four regions using the Lagrangian plume height and the constant plume height z_c . For individual fires errors The error for individual fires can be accessed in the data base database (Goudar et al., 2023). The data shows show that the maximum standard error for the Lagrangian plume height is significantly smaller than for the constant plume height z_c for both the US and Australiadata. This is another indication to use the Lagrangian plume height height of the plume does not vary much as depicted in Figure. 7c. Overall In general, the standard error of the emission estimate is < 20%.

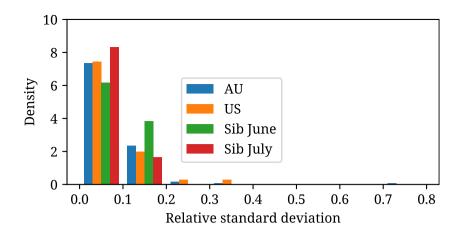


Figure 8. Histogram plot showing density vs relative standard deviation of emissions from 10 ERA5 ensemble velocity fields for four regions.

3.2.2 Other errorsERA5 uncertainties

ERA5 ensemble data (Hersbach et al., 2017) is used to quantify velocity uncertainties. The ensemble includes ten members, and the variation between the members represents random errors, but not systematic errors (Hersbach et al., 2020). Due to the small size of the ensemble, the data cannot encapsulate all the random uncertainties. For every member j of the ensemble, the Lagrangian plume height (Sec. 2.3.3 is used to calculate the emission (E_{lag}^j) for all plumes. Subsequently, the relative standard deviation σ_{vel}/E_{lag} per emission source was computed, where σ_{vel} represents the standard deviation of the emissions of the ten members of the ensemble. Figure 8 shows the density histogram versus the relative standard deviation for all regions. A total of 215 cases among 221 cases have a velocity uncertainty of less than 20%. One Australian case has an uncertainty > 70%, which was due to a single ensemble whose velocities were ≈ 3 times higher than the other 9 ensembles. Although the ensembles do not fully describe the random errors, we observe less than 20% uncertainty in 97% of the cases.

3.2.3 Systematic errors

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One potential error which that cannot be addressed with the standard error is an overall bias in the TROPOMI CO product. Borsdorff et al. (2019b) reported a CO bias of 3.4 ppb for the TROPOMI product compared to the Total Carbon Column Observing Network (TCCON). This corresponds to an a typical relative error < 1.7% for a plume concentration of about 200 ppb in a plume. Assuming the CO bias to be constant over the plume, it yields the same relative error on the emission estimate and so can be neglected Rowe et al. (2022) showed that the integrals of TROPOMI CO data along the plume transects were \approx 7.2% higher than the aircraft measurements after corrections for a few fires in the US. Assuming a worst-case scenario, the constant bias of 7.2% over the plume leads to \approx 7.2% higher emission estimation.

Another error of in this category is the emission uncertainty because of due to the uncertainty in the IS4FIRES injection height of ± 500 m Sofiev et al. (2012). For each fire, we calculate the emission uncertainty

$$\Delta E_{\text{lag}}^{p/m} = \left| \frac{E_{\text{lag}}^{p/m} - E_{\text{lag}}}{E_{\text{lag}}} \right| \tag{10}$$

using plume heights z_{lag}^p and z_{lag}^m , respectively (see Fig. 5b). Analogously, the uncertainties ΔE_{c}^p and ΔE_{c}^m for E_{c} , are computedand the maximum number per region is again given in Table 2 are computed. The uncertainties changes change from fire to fire and can be found in the data (Goudar et al., 2023). Table 2 shows the value of the largest uncertainty per region. For the Siberian region, the maximum uncertainties are small, indicating little vertical variation in the velocity. For US and Australian regions, the corresponding uncertainties are much larger where the uncertainties for and the uncertainties corresponding to the constant emission height exceeds those using the Lagrangian simulation by a factor of are a factor 2-3 times higher than the Lagrangian plume height uncertainties. This hints at a more variable wind field for these regions. Overall, we estimate that this APE error term to be is the largest error contribution with error an error < 100% for each fire.

Finally, we consider errors in the systematic errors in wind velocity that are constant over in the plume domain. The error propagate propagates one-to-one into the error of the flux estimate. Uncertainties of the ERA5 wind fields in the tropospheric boundary layer are not reported. Gualtieri (2022) derived surface near wind errors of 1.76 m/s (root-mean-square ms⁻¹ (root mean square error) for ERA5 data. A typical wind speed at the plume height is 3-11 m/s-ms⁻¹, and although at the plume height, the wind speed error of 1.76 ms⁻¹ might be smaller, we consider this error as a significant error contribution. However, we refrain from quantifying this error because of due to the lack of reliable the knowledgeon the windknowledge.

