

1 **Can the combining of wetlands with reservoir operation largely reduce the risk of**
2 **future flood and droughts?**

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12
13 **Abstract.** Wetlands and reservoirs are important water flow and storage regulators in a river basin;
14 therefore, they can play a crucial role in mitigating flood and hydrological drought risks. Despite the
15 advancement of river basin theory and modeling, our knowledge is still limited about the extent that
16 these two regulators could have in performing such a role, especially under future climate extremes. To
17 improve our understanding, we first coupled wetlands and reservoir operations into a semi-spatially
18 explicit hydrological model and then applied it in a case study involving a large river basin in Northeast
19 China. The projection of future floods and hydrological droughts was performed using the hydrological
20 model during different periods (near-future: 2026-2050, mid-century: 2051-2075, and end-century:
21 2076-2100) under five future climate change scenarios. We found that the risk of future floods and
22 hydrological droughts can vary across different periods, in particular, will experience relatively large
23 increases and slight decreases. This large river basin will experience longer duration, larger peak flows
24 and volume, and enhanced flashiness flood events than the historical period. Simultaneously, the
25 hydrological droughts will be much more frequent with longer duration and more serious deficit.
26 Therefore, the risk of floods and droughts will overall increase further under future climate change even
27 under the combined influence of reservoirs and wetlands. These findings highlight the hydrological

28 regulation function of wetlands and reservoirs and attest that the combining of wetlands with reservoir
29 operation cannot fully eliminate the increasing future flood and drought risks. To improve a river basin's
30 resilience to the risks of future climate change, we argue that implementation of wetland restoration and
31 development of accurate forecasting systems for effective reservoir operation are of great importance.
32 Furthermore, this study demonstrated a wetland-reservoir integrated modeling and assessment
33 framework that is conducive to risk assessment of floods and hydrological droughts, which can be used
34 for other river basins in the world.

35 **Keywords:** Climate change; Hydrologic projection; Floods and droughts; Wetland hydrological
36 services; Reservoir operations; Model integration

37

38 **1. Introduction**

39 Floods and droughts have produced some of the most frequent and serious disasters in the world
40 (Differbaugh et al., 2015; Hirabayashi et al., 2013; UNISDR, 2015). Globally, they account for 38% of
41 the total number of natural disasters, 45% of the total casualties, more than 84% of the total number of
42 people affected, and 30% of the total economic damage caused by all-natural disasters (Güneralp et al.,
43 2015) in the past. As climate change has been accelerating the hydrological cycle, causing more frequent
44 and stronger weather extremes, more floods and droughts have been projected to increase at both global
45 (Chiang et al., 2021; Jongman, 2018) and regional scales (Hallegatte et al., 2013; Wang et al., 2021).
46 Concurrently, the disaster-related loss of ecosystems (e.g., wetlands, forest, and grassland) and their
47 services can mitigate the flood and drought risks to a great extent (Gulbin et al., 2019; Walz et al., 2021).
48 Given this, grey infrastructure such as dams, dikes, and reservoirs, which have often been used to
49 attenuate flood and drought hazards because of their rapid and visible effects, can play an important role
50 in ensuring the water security of a river basin (Alves et al., 2019; Casal-Campos et al., 2015). However,
51 relying solely on grey infrastructure to attenuate floods and droughts has some inadequacies, such as
52 large investments to build and maintain in addition to adverse effects on downstream ecosystems (Maes
53 et al., 2015; Schneider et al., 2017). In this context, Nature-based solutions (NBS) for hydro-

54 meteorological hazards mitigation are becoming increasingly popular (Kumar et al., 2021), because
55 NBS can effectively reduce or even offset the hydrological processes driving floods and droughts (Nika
56 et al., 2020), while making least disturbance to the environment as well as delivering co-benefits which
57 grey infrastructure cannot provide (Anderson and Renaud, 2021; Nelson et al., 2020). Therefore, it is
58 urgent to integrate NBS into the current water management practices to increase basin resilience to
59 hydrological extremes under climate change.

60 Wetlands have the potential to be used as a NBS for improving water storage and hence the resilience
61 of a river basin to hydrological extremes along with grey infrastructures (Thorslund et al., 2017). This
62 is because, similar to man-made dams and reservoirs, wetlands can attenuate flow and alter basin
63 hydrological processes (Lee et al., 2018), such as floods (Wu et al., 2020a) and baseflows (Evenson et
64 al., 2015; Wu et al., 2020b). However, unlike man-made grey infrastructures, wetlands are integral in
65 landscapes and they are connected laterally and vertically with the surrounding terrestrial and aquatic
66 environments through the hydrological cycling of water and waterborne substances (Åhlén et al., 2020),
67 making their water storage and cycling fundamental to estimate a watershed's water balance (Golden et
68 al., 2021; Shook et al., 2021). To understand how and to what extent wetlands can mitigate hydrological
69 processes, two approaches are commonly used: (i) description of individual wetland service at the field
70 scale (e.g., Park et al., 2014) or wetland landscape scale (e.g., Åhlén et al., 2022); (ii) assessment of wetland
71 hydrological services at the regional/watershed scale (Fossey et al., 2016; Wu et al., 2020a, 2020b).
72 However, the former approach only be achieved with field instruments and is mainly used to provide
73 key parameters of wetland processes for model calibration (Fossey and Rousseau, 2016). Recently,
74 several wetland modules have been development and coupled to hydrological (e.g., Soil and Water
75 Assessment Model, HYDROTEL model) to quantify hydrological function of wetlands, particularly the
76 mitigation services on floods and droughts (Evenson et al., 2018; Evenson et al., 2016; Fossey et al.,
77 2015a; Zeng et al., 2020). These wetland hydrological models not only consider the general water budget
78 of a river basin but also consider the perennial and intermittent hydrological interactions between
79 wetlands-to-wetlands and wetlands-to- surrounding landscapes. It is of both scientific and practical
80 interest to project wetland capability in mitigating floods and droughts in response to a changing climate.

81 Reservoirs redistribute large amounts of surface water, thus altering natural hydrological processes,
82 such as flow range, flood and drought patterns, and basin water balances (Boulangue et al., 2021; Chen
83 et al., 2021; Manfreda et al., 2021; Zhao et al., 2016). So far, throughout the world, there are 57, 985
84 reservoirs registered by the International Commission on Large Dams and their total volume has been
85 reached 14, 602 km³ (Eriyagama et al., 2020). Such numerous reservoirs and their large storage capacity
86 should not be neglected in water hazard assessment and hydrological projection because of their
87 significant modification on flood and drought patterns (Boulangue et al., 2021; Brunner et al., 2021). For
88 that reason, scholars called for the need to integrate reservoirs in model-based impact analysis of flood
89 exposure under climate change (Dang et al., 2020a; Yassin et al., 2019). Therefore, there is a growing
90 need in incorporating reservoir operations into basin hydrologic simulations and predictions.

91 Despite the well-established knowledge of flow regulation and water storage functions that wetlands
92 and reservoirs can provide in a river basin, most modeling assessments on floods and droughts at the
93 basin scale do not take the two components into account, or give little emphasis on the combined benefits
94 of them (Brunner et al., 2021; Golden et al., 2021). Nor are the hydrological processes associated with
95 these features implicitly included in the calibration of hydrologic models. Recent studies have suggested
96 that disregarding of the wetlands or reservoir operation would add significant error and larger
97 uncertainties to simulate hydrologic processes (Brunner et al., 2021; Ward et al., 2020). Because
98 wetlands are often abundant across many landscapes, making their water storage and cycling
99 fundamental to estimating a watershed's water balance (Rains et al., 2016; Lee et al., 2018). Therefore,
100 missing this component of water balances could potentially lead to disproportionately large model errors
101 (Rajib et al., 2020). Consequently, integrating the wetlands (Fossey et al., 2015a; Golden et al., 2021;
102 Rajib et al., 2020) or reservoir operation (Dang et al., 2020; Yassin et al., 2019; Zhao et al., 2016) alone
103 into watershed-scale hydrologic models may largely minimize uncertainties and improve model
104 performance. Furthermore, on a global scale, most river basins have wetlands and their river flow has
105 or will experience reservoir regulation (Muller, 2019; Schneider et al., 2017), which elicits a thought-
106 provoking concerns: What will be the changes of future floods and droughts under the combined
107 influence of wetlands and reservoirs? Such concern is important because the omission of wetlands and

108 reservoirs can cause the policy-making process to be imprecise at best and ineffective at worst. However,
109 a reservoir operation and wetland services, integrated basin-scale model rarely exist in the literature.
110 Furthermore, although few studies (e.g., Rajib et al. (2020; Chen et al., 2021; Wu et al., 2021) provide
111 insights into modeling and understanding the flow regulation functions provided by wetlands and
112 reservoirs, however, is it still unclear whether the combining of wetlands with reservoir operation can
113 largely reduce the risk of future floods and droughts.

114 Considering the above-introduced scientific challenges and management deficiencies, we first
115 developed a framework of hydrological modeling coupled with wetland modules and reservoir operation
116 scenarios. We then applied it to a large river basin with abundant wetlands and a large reservoir, the
117 Nenjiang River Basin in northeast China, to address a central question: Can the combining of wetlands
118 with reservoir operation largely reduce the risk of future flood and droughts? The Nenjiang River Basin
119 was selected as a case study here because it has abundant wetlands and a large reservoir, and has
120 undergone intensive anthropogenic activities in the past half century, particularly in the increasing
121 agricultural water consumption and conversion of wetlands to agricultural and other land uses. Our
122 framework and results are expected to bring new insights into future floods and droughts and provide a
123 basis for decision-making to curb the growing impacts of unprecedented and future hydrological
124 extreme conditions.