3.2.4 Verification of emission Emission uncertainty in APE

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We verify our uncertainty estimates by evaluating WRF simulations of a CO plume using APE. The WRF simulation was performed using real atmospheric forcing at 1 km resolution for a fire with the highest FRP (USA, September 12, 2020; see Sec. 3.1). The details on of the WRF simulation can be found in the Appendix Appendix A. Three plumes at three different UTC times shown in Figure 9 a-c were selected, and emissions were estimated by our algorithm. It should be noted that the averaging kernels were not used to degrade to TROPOMI data, and only the enhancements were simulated in the model, thus the background is set to zero by the simulation. The plume height (z_{lag}) was computed calculated as the maximum height where the concentration became zeroand, and the fire sources were the same as the sources used in the WRF simulation. The velocity used in both Lagrangian simulations and emission estimations was inferred from the WRF velocity data. The emissions for these plumes were estimated by APE assuming a constant emission in time and are presented in Table 3. Here, the actual emission is the mean of the known total total known CO emission from all fire sources with time. This considered time is based on the release time of the particles around The averaging interval is defined as the time the particles take to reach the final transectused to compute emission. Additionally, we degraded the simulation grid to the TROPOMI grid shown in Figs. 9e-g.

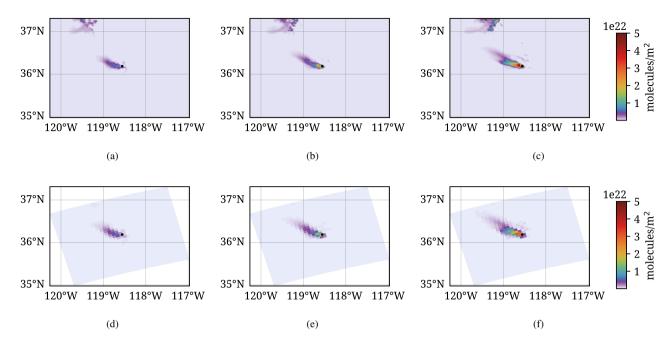


Figure 9. Three selected plumes at three different UTC times (a, d) at 17:00, (b, e) at 18:00 and (c, f) at 19:00. Top (a-c) represent plumes at 1 km grid resolution and bottom (d-f) represent TROPOMI grid resolution.

Table 3. Comparison of actual emissions to the emissions computed at plume height z_{lag} for the three selected plumes shown in Fig. 9. The uncertainty in the table is computed as 100*(Actual - Computed)/Actual.

Time in UTC (H:M)	Actual (kgs ⁻¹)	1 km grid (kgs ⁻¹);	TROPOMI grid (kgs ⁻¹);	
		Uncertainty	Uncertainty	
17:00	28.45	20.26; 28.8%	20.67; 27.3%	
18:00	56.84	34.92; .5 38.5%	34.52; .5 <u>37.5</u> %	
19:00	97.86	99.15; -1.5%	99.36; -1.53%	

The uncertainties of the APE emission estimate ranges between range between -1.5% and + and 38.5% (Table 3). In all three plumes, the velocity and plume height used by APE are is appropriate, however, the emissions computed by our algorithm differed from the actual emissions. This is attributed to the error in the cross-sectional flux method due to the assumption of a constant emissions constant emissions, which might not be the case for a fire. It should be noted that this uncertainty is for one particular case, and it can vary depending upon on the case. For the three selected cases, the CFM method leads to an error of 28.8, 38.5 and 1.5% which is in the range of the derived standard error. The difference in emissions between high resolution (1 km grid) and low resolution (TROPOMI grid) was found to be less than 2%. If the velocity is accurate, then it can be concluded that having higher resolution higher-resolution data does not have much effect on the cross-sectional flux

method. Overall, this analysis suggests that the assumption of a constant emission is the major error source next to errors in the wind field and uncertainties in the injection height.

4 Conclusions and recommendations

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An automated plume detection and emission estimation scheme for CO flux inversion for single point fires was developed by integrating four freely available data sources, such as,: the VIIRS active fire dataset, the TROPOMI CO dataset, the injection height from GFAS and ERA5 meteorological data. The automated plume detection and emission estimation algorithm (APE v1.1) was optimized optimised for one region, and its performance is verified for three months of data for two other regions, Australia and Siberia. For all the regions, 1327 fires had regions and for all the fire sources identified by VIIRS, 16% (882 cases) of the data correspond to clear sky TROPOMI CO data of sufficient quality among 5562 identified fire clusters. Even though with a plume signature. Out of those 882 plumeswere identified in 1327 cases, only 378 plumes were considered as, 309 plumes were too short and about 195 had multiple sources of fire in them. Lastly, the emissions were estimated for only Internal quality filtering of APE reduced the number of estimated fires to 226 cases among 378 plumes. We can conclude that APE can reliably detect and estimate emissions automatically for 97.7 which is 26% of the eases.