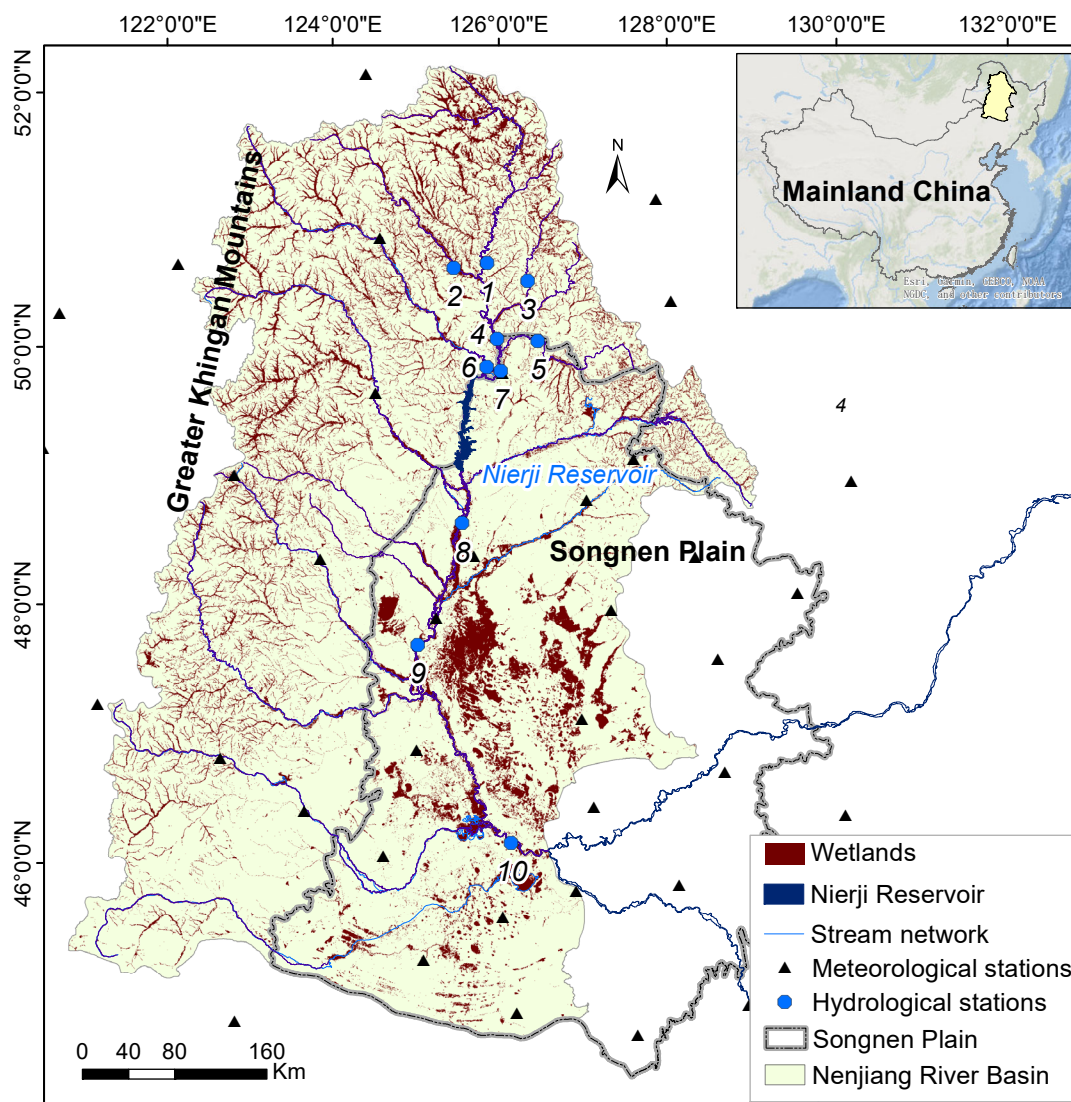
125 2. Methodology

126 2.1 Study area and datasets

127 We conducted this analysis in the Nenjiang River Basin (NRB), a large river basin (291,700 km²)
128 located in the Northeast China (Fig. 1). Long-term annual average runoff depth and volume from the
129 NRB are 97.4 mm and 22.7 billion m³. The river basin is located in the middle-high latitudes and can
130 be characterized by a temperate semi-humid continental monsoon climate. Inter-annual differences in
131 temperature and precipitation are large, i.e., disparate hot and cold periods, and uneven dry and wet
132 conditions (Meng et al., 2019). The average annual temperature across the basin ranges between 2.1-
133 4.5°C. The annual total precipitation within the basin fluctuates from 323.1 to 537.6 mm. Precipitation

134 is mainly concentrated during June-September, which accounts for about 85% of the annual
135 precipitation (Li et al., 2014).

136 The NRB is one of the pivotal wetland areas in China. The basin contains several important wetland
137 conservation areas, among which Zhalong and Nanweng River Wetlands have been designated as a
138 Ramsar Site of International Importance. The wetlands and their contributing drainage areas (see
139 Section 2.2.1 for specific definition) within the subbasins monitored by the ten hydrological stations
140 range from 14 to 23% and from 39 to 56% respectively, demonstrating the large wetland coverage of
141 the NRB and its sub-basins (Table 1). The lower NRB is an important agricultural area of the Songnen
142 Plain, which is one of the three major plains (including the Sanjiang, Songnen and Liaohe Plains) in
143 northeast China. Therefore, understanding potential floods and hydrological droughts under future
144 climate change is crucial for ensuring regional food security and wetland ecological integrity. During
145 the past 60 years, land use and land cover types have drastically changed owing to large-scale
146 development of intensive agriculture and water resources management (Meng et al., 2019). The area of
147 wetlands in the NRB decreased by nearly 23% from 1978 to 2000 (Chen et al., 2021), with only 16.34%
148 remaining today (Table 1), which largely degraded their services (Wu et al., 2021). Along with the
149 reduction in wetland area, the hydrological functions of wetlands in the NRB, such as water storage,
150 flood mitigation and baseflow support, have been considerably reduced (Wu et al., 2021). These wetland
151 services are closely related to flood and drought risks, such as the 1998 mega-flood. In order to
152 effectively deal with the risk of floods and droughts, the Nierji Reservoir was constructed along the
153 mainstream NRB (Fig. 1); it started normal operation in 2006. The drainage area of the reservoir
154 accounts for 22.8% of the NRB. The Nierji Reservoir, located in the upper Nenjiang River (Fig. 1), has
155 flood control and water supply as the primary purposes and hydropower generation and navigation as
156 secondary purposes, thus playing an important role in the distribution of water resources for the lower
157 NRB.



158
 159 **Figure 1.** Location of the Nenjiang River Basin and the distribution of wetlands, river networks, Nierji
 160 Reservoir, and hydrological and meteorological stations within the basin.

161
 162 Table 1 The drainage area of the ten hydrological stations used in this study, area ratios of wetlands and
 163 their contributing areas to the drainage area of the Nenjiang River Basin, northeast China.

ID	River	Hydrological stations	Drainage area (km ²)	Wetland area ratio (%)	Wetland contribution area ratio (%)
1	Mainstream	Shihuiyao	17205	22.2	54.7

2	Duobukuli River	Guli	5490	16.3	57.1
3	Menlu River	Huolongmen	2151	20.8	50.7
4	Mainstream	Kumotun	32229	20.4	54.3
5	Keluo River	Kehou	7310	23.4	56.2
6	Gan River	Liujiatun	19665	13.2	49.9
7	Mainstream	Nenjiang	61249	18.3	54.1
8	Mainstream	Tongmeng	108029	13.1	47.5
9	Mainstream	Fulaerji	123911	13.7	39.0
10	Mainstream	Dalai	221715	16.3	42.4

164

165 The driving datasets used in this study include meteorological data, land-use/land-cover types, soil
166 texture, digital elevation models, drainage network, and observed discharge data. The land-use/land-
167 cover types for 2015 (including wetland types), digital elevation models and digital elevation models
168 with 1 km resolution were obtained from Resource and Environment Science and Data Center
169 (<https://www.resdc.cn/>). The river network was collected from the Geographical Information
170 Monitoring Cloud Platform (<https://www.dsac.cn/DataProduct/Index/30>). Historical daily
171 meteorological datasets including precipitation and air temperature for the period 1963-2020 were
172 obtained from 39 weather stations administered by the National Meteorological Information Centre of
173 China (<http://data.cma.cn>) and 49 weather stations in the upper NRB (Fig. 1) administered by the
174 Nenjiang Nierji Hydraulic and Hydropower Ltd. Company (<http://www.cnnej.cn>). The hydrological
175 data from ten hydrological stations (see Fig.1 and Table 1) were obtained from the Songliao Water
176 Resources Commission, Ministry of Water Resources (<http://www.slwr.gov.cn/>), with the time series
177 extending from 1963 to 2020.

178 In this study, we drove hydrological model using five GCM projections (GFDL-ESM4, IPSL-CM6A-
179 LR, MPI-ESM1-2-HR, MRI-ESM2-0, UKESM1-0-LL) under three Socioeconomic Pathways (SSPs)
180 from the latest CMIP6 (O'Neill et al., 2016). Each of these specific SSPs represents a development
181 model that includes a corresponding combination of development characteristics and influences. The

182 three SSPs that were used herein include SSP126, SSP370 and SSP585, which represent potential
183 futures characterized by green-fueled growth (van Vuuren et al., 2017), high inequality between the
184 countries (O'Neill et al., 2016) and fossil-fueled growth (Kriegler et al., 2017), respectively. We chose
185 the five GCM projections because their high resolution (0.25°) and wide application in previous studies.
186 Given the data requirements of the hydrological model, we downloaded the SSP outputs including daily
187 precipitation, maximum and minimum temperature. We then performed bias correction and spatial
188 downscaling of the SSP outputs. The bias correction of SSP outputs was carried out using the CMhyd
189 software (<https://swat.tamu.edu/software/cmhyd>), in which the widely used Delta Change method in the
190 CMhyd software was used. Delta Change bias-corrects the projected SSP outputs based on the historical
191 statistics and thus conserves the linear spatial-, temporal-, and multi-variable dependence structure in
192 the future climate (Bosshard et al., 2011; Maraun, 2016; Moore et al., 2008; Shafeeque and Luo, 2021).
193 The ANUSPLIN package developed by (Hutchinson and Xu, 2004) was then used to uniformly
194 downscale the output from five bias-corrected GCMs to a resolution of 1-km based on the DEM.
195 Following previous studies (Hagemann and Jacob, 2007; Zhao et al., 2021), the multi-model ensemble
196 means (M_{GCM}) of the daily precipitation, and the maximum and minimum temperature under the SSPs
197 scenarios were then obtained to diminish the uncertainties inherited in a single GCM. MEM was
198 calculated using an equally-weighted average:

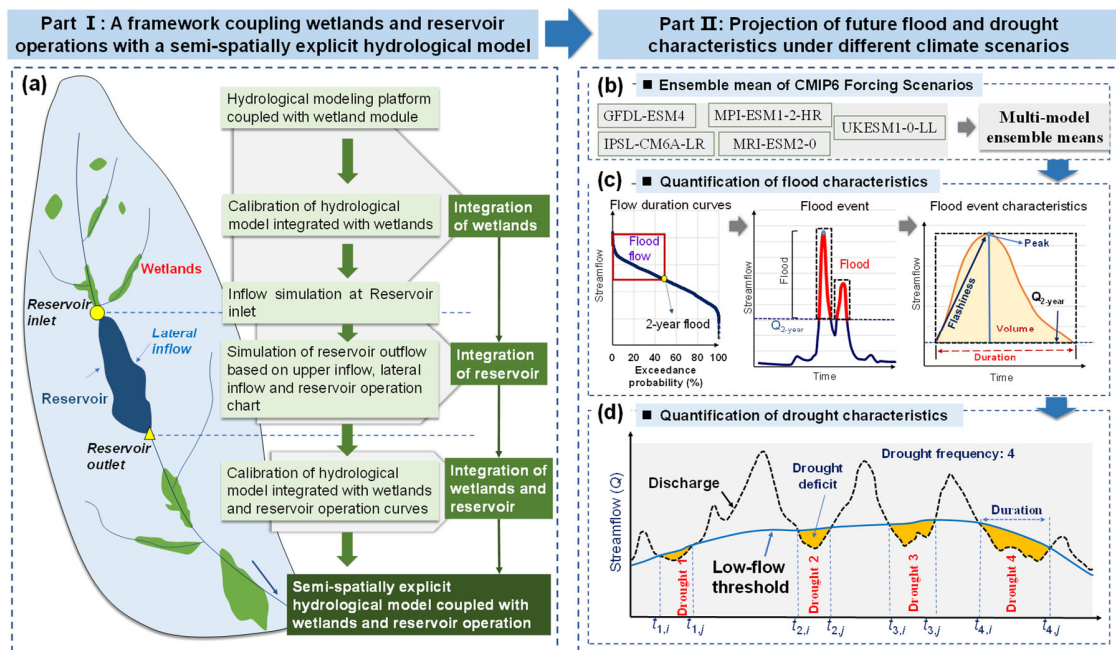
$$M_{GCM} = \frac{1}{N} \sum_{i=1}^N P_i \quad (1)$$

199 where M_{GCM} is the multi-model ensemble means, N is the number of ensemble members (5 in this study);
200 and P_i is the projected climate data of an ensemble member. In this study, the M_{GCM} of five GCMs were
201 used to drive hydrological modeling.
202

203 2.2. Framework of hydrological modeling coupled with wetland modules and reservoir operation 204 scenarios

205 We developed a spatially-explicit hydrological modeling framework that considers wetland
206 hydrological processes and reservoir operations based on HYDROTEL model and reservoir simulation
207 algorithms (Fig.2). Such a modeling framework was based on a distributed coupling implementation at

208 watershed scale from upstream to downstream. Observed streamflow from seven hydrological stations
 209 (see hydrological stations 1-7 in Fig.1) located upstream of the Nierji Reservoir and three hydrological
 210 stations (see hydrological stations 8-10 in Fig.1) installed at downstream of the reservoir, respectively,
 211 were used to calibrate the HYDROTEL model. For the upstream Nierji Reservoir, we calibrated the
 212 HYDROTEL model against observed streamflow of seven hydrological stations with consideration of
 213 wetlands (i.e., hydrologic-wetlands model). Among the seven hydrological stations, the Nenjiang
 214 Station is located at the end of the upstream, where the simulated streamflow was taken as the inflow of
 215 the reservoir. We then computed the reservoir outflow using the simulated inflow, estimated lateral
 216 inflow and reservoir simulation algorithms (see Section 2.2.2), thereby integrating reservoir operation
 217 into the hydrologic-wetlands model to build a hydrologic-wetlands-reservoir model. Based on the
 218 calibrated hydrologic-wetlands-reservoir model, we simulated the outflow of the reservoir (Sect. 2.2.2),
 219 which was used as the input streamflow for downstream model calibration. For the downstream reservoir,
 220 we calibrated the hydrologic-wetlands-reservoir model against observed streamflow of Tongmeng,
 221 Fulaerji and Dalai Stations. Based on this framework, the simulation of basin hydrological processes
 222 coupled with basin scale wetlands and reservoir operations were realized.



223
 224 Figure 2. Framework for projecting future flood and hydrological droughts based on a semi-spatially

225 integrating wetlands and reservoir operation into a hydrological model: (a) a framework coupling
226 wetlands and reservoir operations with a semi-spatially explicit hydrological model; (b) multi-model
227 ensemble means from five GCM projections used for driving modeling framework; (c) methodology
228 for determining a flood threshold, defining flood events, and extracting flood characteristics, and (d) a
229 sequence of runs with examples of drought deficit, duration, and frequency.

230 2.2.1. A semi-distributed hydrological model platform coupled with wetland modules

231 **The PHYSITEL/HYDROTEL modeling platform coupled with two wetland modules (isolated and**
232 **riparian wetlands)** (Fossey et al., 2015b), **has been used to quantify hydrological function of wetlands**
233 (e.g., Fossey and Rousseau, 2016; Blanchette et al., 2019; Wu et al., 2023). PHYSITEL is a Geographic
234 Information System based pre-processing platform for managing hydrological modeling data (Noël et
235 al., 2014; Rousseau et al., 2011). Using general basin data (a digital elevation model, vectorized river
236 network and lacustrine water bodies, and raster-based land use and soil matrix distribution maps),
237 PHYSITEL divides the basin into more detailed hydrological response units, **i.e.**, relatively
238 homogeneous hydrological units (RHHUs) (Fortin et al., 2001). The RHHUs were defined using the
239 algorithm for delineating and extracting hillslopes proposed by (Noël et al., 2014). The hillslopes with
240 same characteristics (e.g., physical geography and hydrological response) were then aggregated within
241 each RHHUs. **In addition, the PHYSITEL platform distinguishes wetlands from other land-use types,**
242 **and then classifies both isolated and riparian wetlands based on an adjacency threshold (i.e., percentage**
243 **of pixels in contact) between the wetlands and the river network (Fossey et al., 2015b). Specifically, if**
244 **more than adjacency threshold (e.g., 1%) of wetland pixels are connected to the river network, they are**
245 **considered as pixels of a riparian wetlands; otherwise, they are referred to as isolated wetlands. It**
246 **subsequently generates data pertaining to isolated and riparian wetlands and their contributing areas.**
247 **The contributing area of wetlands is defined as the sum of the area of all wetland RHHUs and upland**
248 **RHHUs within their immediate catchment areas situated along active fill-spill pathways to the stream**
249 **network (Evenson et al., 2016).** The PHYSITEL platform uses the concept of a hydrologically
250 equivalent wetland (HEW) proposed by (Wang et al., 2008) to integrate isolated and riparian wetlands
251 at the RHHU scale. These typically large RHHUs contain large wetland complexes consisting of various

252 wetland categories such as bogs, fens, marshes, and forested peatlands. After defining the hydrological
253 and wetland parameters, PHYSITEL can directly export the database as part of the input data to
254 HYDROTEL; these data can also be used for other watershed hydrological models.

255 HYDROTEL is a physically-based and semi-distributed hydrological model (Bouda et al., 2014;
256 Bouda et al., 2012; Turcotte et al., 2007) that requires wetland parameter data, land-use type maps, soil
257 texture maps, meteorological data (e.g., daily temperature and precipitation) and daily flows as input.
258 The HYDROTEL model couples the hydrological processes associated with both isolated and riparian
259 wetlands (i.e., the isolated and riparian wetlands modules) at the RHHU scale and calculates the wetland
260 water balance with respect to the surface area of the HEW, contribution area and RHHU. Specifically,
261 for isolated wetlands, the hydrogeological processes are integrated in the vertical water budget (Fortin
262 et al., 2001) at the RHHU scale. For riparian wetlands, the water balance is partially integrated in the
263 vertical water budget of an RHHU and directly connected to the associated river segment via **the**
264 **kinematic wave equation** (Beven, 1981). Based on this, the isolated wetlands modules can realize the
265 vertical water balance processes of hillslope wetlands with land surface runoff processes, while the
266 riparian wetlands modules can realize the interaction of hydrological processes between riparian
267 wetlands and river channels. It should be mentioned that the HEW concept developed by Wang et al
268 (2008) served as the foundation for the integration of riparian wetlands and isolated wetlands into the
269 modeling framework. This concept contends that the features of one HEW (also known as an isolated
270 wetland or riparian wetland) are equivalent to the sum of the characteristics of each wetland inside a
271 RHHU (which could either be hill slopes or elementary sub-watersheds related to one river segment).
272 The following premises apply to this concept: (i) only one isolated and/or riparian HEW per RHHU; (ii)
273 one HEW can be fully integrated within a RHHU; (iii) isolated HEW parameters must be numerically
274 integrated; and (iv) riparian HEW parameters must be numerically integrated and spatially integrated
275 (i.e., located in a specific location on the river segment). Therefore, isolated wetlands and riparian
276 wetlands do not appear to have direct hydrological connection within a RHHU. However, isolated
277 wetlands also have hydrological interactions with riparian wetlands through vertical water balance
278 processes and fill-spill processes (Fossey et al., 2015). Nevertheless, such representations provide a

279 modelling approach that can simulate water balances at the wetland scale while considering their
280 interactions with the surrounding environment (contributing drainage area and hydrological
281 connectivity) (Fossey et al., 2015b). But the hydrological interactions between riparian wetlands and
282 isolated wetlands are not considerate in this study.

283 2.2.2. Simulation of Nierji reservoir operations

284 Based on the simulated runoff at the inlet (the Nenjiang Station), lateral inflow, and the schemes of
285 reservoir operation, we estimated the reservoir outflow using the ResSimOpt-Matlab software package
286 developed by (Dobson et al., 2019) was used to simulate the operation of the Nierji Reservoir.
287 ResSimOpt-Matlab contains three algorithms for reservoir simulation. The first algorithm considers a
288 case when we want to always release a constant amount over the simulation period. This constant
289 amount is the target release that would cover all downstream demand for water, for instance for domestic
290 use and/or irrigation. The second consider a case when we still want to release the target demand but we
291 would also like to (1) apply some hedging (that is, an intentional reduction of the release - even if it
292 would still be feasible to release the target demand - aimed at saving more water and thus facing smaller
293 deficits at later time); and (2) attenuate downstream peak flows for flood control purpose. The third
294 algorithm, which was used in this study, dynamizes the operation rules. A dynamic operation schemes
295 was used in this study to achieve the simulation. Specifically, following (Dobson et al., 2019) and
296 according to actual hydrological conditions, we defined two seasons: the wet season (from June to
297 September) when the risk of flooding is higher and we wanted to release the target demand and provide
298 some storage space for flood control, and the dry season when the risk of flooding is low and the main
299 objective is to sustain ecological baseflows. The required input data to the algorithm includes reservoir
300 inflow (Q_{in}) (m^3/s), the minimum environmental flow (E_{env}), (m^3/s) initial storage (S_0) (m^3), minimum
301 (S_{min}) and maximum (S_{max}) storage (m^3), estimated evaporative losses (E_{vap}) (mm), released discharge
302 (Q_{out}) (m^3/s) and the simulation time-step length (day). Based on the required data, we performed
303 reservoir simulation by implementing the mass balance equation at each simulation time step t :

$$\begin{cases} S_{(t+1)} = S_{(t)} + Q_{in(t)} - E_{vap(t)} - Q_{out(t)} & \text{or} & S_{(t)} + Q_{in(t)} - E_{min(t)} - E_{vap(t)} \\ 0 \leq S_{(t)} \leq S_{max} \\ 0 \leq R_{(t)} \leq \min(S_{(t)} + Q_{in(t)} - E_{min(t)} - E_{vap(t)}, Q_{max}) \end{cases} \quad (2)$$

where S_t is the reservoir storage at time t . S_t and Q_{out} are constrained by the design specifications and operation rules of a reservoir. Specifically, S_t cannot exceed the reservoir capacity S_{max} , while Q_{out} (m^3/s) is constrained by the operation schemes and capacity of the turbines Q_{max} (m^3/s). The excess water, if any, is spilled:

$$Q_{spill(t)} = \max(S_{(t)} + Q_{in(t)} - E_{vap(t)} - Q_{out(t)}) \quad (3)$$

Based on this, the dynamic Q_{out} can be represented using the equation (1) and (2).

We collected information on the reservoir operation including reservoir capacity, control water levels, outflow, the storage-area-water level relationship, the tailwater level-discharge relationship, and the maximum release, along with other data necessary to estimate the outflow. The reservoir inflow is the simulated streamflow at the Nengjiang Hydrological Station, which is at the inlet of the Nierji Reservoir. The minimum storage and maximum storage are 4.9 billion m^3 and 86.1 billion m^3 , respectively. Based on the available data for the study area, the Karrufa method (Kharrufa, 1985) was used to estimate daily evaporative losses from the reservoir. We convert days to seconds so that it would correspond to the flow data. During the wet season, the actual operation schemes for the Nierji Reservoir are as follows: The pre- and post-flood periods are June 1-20 and September 6-30, respectively, with a flood limited water level of 216.0 m; The main flood period is from June 21 to August 25, and the reasonable flood limited water level ranges from 213.4 m to 216.0 m and can be gradually increased. During the dry season, the environmental flow was defined as 25.3% of the daily streamflow during the dry season over the years based on the designed operating curves of the reservoir operation chart.

2.2.3. Model calibration, validation and performance assessment

For all above scenarios, we calibrated the HYDROTEL model against observed streamflow at a daily time step over 8 years, including a 1-year warm-up (2010.10.01-2011.09.30) and a 7-year calibration (2011.10.01-2018.09.30) periods. The same model settings (i.e., key parameters, simulation periods,

328 fitting algorithm, and objective function, etc.) were used for the calibration processes under the both
329 presence and absence scenarios. Following (Arsenault et al., 2018), the model was calibrated using full-
330 time observations without additional validation, as the former allows for more reliable parameters and
331 maximizes the accuracy of the model. The dynamically dimensioned search algorithm (DDS) developed
332 by (Tolson and Shoemaker, 2007) was used to calibrate the 13 most sensitive parameters of the model
333 as proposed by (Foulon et al., 2018). Based on the maximizing of Kling-Gupta efficiency (KGE) (Gupta
334 et al., 2009), automatic calibrations using DDS were carried out utilizing 10 optimization trials (250 sets
335 of parameters per trial). Then, the best set of parameter values out the 10 trials were selected following
336 (Foulon et al., 2018). The KGE was chosen as the objective function because previous research has
337 shown that it can improve flow variability estimates when compared to the NSE (Fowler et al., 2018;
338 Garcia et al., 2017).