The key to automatically detect 882 cases. Finally, the visual filter on APE's output of 226 cases showed a true-positive confidence level of 97.7% (221 cases). One key element of automated APE detection of fire plumes in the TROPOMI CO dataset and estimate emissions was knowing the fire source a priori. In the present study, is prior knowledge of potential fire locations coming from the VIIRS active fire data product from Suomi NPPhelped in identifying fire source which was used to detect plumes in TROPOMI CO data. This . It highlights the potential for flying the satellites to fly the Suomi NPP and SP5 in the same orbit and in satellites in a loose formation with a temporal separation of 3.5 minutes to identify fires.

We To estimate CO fire emissions, we employed the CFM. Here, we considered three different assumption assumptions on plume heights, first a constant plume height at 100 m altitude, second a constant plume height at the GFAS injection height, and third a varying plume height using a Lagrangian model. The varying plume height was chosen based on approach best reflects the characteristics of a fire. If a fire is at its peakthen the airaround is heated and rises in the atmosphere, strong convection leads to an upsurge of air, and at the same time, it will be transported away is transported downwind from the fire source. Thus, one can observe the CO rise with distance from the source. It should be noted Note that we assume that the ERA5 velocity fields incorporate this heating effect to a certain extentsome extent, as it assimilates the surface temperature from observed by satellites. In our simulations, the plume height varied by more than 500 m in a downwind direction for 30 eases out of 221 cases, and all 30 cases were in US the United States and Australia. The plume height variation variation in plume height was found to be minimal in Siberia.

The assumption of plume height at 100 m led to unreliable emission estimates and was discarded. The difference in estimation emission for the constant injection height from GFAS and of the GFAS and the varying plume height, was observed to be less than 4% for the Siberian region. Larger differences we observed We observed larger differences for the US and Australia, where the maximum uncertainty using the varying plume height is half of that using a constant plume height. However, for

many plumes the difference in the uncertainty estimate was minor. Based on this Based on these findings, we decided to use the Lagrangian model for the plume height as processing baseline of the processing baseline for APE.

Overall, we estimated estimate the uncertainty of our product with the standard error a standard error of < 20%, which mainly accounts for errors in the spatial under-sampling due to spatial undersampling of the CO field by TROPOMI and the assumption of a constant emission for the time frame relevant for the to plume formation. The TROPOMI CO data is are of high quality regarding with respect to precision and biasand does. Based on TCCON, the TROPOMI CO data do not provide any significant contribution to the CO emission estimate of APE. Additionally, we analyzed analysed emission errors due to the GFAS uncertainty in the injection height uncertainty of injection height from GFAS. Depending on the meteorological situation at in the different regions, errors are < 100%. Errors due to wrong wind information is also considered to be important but could not be specified. The random error in the meteorological data (wind velocities) was described using the ERA5 ensemble data and was found to be less than 20% for 97% of the cases. Systematic errors due to the wind for every fire case were also considered important, however, they cannot be specified, as the ERA5 data product does not provide an uncertainty estimate of the provided wind field at the plume heightestimate of systematic wind errors.

Finally, for the first time the presented the presented APE algorithm is appropriate to estimate for estimating CO emissions from single isolated fires from TROPOMI/VIIRS data by using a fully automated algorithm. It is considered as a baseline for future APE upgrades to optimize optimise automated emission estimates of CO point sources. As a next step, we consider (1) the processing of the entire CO TROPOMI data set, (2) expanding emission estimations for multiple fire sources and develop developing an improved inversion scheme. This can, which can be done by developing algorithms that maps map the simulated tracer particles from Lagrangian simulations to the TROPOMI CO concentrations to compute emissions , and (4) verifying comparing the emissions predicted by APE to available in-situ emissions emission databases.

Appendix A: The WRF Model description

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The WRF model configured in a two-domain configuration is applied in the tracer mode to simulate the transport and dispersion of CO emitted by a wildfire in the US. The outer and inner domains are run at a horizontal grid spacing of $5 \times 5 \text{ km}^2$ and $1 \times 1 \text{ km}^2$, respectively. The model domains are centered centred at 36.16225^0 N , 119.1528^0 E and have 43 vertical levels stretching that extend from the surface to a model top of 50 hPa. The outer domain has $200 \times 200 \text{ grid points}$, while the inner domain has $400 \times 350 \text{ grid points}$ in the west-east and north-south directions. The meteorological initial and boundary conditions for the outer domain are based on the Global Forecast System (GFS) forecasts available every 3 hours at a horizontal grid spacing of $0.25^0 \times 0.25^0$. The static geographical fields and the GFS output are mapped onto the WRF domains using the WRF preprocessing system (WPS). The physical parameterizations follow Kumar et al. (2021)except the cumulus parameterization except for the cumulus parameterization that is turned off in the inner domain.