339 It should be noted that we calibrated the HYDROTEL model against observed streamflow under with
340 and without wetland scenarios. For the without wetland scenarios are defined as follows: When the
341 wetland modules are turned off in HYDROTEL, wetland areas are not removed, but they are treated as
342 the land cover of saturated soils. Such a saturated soil is fixed and does not participate in hydrological
343 processes such as water yielding and runoff routing, and thus their explicit storage properties are not
344 accounted for in the modeling. This is a basic assumption that has been used in several studies using
345 models such as SWAT (Liu et al., 2008; Wang et al., 2008; Evenson et al., 2015), Mike 11 (Ahmed,
346 2014) and HYDROTEL (Fossey et al., 2016; Fossey and Rousseau, 2016a, b; Wu et al., 2019, 2020a,
347 2021), to quantify the hydrologic services provided by wetlands (flood mitigation, flow regulation and
348 baseflow support etc.).

349 To determine whether coupling the wetland module and the reservoir can improve the model
350 performance, we compared (1) the efficiency of the model in simulating daily flow processes; and (2)
351 the capability of the model to simulate floods and hydrological droughts in the presence or absence of
352 the wetlands and the combination of wetlands and reservoir. Following the recommendations of (N.
353 Moriasi et al., 2007) and (Moriasi et al., 2015), four performance criteria were selected to assess model
354 performance with regards to simulated daily flows with and without the presence of the wetland modules

355 and reservoir operation, namely the Nash-Sutcliffe efficiency (NSE) (Nash and Sutcliffe, 1970),
356 Correlation Coefficient (CC), the root-mean square error (RMSE) and the percent bias (Pbias). We used
357 multiple performance criteria because it may be unreliable to rely on a single objective function to
358 determine whether the model performs well (Fowler et al., 2018; Pool et al., 2018; Seibert et al., 2018).
359 It should be noted that although NSE as an objective function has shortcomings in model calibration, it
360 can still provide an important reference for the evaluation of simulation results as a performance
361 criterion as suggested by Moriasi et al. (2007, 2015). In addition, we compared model performance
362 considering daily hydrograph changes. Furthermore, flood and drought features were extracted (see Sect.
363 2.4.2 and 2.4.3) and used to discern whether, and to what extent, the coupled wetland modules and
364 reservoir simulations could improve the model's ability to simulate droughts and floods.

365

366 2.3. Projection of future flood and drought characteristics under different climate scenarios

367 The calibrated hydrologic-wetland-reservoir model was used to simulate streamflow driven by multi-
368 model ensemble means from the latest CMIP6 and to derive drought and flood characteristics. The flood
369 and drought characteristics were then compared against historical periods to discern how future
370 hydrological extremes will be changed under the influence of wetlands and reservoirs (see Part II in
371 Fig.2).

372 The future simulated streamflow at the Nenjiang and Dalai hydrological stations driven by the
373 ensemble mean of bias-corrected CMIP6 Forcing Scenarios (see Section 2.1) were selected to derive
374 drought and flood characteristics. The Nenjiang Station was chosen because it is located at the outlet to
375 (mouth of) the upper NRB and the inlet to the Nierji Reservoir, whose flood and drought patterns are
376 mainly driven by wetlands and climate change. Moreover, changes in drought and flood characteristics
377 of the Nenjiang Station are critical to the operation of the reservoir immediately lower reach. The Dalai
378 Station, located at the outlet of the entire NRB, was used as a proxy to characterize future flood and
379 drought evolution for the whole basin under the combined influence of the wetlands and reservoir. Using
380 the calibrated hydrologic-wetland-reservoir model, we carried out the simulation of hydrological
381 processes for the historical period (1971-2020) and under the constraints of the SSP126, SSP370 and

382 SSP585 scenarios. We then extracted flood and hydrological drought characteristic indices from the
383 simulations to conduct a comparative analysis of their temporal evolution for the near-future (2026-
384 2050), mid-century (2051-2075) and end-century (2076-2100). The purpose of subdividing the analysis
385 into three time periods was to compare whether, or to what extent, flood and drought characteristics
386 increase or decrease for different future time periods as compared to a historical period.

387 In this study, we characterized floods in terms of four indices consisting of flood peak, flood volume,
388 duration, and flashiness (Fig. 2c). The 2-year flood streamflow was used as a threshold for defining
389 flood events, as it has been often used as a substitute of the threshold for bankfull discharge in previous
390 studies (Cheng et al., 2013; Wu et al., 2020; Xu et al., 2019). Daily streamflows that were greater than
391 the 2-year flood threshold were considered as flood flows. Flood flows occurring on multiple
392 consecutive days were considered as a single flood event. The flood indices, i.e., flood peak, volume,
393 duration, and flashiness were derived with respect to event hydrographs. Flood volume is the cumulative
394 flow from the initial to the end of a flood event with respect to the 2-year flood streamflow level, and
395 represents the flood intensity for different flood events (Wang et al., 2015). The annual total flood
396 volume is the total amount of water associated with all flood events during a water year. We calculated
397 the annual total flood volume based on flood duration and the average amount of streamflow per event
398 in a water year. Flood duration varies for different floods and is, therefore, an important characteristic
399 of a flood event. We summed the flood duration of each event in a water year to obtain the annual flood
400 days. In addition, the annual maximum peak flow was derived from the daily flows to investigate
401 changes in the characteristics of extreme floods. We extracted the 2-year flood threshold for a
402 hydrological station based on the streamflow-exceedance probability curve. Flashiness is a measure of
403 flood severity and is defined as the difference between the peak discharge and action stage discharge
404 normalized by the flooding rise time (Saharia et al., 2017).

405 We characterized hydrological drought characteristics using four indices consisting of the number of
406 droughts, annual drought days, drought duration and deficit (Fig. 2d). A threshold method was used to
407 define hydrological drought events because it can determine the start and end of a hydrological drought
408 event, which allows further assessment of drought characteristics, such as frequency, duration, and

409 **intensity of a drought event** (Cammalleri et al., 2017). It is based on defining a flow threshold (discharge,
 410 Q , m^3/s), below which a hydrological drought event is considered to occur (also known as a low flow
 411 spell). A daily variable threshold, defined as an exceedance probability of the 365 daily flow duration
 412 curves was used to derive drought events from daily streamflow records (Fleig et al., 2006; Hisdal and
 413 Tallaksen, 2003). For rivers with perennial flow, relatively low streamflows ranging from Q_{70} to Q_{95}
 414 have been used as a reasonable threshold (Tallaksen and van Lanen, 2004; Zelenhasić and Salvai, 1987).
 415 In this study, we chose the 90th percentile (Q_{90-n}) streamflow as the daily threshold, **which also used as**
 416 **the threshold for identifying droughts in future climate change scenarios.** The Q_{90-n} of all days was
 417 determined based on the observed historical daily streamflow.

418 To enable the comparison across different modeling scenarios (i.e., historical scenarios and future
 419 climate change scenarios), we derived drought days, deficit, duration, and number from identified
 420 hydrological drought events to characterize their patterns. Drought volume deficit was calculated by
 421 subtracting daily streamflow from the threshold level (Q_{90-n}) during a drought event, and it presents the
 422 severity of the drought compared to the normal streamflow conditions. Drought duration was the
 423 cumulative number of days during a drought event, i.e., the number of days from the beginning to the
 424 end of the drought. Annual drought days were then the cumulative drought duration in a year. **The**
 425 **number of droughts is expressed by the number of drought events during a study period.** In addition, the
 426 annual minimum flows of each water year were extracted and used to determine the model's ability to
 427 simulate very low flows. The drought volume deficit was calculated as:

$$428 \quad D_k = \sum_{t_i}^{t_j} (Q_{90,t} - Q_n) \cdot 60 \cdot 60 \cdot 24 \quad (4)$$

429 where D_k is the drought volume deficit (m^3) of a drought event k at a hydrological station and $t_{k,i}$ and $t_{k,j}$
 430 are the initial and final time steps of the run, respectively. Q_n is the daily streamflow of n day of the year
 431 (1-365). The corresponding drought duration is computed as $t_j - t_i + 1$.

432 For hydrological drought events that occur relatively close in time, the inter-event time method
 433 introduced by (Zelenhasić and Salvai, 1987) was used to separate events. This method defines a
 434 minimum gap period, t_c , and assumes that if the inter-event time $(t_j - t_i + 1) < t_c$, then the consecutive events

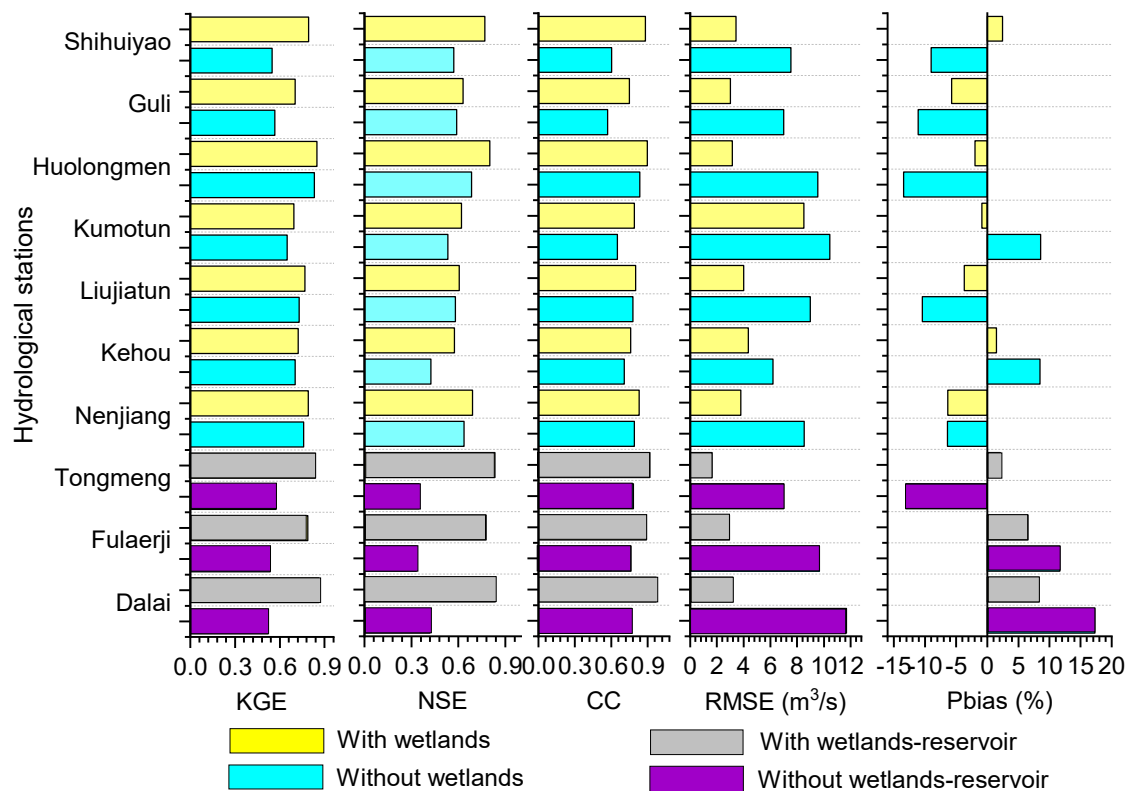
435 are interdependent and merged. In this case, the total drought deficit volume is the sum of the individual
436 deficit values, and the event duration is the so-called real drought duration (sum of the single event
437 duration, excluding excess periods). For this study, t_c was set equal to 7 days as recommended by
438 (Cammalleri et al., 2017).

439

440 **3. Results**

441 3.1 Model performance on daily streamflow and hydrography

442 Fig. 3 depicts model performances for calibration results in the presence or absence of the wetlands
443 and the combination of wetlands and reservoir at the ten hydrological stations in the NRB. In the case
444 of whether the wetlands were present or absent, the simulated daily streamflow results all achieved the
445 acceptable performance criteria ($NSE > 0.5$ and $Pbias \leq \pm 15\%$) suggested by (Moriasi, 2007) and
446 (Moriasi et al., 2015) at the Shihuiyao, Guli, Huolengmen, Kumotun, Kehou, Liujiatun and Kumotun
447 stations. However, compared with the calibrated results of the model without wetlands, the simulation
448 efficiency under with wetland scenario improved to varying degrees. Specifically, the relative
449 improvement (i.e., the relative change) of KGE values at Shihuiyao, Guli, Huolengmen, Kumotun,
450 Kehou, Liujiatun, Kumotun, Tongmeng, Fulaerji and Dalai were 44%, 24%, 2%, 6%, 5%, 3%, 4%, 46%,
451 47% and 67%, respectively. In addition, the NSE and CC values were generally larger in the presence
452 of wetlands than those in the absence of wetlands, and the RMSE and Pbias values are generally smaller
453 than those in the absence of wetlands, showing that integrating wetlands into the hydrological model
454 can slightly improve the model calibration results.



455

456 Figure 3. Model performances for calibration results for the with/without wetlands and reservoir
 457 scenarios at the ten hydrological stations in the Nenjiang River Basin. The KGE, NSE, CC, KGE,
 458 RMSE and Pbias refer to Kling-Gupta efficiency, Nash-Sutcliffe efficiency, Correlation Coefficient,
 459 Root Mean Square Error, and the percentage bias, respectively.

460

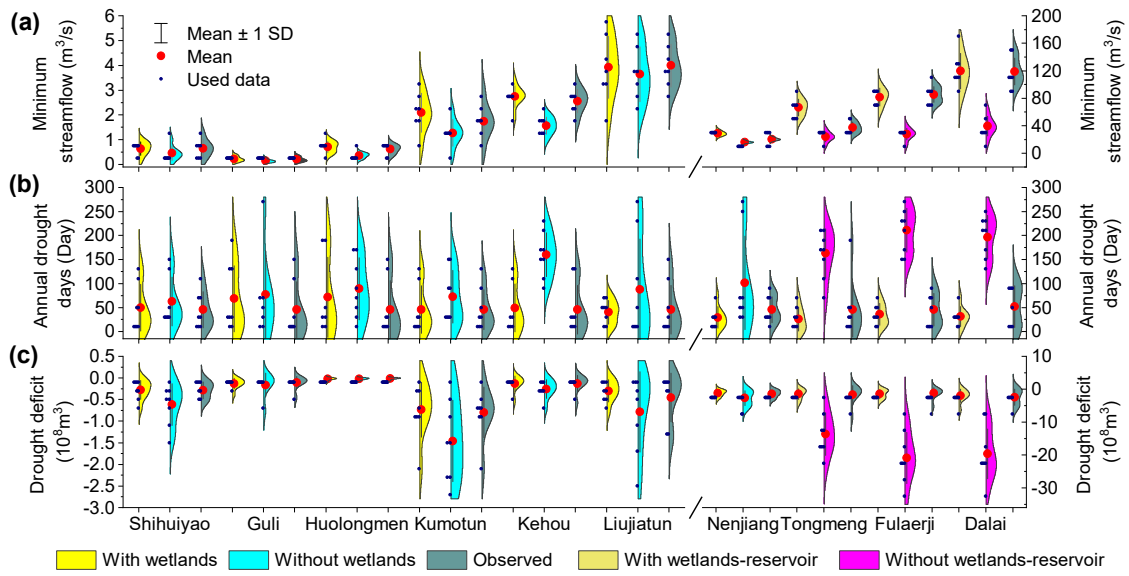
461 For the lower reaches of Nierji Reservoir (i.e., the Tongmeng, Fulaerji and Dalai stations, representing
 462 inclusion of the wetlands and the reservoir operation into hydrological modeling), the NSE and CC
 463 values were greatly higher and RMSE and Pbias values were substantially lower when the wetlands and
 464 reservoir were considered, in comparison to the case without wetlands-reservoir (Fig. 3). In fact, in the
 465 scenario without wetlands-reservoir, the simulated daily streamflow results failed the acceptable
 466 performance criteria ($NSE > 0.5$ and $Pbias \leq \pm 15\%$ as suggested by Moriasi (2007) and Moriasi et al.
 467 (2015). In addition, the simulated daily streamflow in the no-wetland and no wetlands-reservoir
 468 scenarios both overestimated the high flows, especially those during the flood periods; during the low
 469 flow periods, the low flows were underestimated (Please refer to Fig. S1 in Supplementary materials).

470 Further, the simulated hydrographs under the wetland and wetlands-reservoir scenario were in much
471 better agreement with the hydrographs of observed streamflow, especially during floods and the low-
472 flow period (Please refer to Fig. S2 in Supplementary materials). These results indicate that inclusion
473 of the wetlands and the operation of reservoirs can greatly improve model capacity to replicate basic
474 hydrograph characteristics and capture hydrological extremes (e.g., high and low flows).

475 3.2 Model capacity to replicate flood and drought characteristics

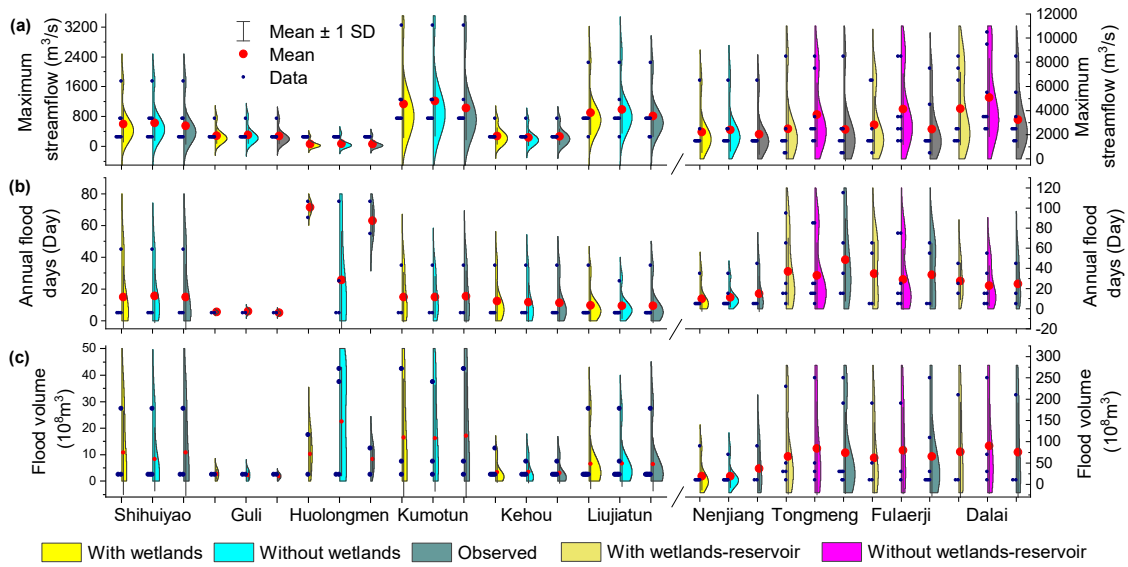
476 The simulated annual minimum streamflow for the wetlands/wetlands-reservoir scenarios were, in
477 general, slightly overestimated or approximately equivalent to the observations compared to the
478 scenarios that did not include the wetlands/wetlands-reservoir (Fig. 4 and 5). However, the simulation
479 results without wetlands clearly underestimated minimum streamflow (Fig. 4a), distinctly overestimated
480 annual drought days (Fig. 4b) and drought deficit (Fig. 4c) compared to the simulation results for the
481 scenario with wetlands at the ten hydrological stations. In addition, the simulated annual maximum peak
482 flow (Fig. 5a), flood days (Fig. 5b) and volume under (Fig. 5c) the with/without wetland scenarios are,
483 in general, approximately comparable to observations at the Guli, Kumotun, Kehou, Liujiatun and
484 Nenjiang hydrological stations (Fig. S4). Specifically, for the upstream Nierji Reservoir, it is apparent
485 that if wetlands are not considered, the number of annual flood days will be overestimated, whereas
486 flood volume will be substantially underestimate at the Huoloengmen Station. For the lower reach of
487 Nierji Reservoir, lack of integrating the wetlands and reservoir into the simulation can lead to a notable
488 underestimation of annual flood days, and a substantial overestimation of the annual maximum peak
489 flow and flood volume. These results demonstrate that integrating wetlands and the combination of
490 wetlands and the reservoir into the model can help improve model performance with regards to flow
491 during the calibration process, and enhances the model's capability of depicting streamflow processes
492 as well as capturing flood and drought characteristics.

493



494

495 Figure 4. Annual minimum streamflow, drought days and deficit derived from observed records and
 496 simulated streamflow at ten hydrological stations in the Nenjiang River Basin. The with and without
 497 wetlands/wetlands-reservoir refers to streamflow simulation based on the presence or absence of
 498 wetlands/wetlands-reservoir.



499

500 Figure 5. Annual maximum peak flow, flood days and volume derived from observed records and simulated
 501 streamflow at ten hydrological stations in the Nenjiang River Basin. The with and without wetlands/
 502 wetlands-reservoir refers to streamflow simulation based on the presence or absence of wetlands/
 503 wetlands-reservoir.