Biomass burning emissions are obtained from the Fire Inventory from NCAR NCAR Fire Inventory (FINN; Wiedinmyer et al., 2011) version 2.5 and are distributed vertically online using a plume rise parameterization parameterization developed by Freitas et al. (2007). This parameterization parameterization selects fire properties appropriate for the land use in every grid box

containing fire emissions and simulates the plume rise explicitly using the environmental conditions simulated by WRF. Since we are using the model in the tracer mode, the chemical evolution of the plume is not simulated. To describe the loss of CO in the model, we allow the CO fire emissions to decay with an e-folding lifetime of 30 days. No other source (anthropogenic emissions, biogenic emissions or photo-chemical production from hydrocarbons) is included in the simulation. The model run started on 12 Sep September 2020 at 12 UTC and stopped at on 13 Sep September at 00 UTC. We used a time step of 20 s for the outer domain and 4 s for the inner domain. The model output is saved every min-minute and used for further analysis.

Appendix B: Algorithm and Simulation details

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Algorithm 1 APE algorithm: Pseudo-code

Require: region and time do

Find Fire sources from VIIRS data (Sec. 2.1.1)

for Each fire source do

Extract TROPOMI CO data granule (Sec. 2.1.2)

if Data is good then

Detect plume by plume detection algorithm (Sec. 2.2)

if Plume is detected then

Estimate emission (Sec. 2.3)

end if
end if
end for
```

Table B1. Considered region and time. The region is rectangular and is constructed based on the origin and the width and height. The origin is always the south-western point of the region.

Label	Region Origin [(lon, lat)]	Region Size	Time	L2 product version	VIIRS, ERA5 and
		(Width, Height)			GFAS Data access
US	140°W, 20°N	80°, 45°	Sept 2020	1.03.02	10 Oct 2020
AU	70°E, 53°S	55°, 27°	Oct 2019	1.03.02	10 Oct 2020
Sib	113 ⁰ E, 44 ⁰ N	41°, 34°	June 2021	1.04.00	5 Feb 2023
Sib	113 ⁰ E, 44 ⁰ N	41°, 34°	July 2021	2.02.00	5 Feb 2023

Code availability. APE v1.1 code is archived on Zenodo (https://doi.org/10.5281/zenodo.7361559).

Table B2. Filtering from fire clusters to good CO data. The total column is the same as the fire clusters in Table 1.

Region	Grid size	Quality	Multiple Clusters	Good Data	Total
US	442	373	53	213	1081
AU	1020	512	87	385	2013
Sib Jun	37	249		130	416
Sib Jul	34	1419		599	2052
All Regions	1533	2553	140	1327	5562

Table B3. Filtering from good data to plume detection Table 1. The total should represent the good CO data available.

Region	No enhancements	Short plumes	Other clusters	Detected plumes	Total
US	42	41	51	79	213
AU	57	62	94	172	385
Sib Jun	22	25	12	71	130
Sib Jul	324	181	38	56	599
All Regions	445	309	195	378	1327

Table B4. Filtering from plume detection to emission estimation Table 1. The total should represent the plume detection cases.

Region	No Injection	Background	Plume	Velocity	Emission	Total
	height	Subtraction	alignment	<2 ms^-1 ms ⁻¹	estimation	
				ms^{-1}		
US	14	14	11	3	37	79
AU	20	4	13	6	129	172
Sib Jun	15	2	15	4	35	71
Sib Jul	8	9	12	2	25	56
All Regions	57	29	51	15	226	378

Data availability. The TROPOMI CO dataset of this study is available for download at ftp://ftp.sron.nl/open-access-data-2/TROPOMI/tropomi/co/ (last access: 5 Feb October 2023). The IS4FIRES injection height and the 3-d velocities at 127 model levels were obtained from the Global Fire Assimilation System (GFAS) database and the European Center for Medium range Weather Forecasts (ECMWF) Reanalysis v5 (ERA5), respectively on 10 October 2020. The Visible Infrared Imaging Radiometer Suite 375m thermal anomalies/active fire product was also accessed on October 2020 (https://firms.modaps.eosdis.nasa.gov/active_fire/). The processed data is available as DOI:10.5281/zenodo.7728874.

Author contributions. MG developed the code and performed the analysis with the inputs from TB and JL. JA master's thesis served as a feasibility study for this work. RK performed WRF simulations and gave inputs on its data analysis. All co-authors commented and improved the paper with a special mention to JL and TB.

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