503

504 3.3 Projection of future floods

505

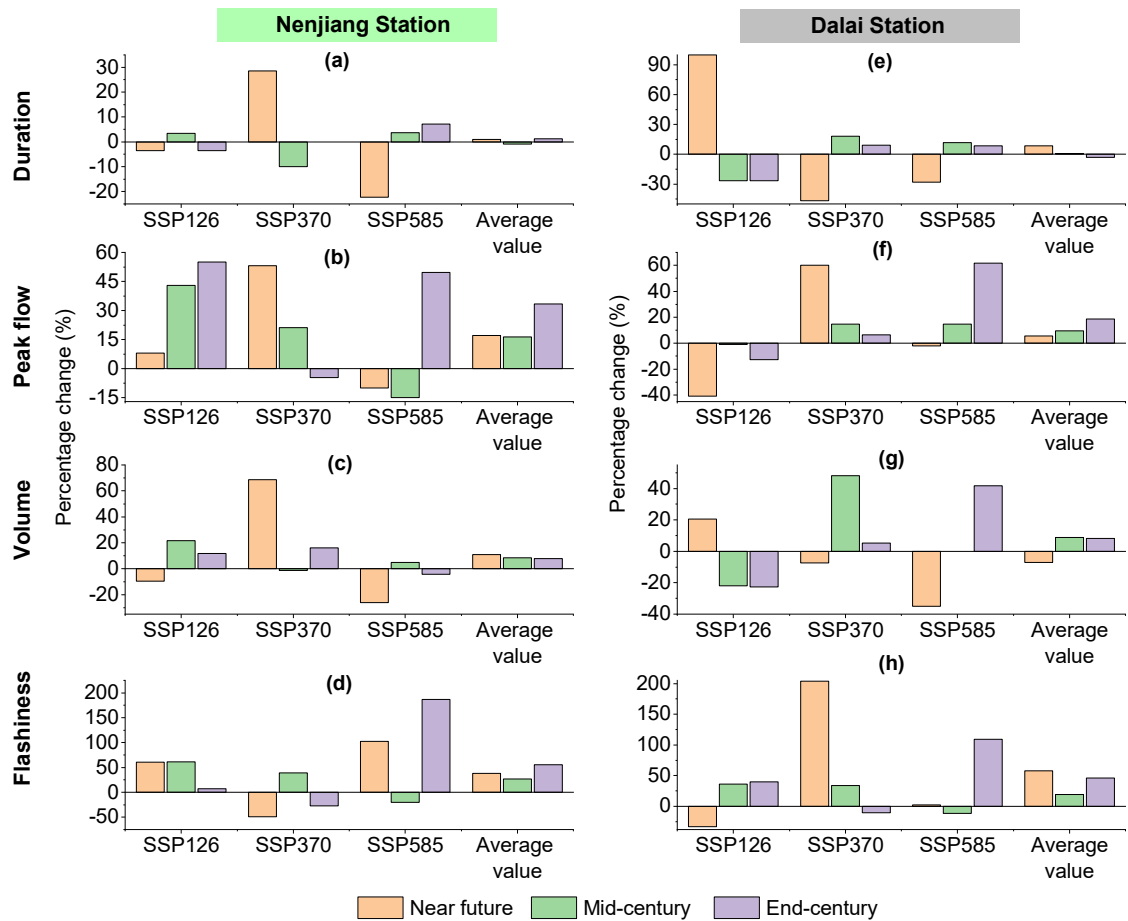
506 A comparison between historical and projected flood characteristics at the Nenjiang Station
 507 (representing inclusion of wetlands into hydrological modeling) shows an overall increase in flood risks
 508 in the upper NRB. The flood duration, peak flow, volume and flashiness generally exhibit larger
 509 fluctuations in most of the scenarios (different SSPs and three periods as shown in Fig. S3 and Table
 510 S1). In addition, the averaged increase in flood duration, peak flow, volume and flashiness ranges from
 511 0.9 to 1.2%, from 16 to 33%, from 8 to 111% and from 26 to 55%, respectively (Fig. 6). Specifically,
 512 the extreme values of flood duration are much larger during the near future and end-century under the
 513 SSP126 scenario, the end-century under the SSP370 scenario and the mid- and end-century under the
 514 SSP585 scenario (Fig. 6a). Apart from a slight decrease during the near future and mid-century under
 515 the SSP585 scenario, peak flow will increase through time in the SSP126, SSP370 and SSP585 scenarios
 516 (Fig. 6b). Simultaneously, the flood volume will experience the greatest increase of 68% during the near
 517 future under the SSP585 scenario, followed by a 22% increase during the mid-century under the SSP126
 scenario (Fig. 6c). In terms of flashiness, the floods will be more severe under the constrains inherent

518 in the SSP126 and SSP585 scenarios and less severe given the conditions in the SSP370 scenario, as
519 compared to the historical period (Fig. 6d).

520 It should be noted that the flood duration, peak flow, volume and flashiness can decrease in the future,
521 as compared to the historical period (Fig. 6). For example, flood duration will slightly decrease during
522 the near future and end-century under the SSP126 scenario, largely decrease during the near future under
523 the SSP585 scenario, respectively. Under the SSP585 scenario, the flood peak flow will experience a
524 decrease with the percentage change values of 15% during the mid-century and the volume will reduce
525 26% during near future. In addition, future flood flashiness will be reduced by 49% and 28% in the near
526 future and the end-century under the SSP370 scenario respectively, and by 21% at mid-century under
527 the SSP585 scenario.

528 The changes in the historical and future flood duration, peak flow, volume and flashiness at the Dalai
529 Station (representing inclusion of downstream wetlands and reservoir operation into hydrological
530 modeling) is shown in Fig. S3 and Fig.4 e-h. Similar to the Nenjiang station, the flood duration, peak
531 flow, volume and flashiness at the Dalai station also exhibit divergent change trends across different
532 SSPs and three periods, as compared to the historical periods. Flood duration is projected to increase
533 largely in the near-future period for the SSP126 scenario, both in the mid-century and end-century for
534 the SSP370 scenario (Fig. 6e). The peak flow will broadly decrease for the SSP126 scenario, and
535 increase for the SSP370 and 585 scenarios (Fig. 6f). Flood volume shows divergent change trends under
536 the three SSPs (Fig.4g). For the SSP126 scenario, flood volume will grow in the near-future and
537 diminish in the mid- and end-century. Flood volume will decrease in the near-future, increase in the
538 mid-century, and increase slightly in the end-century under the SSP370 scenario. However, following
539 an apparent reduction in the near-future, flood volume is anticipated to have no discernible change trend
540 in the mid-century and a clear increasing trend in the end-century for the SSP585 scenarios. Flashiness
541 will be reduced in the near future and will increase in the mid-century and end-century for the SSP126
542 scenario (Fig. 6h). For the SSP370 scenario, flashiness will increase substantially with percentage
543 changes of 204% in the near-future, Moreover, for the SSP585 scenario, flashiness will experience a
544 considerable increase with percentage changes of 109% in the end-century respectively. In terms of the

545 averaged percentage change values, the peak flow and flood flashiness will overall increase; the flood
 546 volume will reduce in the near future and rise in the mid-century and end-century; and flood duration
 547 will experience a slight increase to a minor decrease.
 548

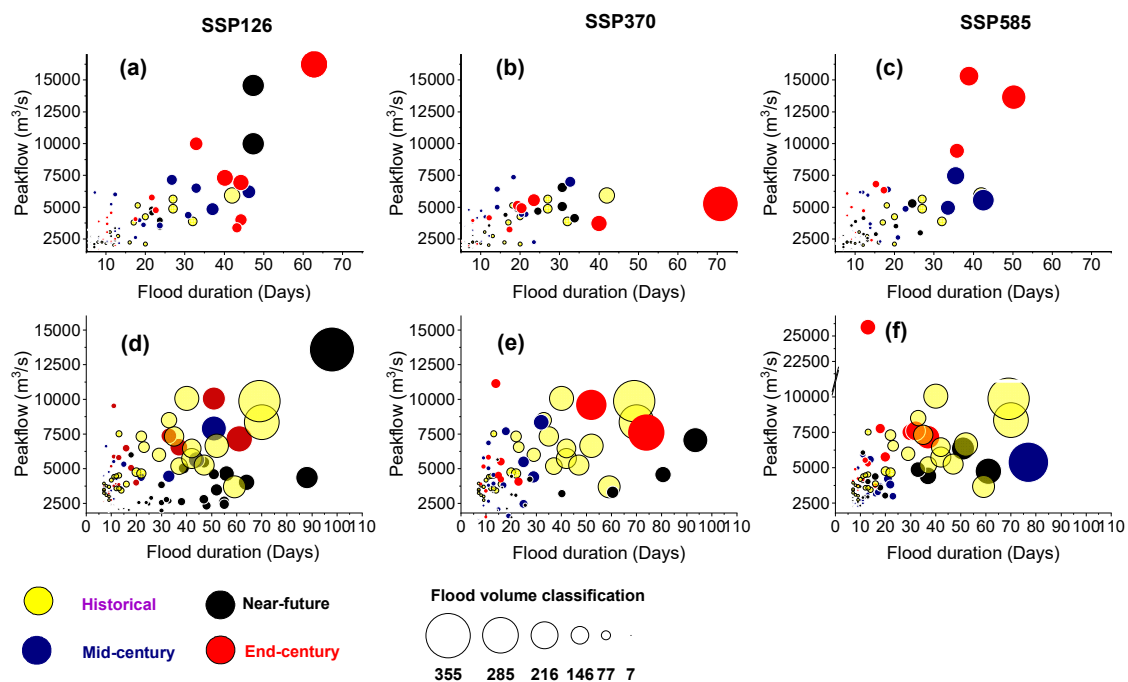


549
 550 Figure 6. Projected percentage changes (relative to historical period during 1971-2020) in flood duration,
 551 peak flow, volume and flashiness at the Nenjiang (the left column) and Dalai (the right column) Station. The
 552 near-future, mid-century and end-century refer to the 2026-2050, 2051-2075 and 2076-2100 under the
 553 Socioeconomic Pathways (SSP) 126, SSP370 and SSP585 scenarios. The average values were calculated
 554 based on the projected percentage changes in the three SSP scenarios.

555
 556 To further investigate flood risks in the NRB under future climate change, the flood duration-peak
 557 flow-flow volume relationships at the Nenjiang and Dalai stations for the SSPs were compared to those

558 of the historical period and analyzed (Fig. 7a-c). Compared with historical flood risk, extreme flood
 559 events with longer and larger volumes will occur more frequently at the Nenjiang Station for the SSP126
 560 and SSP585 scenarios (Fig. 7a and 7c). It is noteworthy that the flood peak-volume-duration
 561 relationships between the historical period and SSP370 scenario are approximate equal, with the
 562 exception that longer duration and larger volume floods will occur during the end-century period (Fig.
 563 5b). In addition, extreme flood events will occur mainly in the near-future for the SSP126 scenario and
 564 during the mid- and end-century for the SSP585 scenario. Moreover, for the SSP370 and SSP585
 565 scenarios, floods will become shorter in duration, and possess a lower peak flow and flood volume in
 566 the near-future. Thus, the upper NRB will experience more severe flood events to a large extent under
 567 most future climate change.

568



569

570 **Figure 7.** Historical and projected flood duration-peak flow-volume relationships at the Nenjiang (the
 571 first row) and Dalai (the second row) Station. The historical period refers to 1971-2020 and the near-
 572 future, mid-century and end-century refer to the 2026-2050, 2051-2075 and 2076-2100 under the
 573 Socioeconomic Pathways (SSP) 126 (the first column), SSP370 (the second column) and SSP585 (the

574 **third column) scenarios.**

575

576 The duration-peak flow-volume relationships of extreme flood events under future climate change
577 scenarios are closer to those of the historical period at the Dalai Station than at the Nenjiang Station
578 (Fig. 7d-f). For the three future SSPs, the flood events with longer duration, higher peak flows or larger
579 volume than the historical period will occurred infrequently, and the duration, flood volume and peak
580 flow of the other shorter and lower magnitude flood events will generally be attenuated. However, very
581 extreme flood events are projected to occur in the near-future under the conditions of scenario SSP126
582 (Fig. 7d). Likewise, future climate change under the SSP370 scenario and 585 scenarios are projected
583 to result in longer flood events in the near-future and mid-century, respectively (Fig. 7e and 7f).
584 Therefore, the future flood risk can be effectively attenuated to a great extent by the combined influence
585 of wetlands and reservoir. However, the fact that extreme flood events that will still occur in the future.

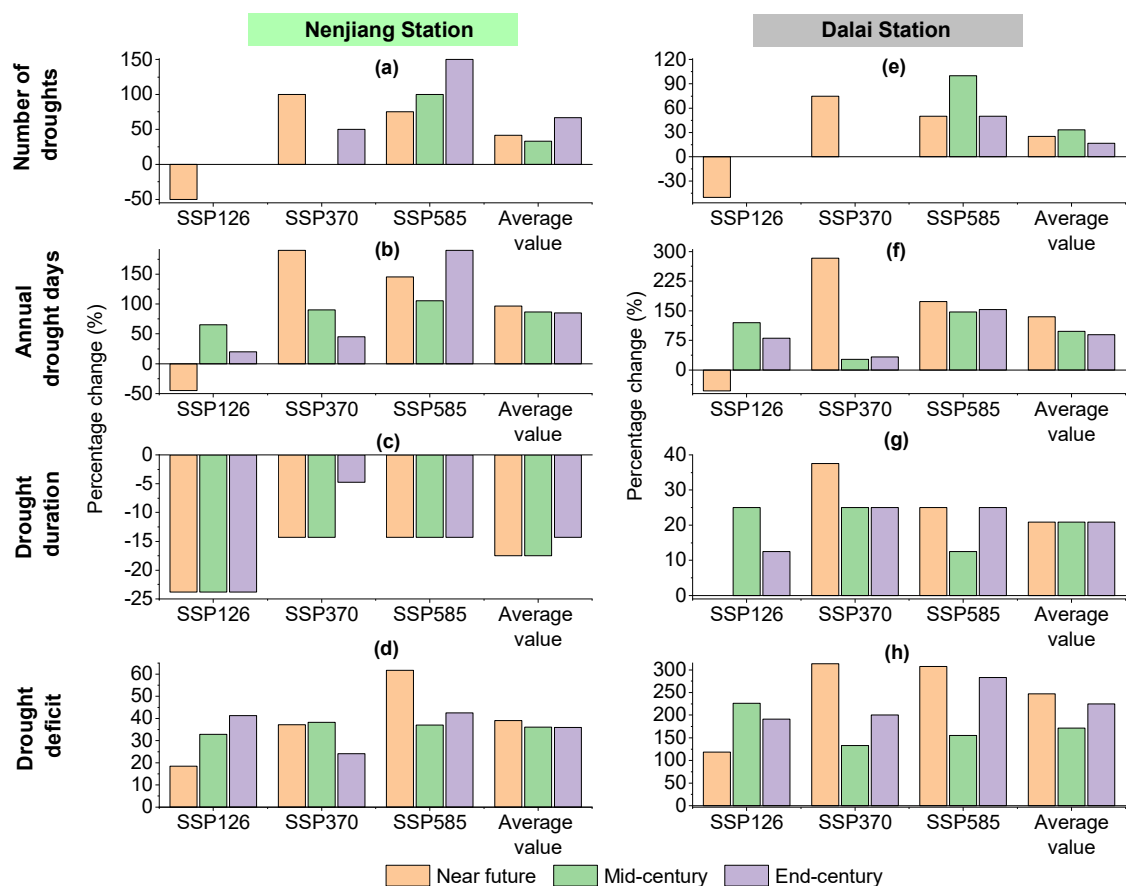
586 3.4. Prediction of future hydrological droughts

587 The comparison between historical and projected hydrological drought indices shows that the risks
588 of hydrological droughts will be increased to some extent under future climate change for both Nenjiang
589 and Dalai stations. Specifically, in addition to a reduction in the number of droughts and annual drought
590 days in the near future for the SSP126 scenario, the number of droughts (Fig. 8a and 8e), annual drought
591 days (Fig. 8b and 8f) and drought deficit (Fig. 8d and 8h) will overall increase in other periods for three
592 scenarios (Figure S4 and Table S2). **It is clear that the number of droughts will be equivalent to the**
593 **historical period in the mid-century and end-century for the SSP126 scenario and in the mid-century for**
594 **the SSP585 scenario.** For all other scenarios, the number of droughts will increase. In terms of the mean
595 percentage change values, there is a general trend towards an increase in the number of droughts and
596 annual drought days, which indicate that future drought events will be more frequent and there will be
597 more days per year affected by drought. The predicted extreme values show that the future duration of
598 drought at Nengjiang station may shorter than the historical period, but the degree of **shortening**
599 presented in different SSP scenarios varies (Fig. 8c and 8g). For the Dalai station, the longest drought
600 durations would all exceed historical extremes in the end-century for the **SSP126** and **SSP585** scenario,

601 and in the near future for the SSP370 scenario. The percentage change values display that drought
 602 duration will be reduced at the Nenjiang station and will be extended at the Dalai station for all the SSP
 603 scenarios. Drought deficit at the Dalai station will increase by 39%, 36% and 36% and in the near future,
 604 mid-century and end-century. For the Dalai station, drought deficit will increase further in the three
 605 periods with 39%, 36% and 36%, respectively.

606 A comparison of the percentage change values between the Nenjiang and Dalai stations shows that,
 607 apart from a reduction of the number of drought events, the risk of drought to be experienced at Dalai
 608 is considerably stronger than at Nenjiang. Specifically, the percentage change in the annual drought
 609 days, drought duration and deficit will increase from 85-97% to 89-134%, from -17- -17% to 21%, and
 610 from 36-39% to 171-247%, respectively.

611



612

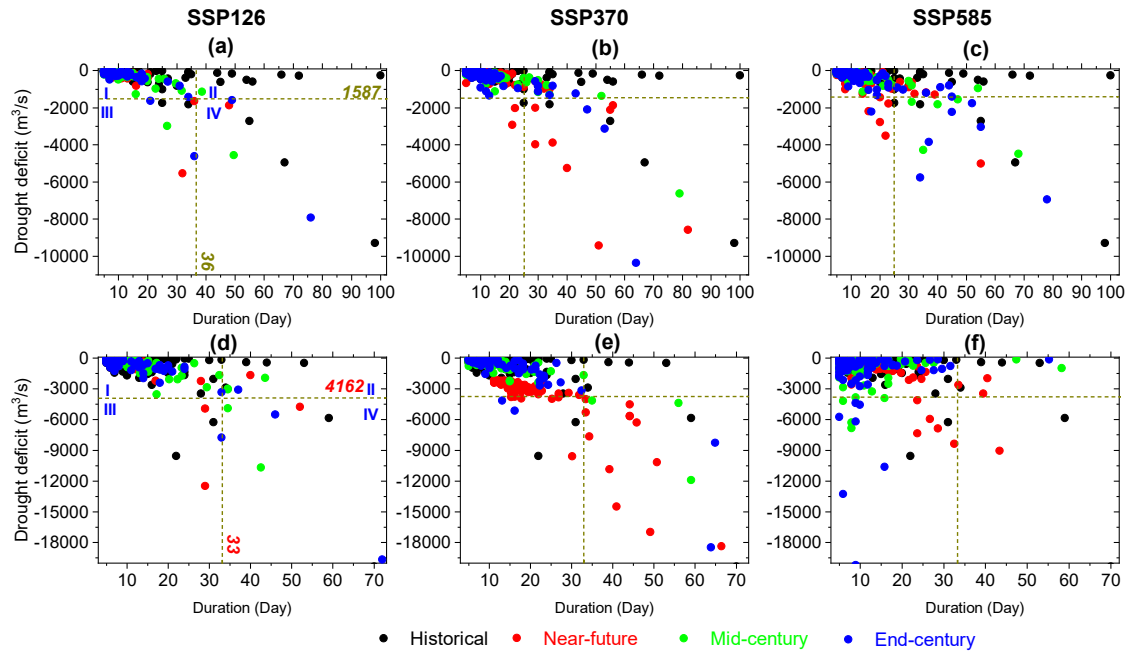
613 Figure 8. Projected percentage changes (relative to historical period during 1971-2020) in hydrological

614 drought characteristics at the Nenjiang (the left column) and Dalai (the right column) Station. The near-future,
615 mid-century and end-century refer to the 2026-2050, 2051-2075 and 2076-2100 under the Socioeconomic
616 Pathways (SSP) 126, SSP370 and SSP585 scenarios. The average values were calculated based on the
617 projected percentage changes in the three SSP scenarios.

618

619 To further analyze the temporal evolution of droughts in the Nenjiang River Basin under future
620 climate change, drought events were classified into four types in terms of duration and deficit, i.e., short-
621 term light droughts, long-term light droughts, short-term severe droughts, and long-term severe droughts
622 (see Fig. 9 for details). This four-part classification was then used to compare and analyze the changes
623 in the temporal characteristics of drought events under the different SSP scenarios. Similar to the
624 drought characteristics during the historic historical period, the majority of drought events for the
625 SSP126, SSP370 and SSP585 scenarios are short-term light droughts (Fig. 9a-c), i.e., the upper NRB
626 will still be dominated by short-term light droughts under future climate change. However, these
627 droughts will be slightly aggravated and marginally longer. In addition, long-term light droughts will
628 occur rarely under the conditions inherent in scenarios SSP126 (Fig. 9a) and SSP370 (Fig. 9b), and
629 occur relative frequently in the SSP585 scenario (Fig. 9c). However, compared with the historical period,
630 the overall number of long-term light droughts will largely decrease, but the deficit will increase slightly
631 under future climate change. In addition, short-term severe droughts will increase substantially, along
632 with their deficit. The number of long-term severe droughts for the SSP126 scenario is approximately
633 the same as in the past, but the duration will be substantially reduced. For scenarios SSP370 and SSP585,
634 the number of long-term severe droughts will increase more than during the historical period, but the
635 duration will be markedly less, and the deficit will be reduced to some extent. In terms of the different
636 the sub-periods, severe droughts in the upper NRB will be more severe during the near-future and end-
637 century periods, and relatively less severe in the mid-century period in comparison to the historical
638 period. However, overall, the droughts will be of shorter duration and characterized by an increased
639 deficit under future climates.

640



641

642 Figure 9. Historical and projected duration-deficit relationship of each hydrological droughts at the Nenjiang
 643 (the first row) and Dalai (the second row) Station. The historical period refers to 1971-2020 and the near-
 644 future, mid-century and end-century refer to the 2026-2050, 2051-2075 and 2076-2100 under the
 645 Socioeconomic Pathways (SSP) 126 (the first column), SSP370 (the second column) and SSP585 (the
 646 third column). The dark yellow lines in the horizontal and vertical directions refer the 95% threshold lines
 647 for drought deficit and duration values, respectively. I, II, III and IV refer to short-term light droughts, long-
 648 term light droughts, short-term severe droughts, and long-term severe droughts, respectively.

649

650 Droughts brought about by future climate change at the Dalai Station located along the lower reaches
 651 of the NRB will continue to be dominated by short-term slight droughts (Fig. 9d-f). For the SSP126
 652 scenario, the duration and deficit of the short-term slight droughts will be approximately the same as
 653 those during historical times (Fig. 9d). However, the duration and deficit of short-term slight droughts
 654 will increase given the conditions specified in the SSP370 (Fig. 9e) and SSP585 (Fig. 9f) scenarios. The
 655 duration of short-term slight droughts will increase the most for scenario SSP370. In addition, under all
 656 three SSP scenarios, long-term slight droughts will, in general, be reduced. In fact, under the SSP370
 657 scenario, long-term slight droughts will not occur. The number of short-term severe droughts will

658 generally tend to increase, with the most pronounced increase under the SSP585 scenario, followed by
659 the SSP370 scenario. A slight increase will occur under the SSP126 scenario. However, long-term
660 severe droughts will increase substantially under the SSP126 and SSP370 scenarios. In particular, under
661 the SSP370 scenario, the duration of long-term severe droughts will be exceptionally prolonged, and
662 the severity will be extraordinarily increased, indicating that the risk of droughts of long duration and
663 with a severe deficit will climb abnormally in some year. For example, under the conditions set by the
664 SSP370 scenario, the deficit of long-term severe droughts will reach -18,169 m³ and -18,457 m³ during
665 the near-future and end-century periods. For the SSP585 scenario, long-term severe drought will occur
666 only once in the near-future, which is equivalent to the historical period. These results indicate that the
667 risk of future hydrologic droughts along the lower NRB will further increase even under the combined
668 influence of reservoirs and wetlands.

669

670 **4. Discussion**

671 4.1. Integrating wetlands and reservoir operation into basin hydrologic modeling and basin water 672 management

673 A series of studies have shown that the simulation and prediction of floods and droughts faces many
674 challenges, such as the scarcity of hydrometeorological driven data (Foulon et al., 2018), model errors
675 (Golden et al., 2021; Smakhtin, 2001; Staudinger et al., 2011) and anthropogenic disturbances (e.g.,
676 reservoir operation) (Brunner, 2021; Brunner et al., 2021). In this study, we developed a spatially
677 explicit hydrological model that considers wetland hydrological processes and reservoir operations
678 through coupling a distributed hydrological modeling platform with wetland modules and reservoir
679 simulation algorithms. We found that coupling wetland alone or coupling wetlands and reservoir with
680 hydrological model can improve model calibration results and model performance of capturing flood
681 and drought characteristics in a large river basin. Such model performance improvement can provide
682 important information for developing downstream water resources management. Previous studies have
683 shown that climate change is further exacerbating the risk of hydrological extremes, leading to an

684 expanding of flood and drought affected area (Diffenbaugh et al., 2015; e.g., Hirabayashi et al., 2013;
685 Wang et al., 2021), which increase the complexity of accurate prediction and the challenge for effective
686 mitigation. Give that, projecting flood and drought risks in response to a changing climate requires
687 robust hydrologic models that take into account the important factors within a watershed that can largely
688 influence basin hydrological processes (Golden et al., 2021). Therefore, in basins that coexist with high-
689 coverage wetlands and multiple reservoirs, it is necessary to integrate wetlands and reservoir operation
690 into basin hydrological simulation, thus providing practical support for extreme hydrological risk
691 mitigation and water resource management under a changing climate.

692 Although our developed framework demonstrates good modeling results, uncertainties could exist in
693 the assessment. Aspects such as the accuracy and error of the input data (Lobligeois et al., 2014), the
694 choice of the objective function (Fowler et al., 2018), the length of the period considered during
695 calibration (Arsenault et al., 2018), and the model structure (Melsen et al., 2019) can all affect the
696 performance of a model to replicate streamflow, thus impacting flood and drought predictions and water
697 management under future climate change. In addition, due to a lack of wetlands water balance
698 monitoring data, this study only used river station data (which only considered the cumulative
699 hydrologic effect of upstream wetlands) for model calibration. Therefore, there are ongoing efforts to
700 obtain sufficient observations on wetland area dynamics and evapotranspiration, water depth and
701 volume, soil water content and actual observations to better calibrate/validate watershed hydrological
702 models, which are expected to better provide key parameters for further improving the model's capacity
703 to capture flood and drought patterns and better serve basin water management. In addition, the several
704 SPPs employed to drive the simulation framework, including SSP126, SSP370, and SSP585 scenarios,
705 can introduce uncertainty into future flood and drought risk projections. Because of the internal
706 variability and uncertainties inherent in the existing climate models (Qing et al., 2020; Martel et al.,
707 2022), the projection findings under different scenarios were inconsistent, creating a challenge for pro-
708 active management and mitigation decisions. Despite the climate models' recognized flaws and
709 uncertainties, the general concordance between models and observations over many regions suggests
710 some improved confidence in their utility for understanding and mitigating future drought and climate

711 change (Cook et al., 2020).

712 4.2. The combining mitigation efficiency of wetlands and reservoir operation

713 The relative changes (compared with historical periods) of future flood and drought indices (Fig. 6
714 and 8), duration-peak flow-volume relationships (Fig. 7) and duration-deficit relationship (Fig. 9) differ
715 between the Nenjiang and Dalai stations under the same SSP scenario or in the same period, indicating
716 that reservoirs and downstream wetlands can modify the continuous propagation of upstream flood and
717 hydrological drought risks to the downstream. First, reservoirs and downstream wetlands can help to
718 reduce the risks of future floods and droughts to some extent, namely partially reduce flood peak flow
719 and flashiness, and decrease the number of droughts, annual drought days and drought deficit. Second,
720 reservoirs and downstream wetlands cannot completely eliminate flood and drought risks. Because the
721 flood duration and volume will overall increase at the Dalai station, especially that the extreme floods
722 will be more frequent in the future (Fig.7). Further, in addition to the number of droughts, the percentage
723 change values of the annual drought days, drought duration and deficit relative change at the Dalai
724 Station are greater than those of Nenjiang Station (Fig.8). This imply that the mitigation effects on
725 hydrological droughts is minimal. Such findings suggest that future climate change will lead to an
726 increase in the risk of hydrologic failure of existing reservoir and wetlands, thus posing large challenges
727 for future socio-and eco-hydrological systems in the downstream NRB.

728 Wetlands are typically viewed as green infrastructures and reservoirs are generally regarded as
729 important gray infrastructures. Although our study showed that the combining of reservoirs and
730 wetlands does not completely eliminate the risk of future hydrological extremes, they continue to play
731 an important role that cannot be ignored. The reservoir's inherent constraints are one factor contributing
732 to this likelihood of hydrological failure. This is because reservoirs only control floods and droughts
733 that occur downstream of them, limiting their effects to the regional scale (Brunner, 2021). The
734 regulation becomes less effective with distance increased due to "dilutions" effect caused by inflows
735 from downstream tributaries (Guo et al., 2012). Reservoirs cannot, however, play a considerable role in
736 basins where tributaries exist downstream, particularly those sub-basins prone to drought and flooding.
737 From these perspectives, widely distributed wetlands can provide a complementary and vital function

738 by providing biological function and hydrological regulation in regions where reservoirs are unable to
739 have an impact. On the other hand, the limited capacity of existing wetlands to regulate hydrology
740 increases the risk of hydrological failure to some extent. This is because, compared with the historical
741 period, the existing wetlands in the NRB have been seriously degraded, such as the weakening of the
742 connectivity between riparian wetlands and the river channel, and the increased fragmentation of
743 wetlands, among other changes (Chen et al., 2021). These degraded wetlands cannot play an effective
744 role in mitigating floods and droughts under future climate change.

745

746 4.3. Implications for flood and drought risk management under climate change

747 This modeling study predicts higher flood and drought risks in the NRB under the combined influence
748 of wetlands and reservoirs. This could impose a great challenge to the operation of the Nierji Reservoir
749 dam, i.e., to its effective operation for flood mitigation and drought alleviation. To curb the flood and
750 drought risks caused by future climate change in the NRB, it is urgent to improve the water regulation
751 capacity of the lower NRB. Although the Nierji Reservoir, as previously argued, plays an important role
752 in reducing floods and droughts, the potential for extreme hydrological events in the future necessitate
753 the application of various combinations of measures with different scales of implementation (i.e., hybrid
754 measures). We insist that the first remedial measure to be undertaken should be the implementation of
755 wetland restoration and protection projects, because studies have demonstrated that wetland coverage
756 and their spatial pattern can affect both basin physical conditions and human decision-making attitudes
757 toward risk (Gómez-Baggethun et al., 2019; Javaheri and Babbar-Sebens, 2014; Martinez-Martinez et
758 al., 2014; Zedler and Kercher, 2005). Given that the spatial location of wetlands within a river basin is
759 also important in determining the efficiency of their mitigation services (Gourevitch et al., 2020; Li et
760 al., 2021; Zhang and Song, 2014), optimization of wetland spatial patterns should be considered and can
761 be carried out to further enhance the role of wetlands in flood and drought defense.

762 In our view, the second important remedial measure that should be implemented is to improve the
763 existing reservoir operation schemes based on accurate hydrological forecasting. This requires, on one
764 hand, coupling of wetlands with hydrological processes and models to improve the simulation accuracy

765 of the upstream incoming water (i.e., runoff from the Nenjiang Station) to provide scientific support for
766 reservoir operation decisions. Concomitantly, it is necessary to modify the existing schemes for optimal
767 reservoir operation to improve the system's capacity to deal with extreme flood and drought risks.
768 Because the percentage increase in flood (Fig. 6) and drought indicators (Fig. 8) demonstrated that the
769 existing reservoir operation schemes are not effective in mitigating the risks associated with future
770 climate change-induced floods and droughts. Therefore, we need to re-examine and evaluate the flood
771 and drought risks in the NRB under future climate change and propose optimal operation schemes that
772 can maximize the reduction of flood and drought risks by the Nierji Reservoir. Traditionally, the water
773 level of a reservoir should be maintained at the designed flood limited water level during the flood
774 season, which does not consider river flow forecast. (Ding et al., 2015) analyzed a concept that provides
775 a dynamic control of the maximum allowed water level during the flood season, for the Nierji Reservoir
776 dam. A reasonable approach to tackle this issue could be to considerate forecast uncertainty and
777 acceptable flood risk to minimize the total loss caused by flood and drought. Further modeling studies
778 with multi-objective optimization algorithm can help identify an optimum reservoir operation for best
779 economic and ecological outcomes.

780 **5. Conclusions**

781 This study projected future flood and drought risks by considering the combined impacts of wetlands
782 and reservoirs. To achieve this, we developed a hydrological modeling framework coupled wetlands
783 and reservoir operations and then applied it in a case study involving a 297,000-km² large river basin in
784 northeast China. With this framework, we found that coupling wetlands and reservoir operations can
785 slightly increase model calibration results and efficiently improve model capacity to capture both flood
786 and hydrological drought characteristics in a river basin. The upper NRB will experience more severe
787 flood and hydrological droughts and can impose a great challenge to the effective operation of
788 downstream reservoir under the predicted future climate change scenarios. The risk of future floods and
789 hydrologic droughts along the lower NRB will further increase even under the combined influence of
790 reservoirs and wetlands. These results demonstrated that the risk of floods and droughts will overall

791 increase further under future climate change even under the combined influence of reservoirs and
792 wetlands, showing the urgency to implement wetland restoration and develop accurate forecasting
793 systems. To fully understand how wetland and reservoir operations may be influential and maintain an
794 acceptable level of risk, it is therefore necessary to consider an optimization of wetland spatial patterns
795 and reservoir operations simultaneously, thus achieving a collaborative optimization management to
796 maximum basin resilience to floods and hydrological droughts. Further, the effects of combining **NBS**
797 (e.g., wetlands) with traditional engineering solutions (e.g., reservoir) should both be useful and
798 necessary in the future for management decisions.

799

800 **Data Availability**

801 The data used in this study are openly available for research purposes. The five GCM outputs (GFDL-
802 ESM4, IPSL-CM6A-LR, MPI-ESM1-2-HR, MRI-ESM2-0, and UKESM1-0-L) used in this study are
803 publically available and were provided by the Inter-Sectoral Impact Model Intercomparison Project
804 (ISIMIP) (<https://esg.pik-potsdam.de/search/isimip/>). The CMhyd software is available at
805 <https://swat.tamu.edu/software/cmhyd>. The land-use/land-cover types, soil texture, and digital elevation
806 model for China can be downloaded from <https://www.resdc.cn/>. Data from the 88 weather stations
807 administered by National Meteorological Information Centre of China can be download at
808 <http://data.cma.cn>.

809

810 **Author contribution**

811 **Yanfeng Wu:** Conceptualization, Writing, Data analysis, Methodology, Software. **Jingxuan Sun:**
812 Formal analysis, Investigation, Data analysis and Plotting. **Boting Hu:** Software, Visualization, Data
813 analysis. **Y. Jun Xu:** Writing - review & editing. **Alain N. Rousseau:** Writing - review & editing.
814 **Guangxin Zhang:** Conceptualization, Supervision, review & editing.

815

816 **Competing interests**

817 The authors declare that they have no known competing financial interests or personal relationships
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819

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